

Joint Task Force of ICP Waters and ICP IM 2017

– presentations Wednesday 8 May 2017 - Thursday 11 May 2017

Biology responses to air pollution

Environmental drivers of leaf litter decomposition in streams,
Andreas Bruder and Julien Cornut, Switzerland

Species sensitivity to acidification in highly endemic regions of South
Africa

Londiwe M Khuzwayo, South Africa

Nitrogen deposition impacts in the Austrian IM site Zöbelboden

Ika Djukic, Austria

Gas supersaturation may cause effects on the biota comparable to
acidification

Gaute Velle, Norway

Co-analysis of coniferous forest state parameters and atmospheric
deposition data series obtained by ICP IM and EMEP at European part
of Russia

Ekaterina Pozdnyakova, Russia

Recovery of benthic algal assemblages from acidification - how long
does it take, and is there a link to eutrophication?

Jakub Hruska, Czech Republic

Critical loads/Dynamic modelling

Soil modelling study VSD+Opens in new window

Maria Holmberg, Finland, ICP IM Programme Centre

Acidification and Critical Loads of Surface Waters in European Territory
of Russia and Western Siberia

Tatyana Moiseenko, Marina Dinu, Russia.

Relationships between critical load exceedances and empirical impact
indicators - Update of N assessment

Jussi Vuorenmaa, Finland, ICP IM Programme Centre

Common Task Force meeting for both ICPs

Chemical intercomparison

Øyvind Garmo, Norway, ICP Waters Programme Centre

Separate meetings: ICP IM TF meeting

Explanatory notes presentation ICP IM TF

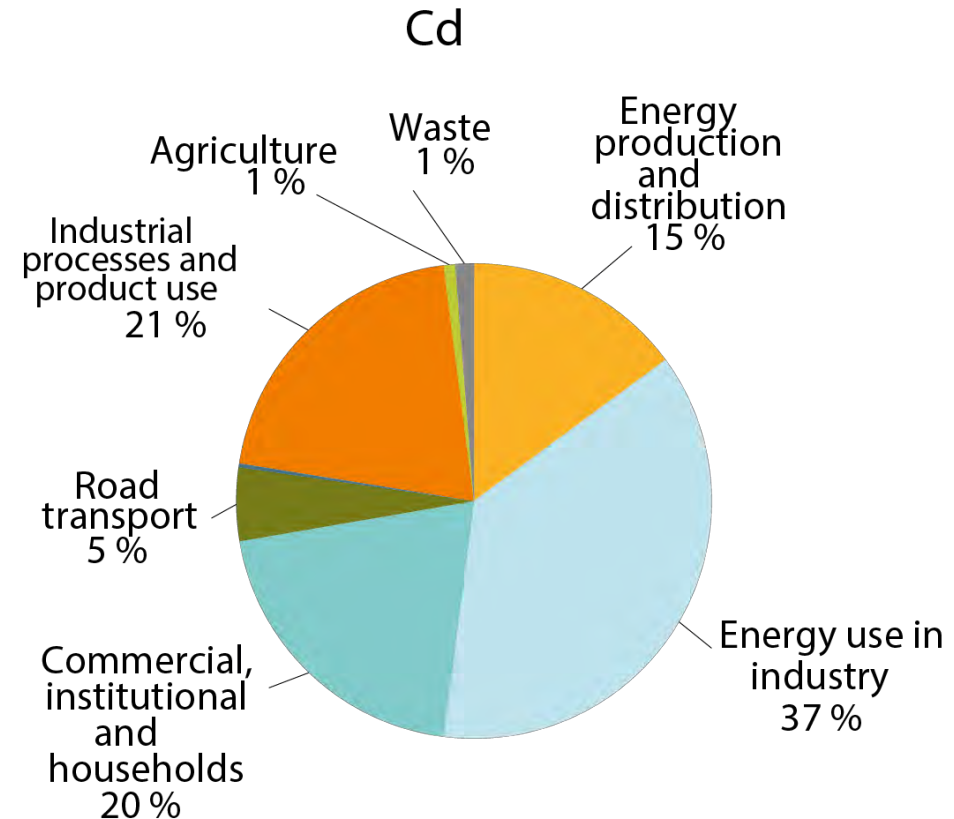
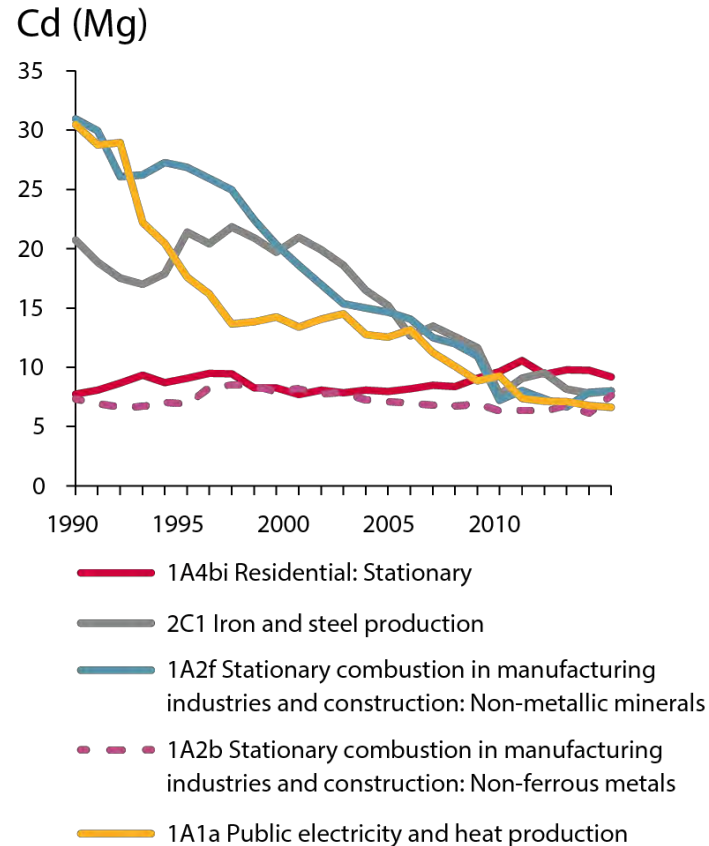
Lars Lundin, Sweden, ICP IM Chair

ICPIM database status

Sirpa Kleemola, Finland, ICP IM Programme Centre

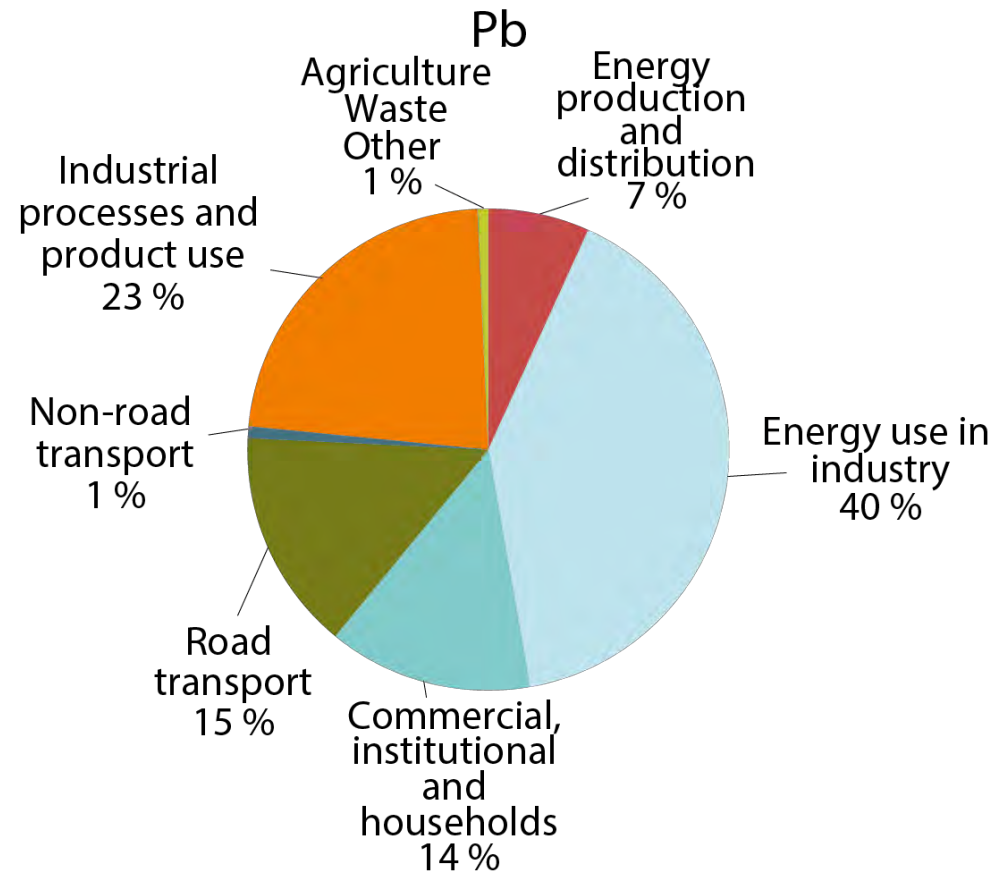
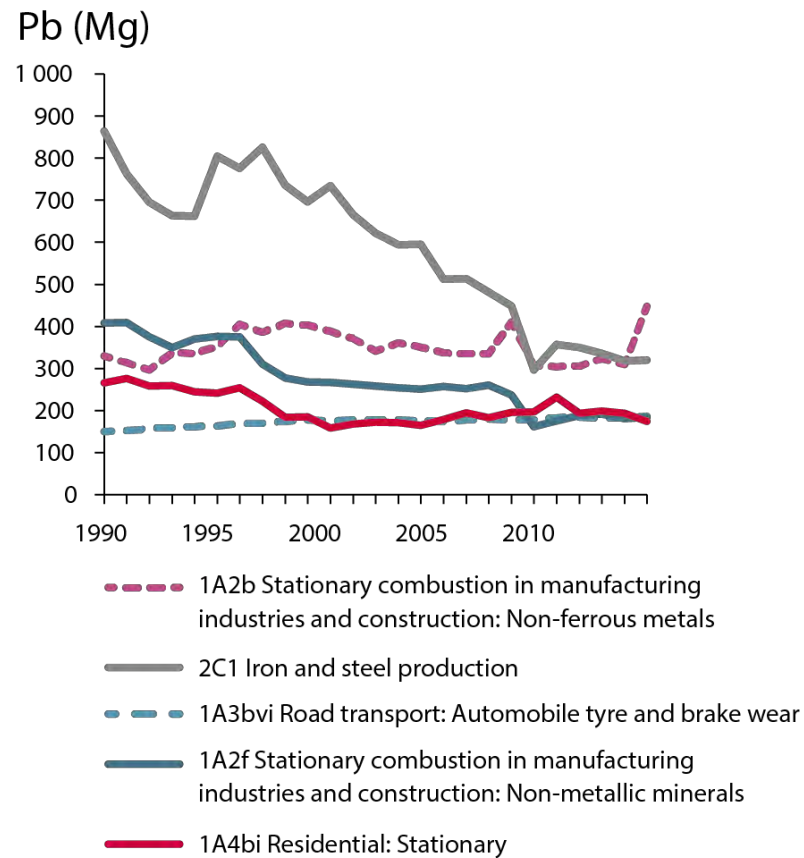
Cd emissions in the EU

From the *European Union emission inventory report 1990–2014 under the UNECE Convention on Long-range Transboundary Air Pollution (LRTAP) (2016)*



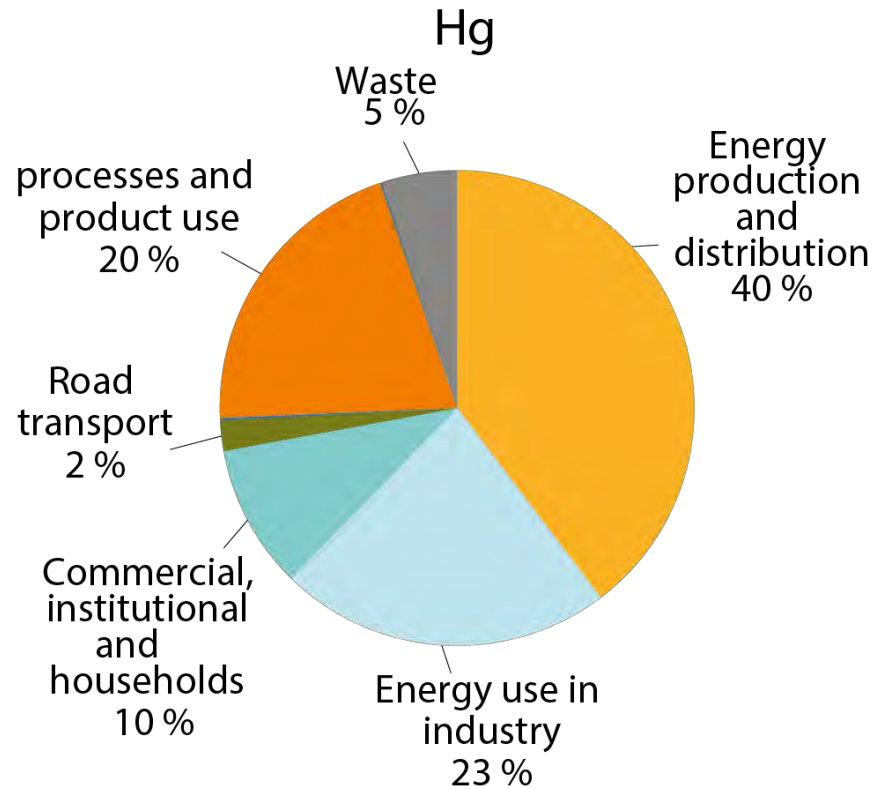
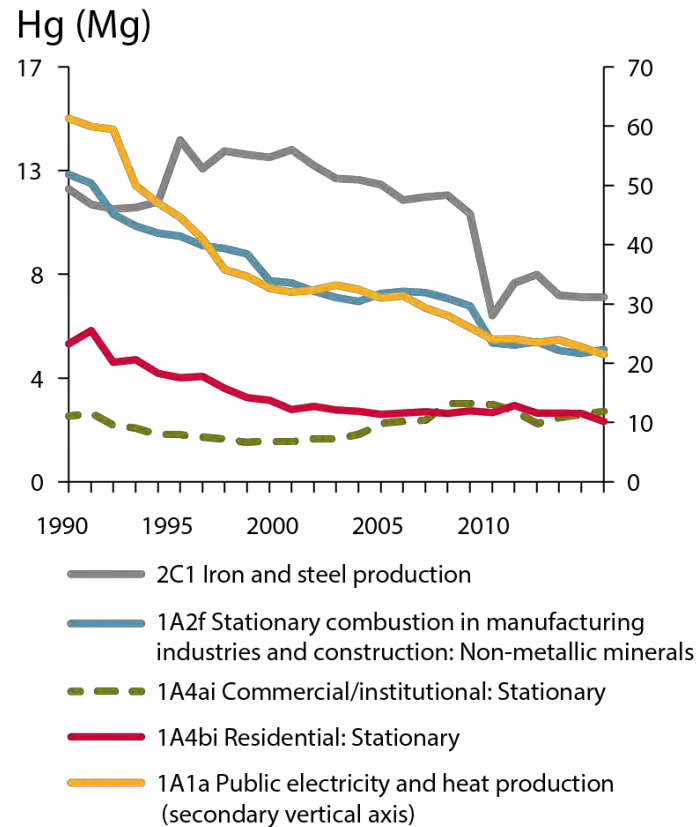
Pb emissions in the EU

From the *European Union emission inventory report 1990–2014 under the UNECE Convention on Long-range Transboundary Air Pollution (LRTAP) (2016)*



Hg emissions in the EU

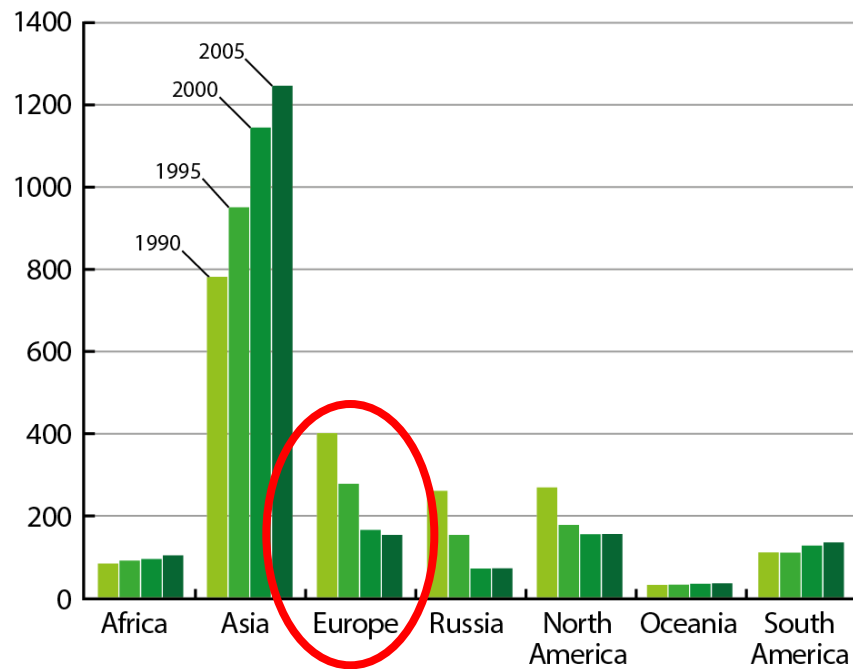
From the *European Union emission inventory report 1990–2014 under the UNECE Convention on Long-range Transboundary Air Pollution (LRTAP) (2016)*



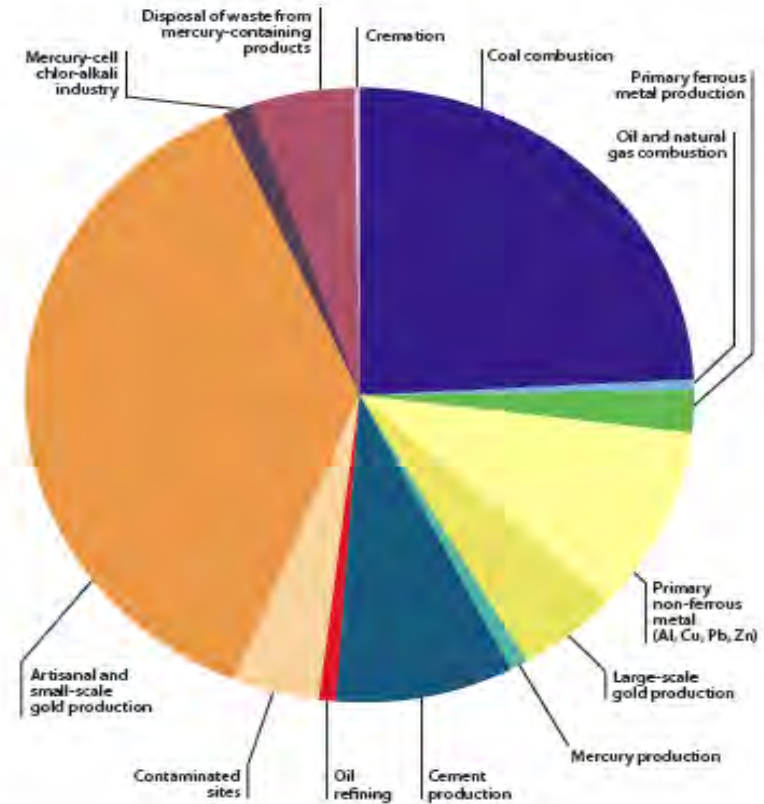
Global Hg emissions

From UNEP, 2013. *Global Mercury Assessment 2013: Sources, Emissions, Releases and Environmental Transport.*

Emissions to air, tonnes

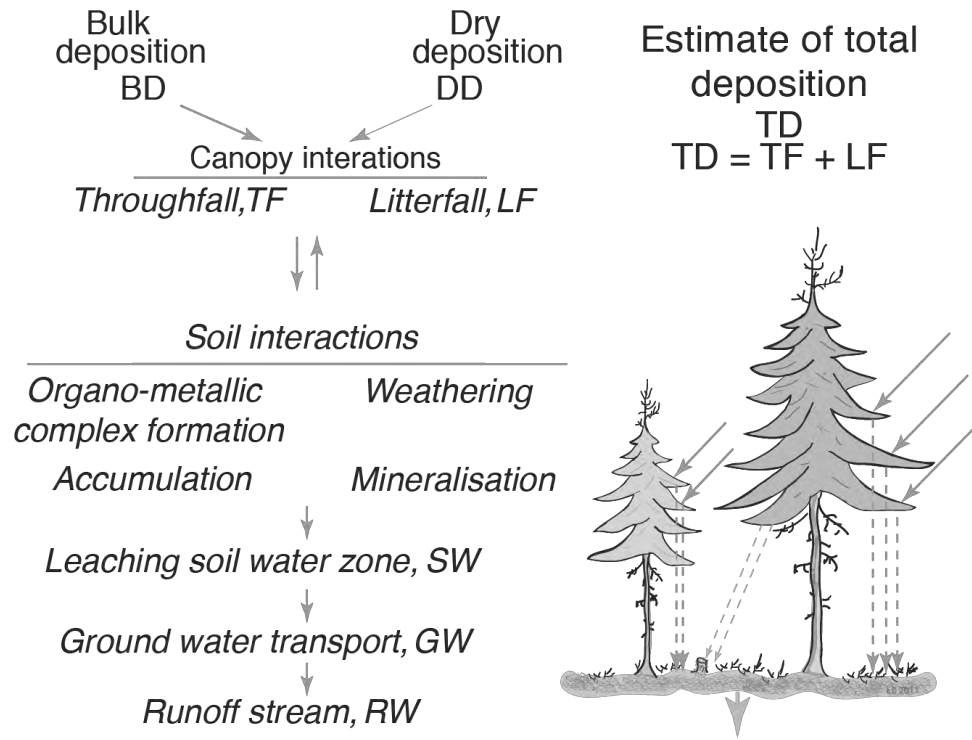


Estimates of annual anthropogenic mercury emissions from different continents/regions, 1990-2005.



Relative contributions to estimated emissions to air from anthropogenic sources in 2010.

Processes involving metals in forest catchment mass balance



$$\text{Retention} = 1 - \text{RW} / \text{Input}$$

$$\text{Input} = \text{TF} (+ \text{LF for Hg and Pb})$$

Retention of heavy metals in catchments:

Cd: (0) – 91%

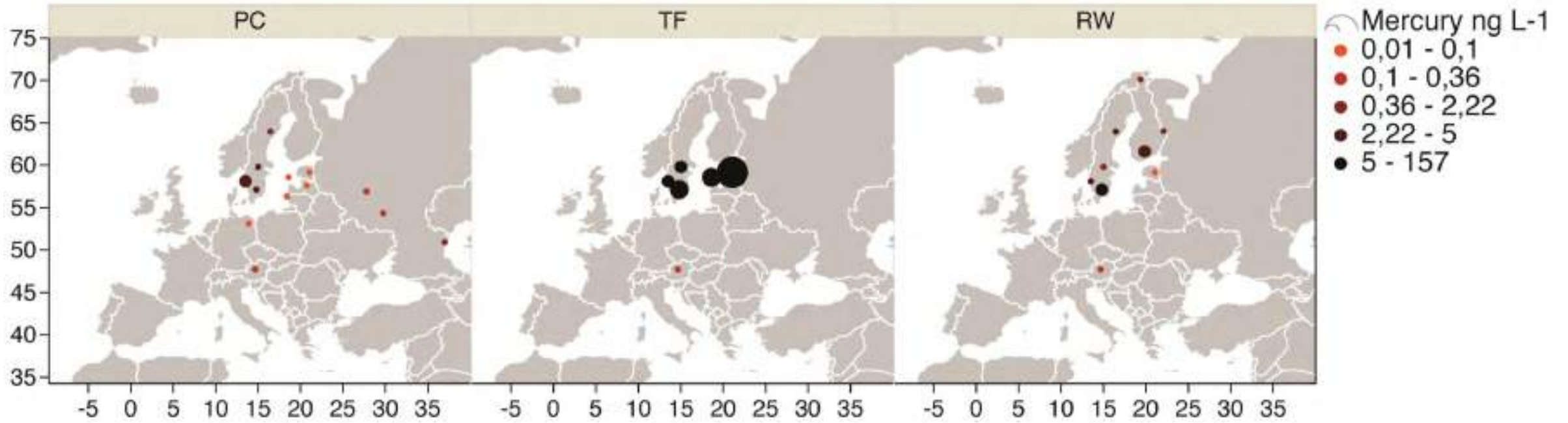
Pb: 74 – 94 %

Hg: 86 – 99 %

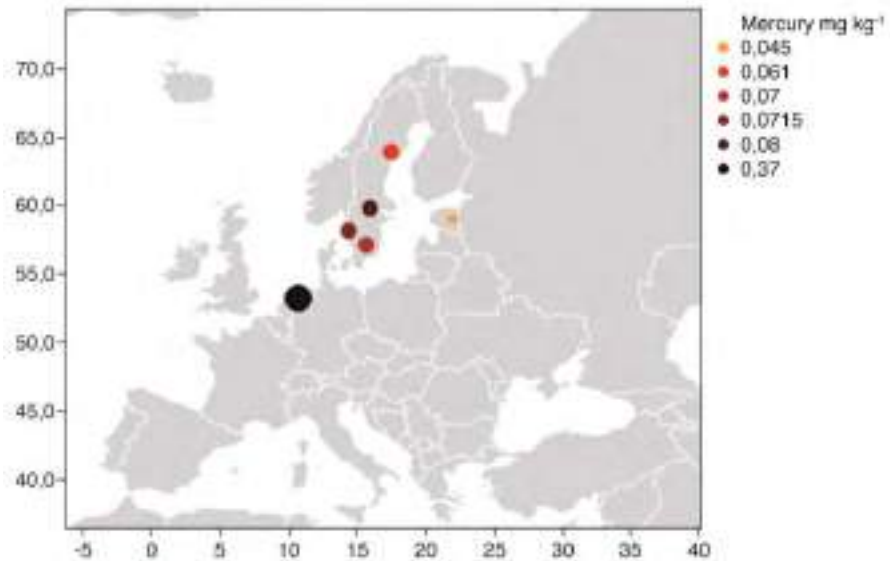
Cu: 80 – 97 %

Zn: 38 – 96 %

Mercury

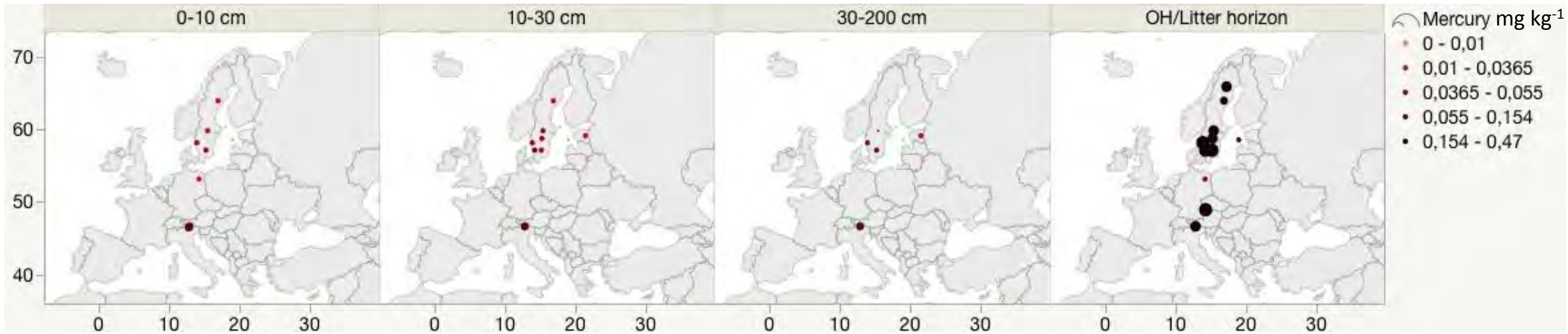


Litterfall



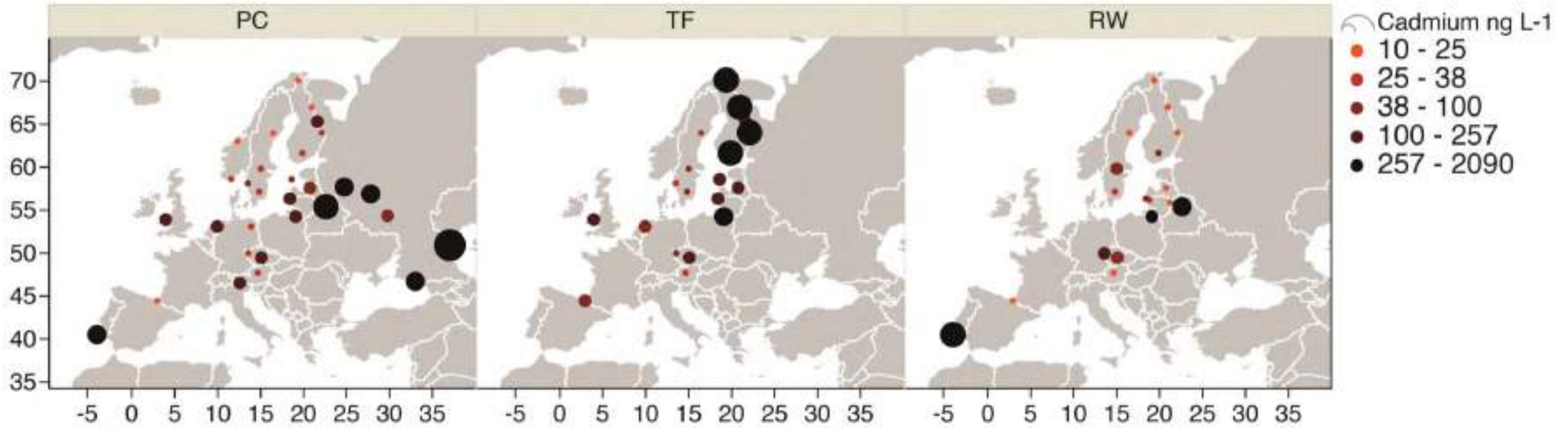
Hg Enrichment	mean	SD	n sites
RW/TF	0.40	0.40	4
TF/PC	160	349	5

Mercury Soil chemistry

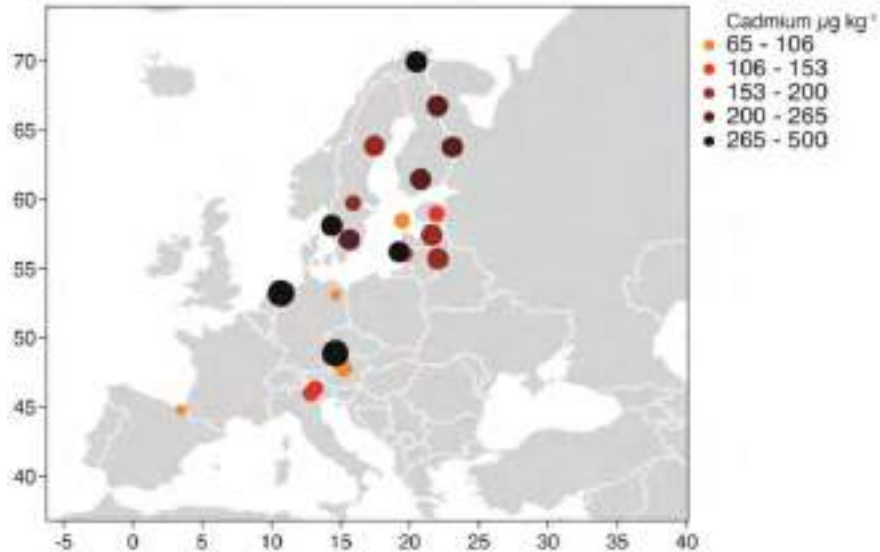


Mercury	Depth (cm)	Mean	median	min	max	n sites
BD ng L ⁻¹		1.23	0.25	0.01	4.6	13
TF ng L ⁻¹		85.3	33.5	18.8	250	6
LF mg kg ⁻¹		0.12	0.07	0.05	0.37	6
RW ng L ⁻¹		2.90	2.21	0.1	7.4	9
SC mg kg ⁻¹	0 – 10	0.05	0.04	0.005	0.15	8
	10 – 30	0.05	0.04	0.005	0.15	11
	30 – 200	0.03	0.01	0.005	0.12	10
	OH/litter	0.23	0.23	0.06	0.47	12

Cadmium



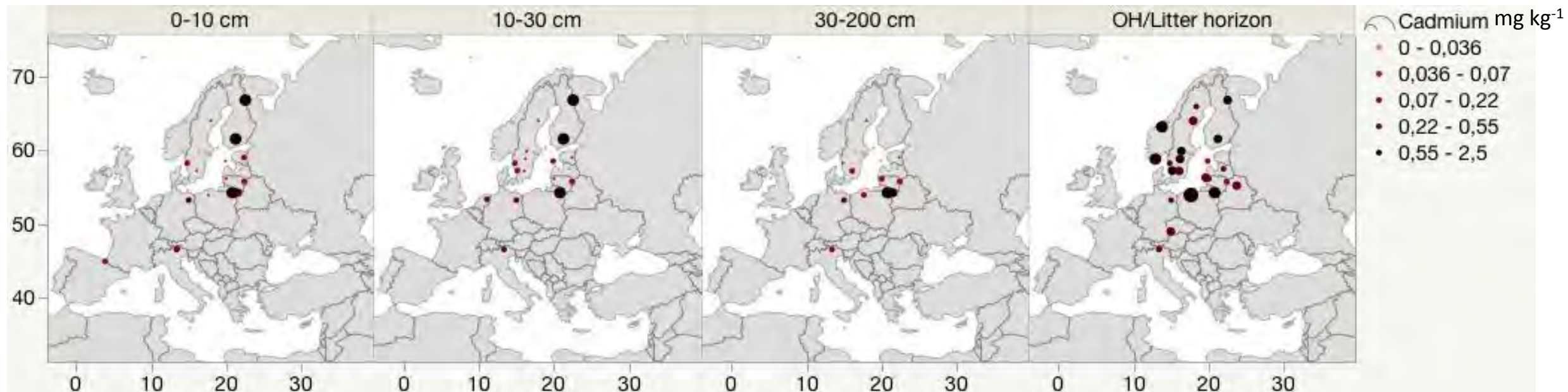
Litterfall



Cd Enrichment	mean	SD	n sites
RW/TF	0.64	1.13	14
TF/PC	9.9	17.4	19

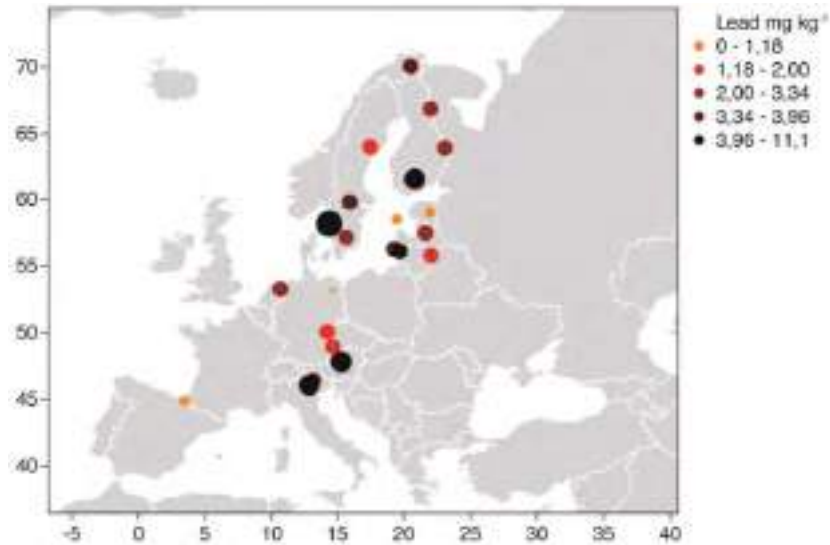
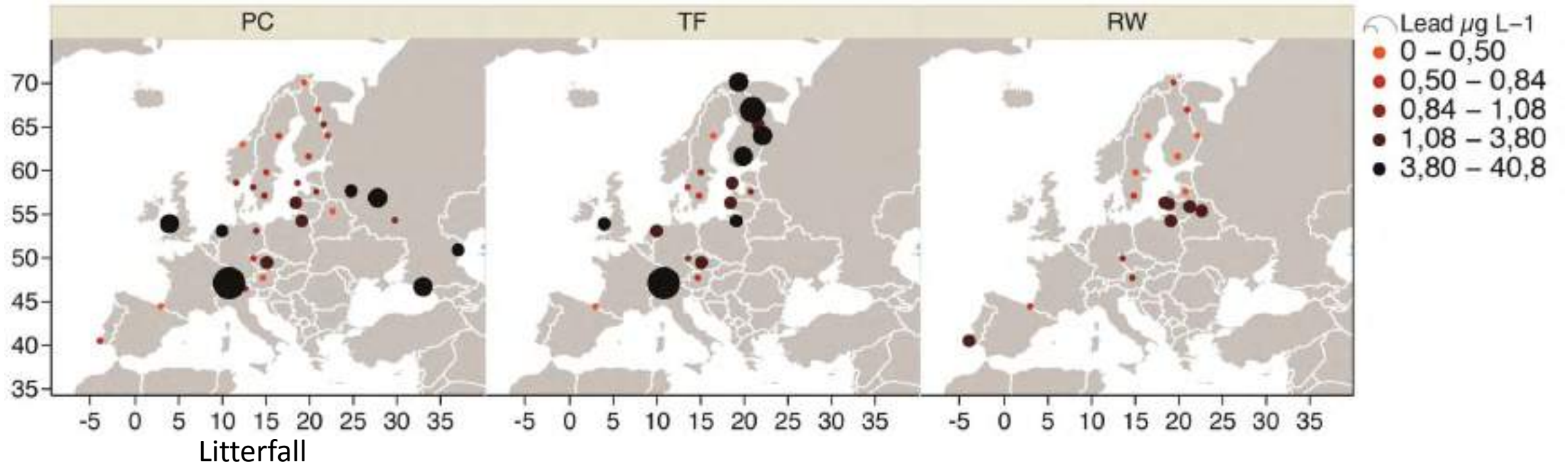
Cadmium

Soil chemistry



Cadmium	Depth (cm)	Mean	median	min	max	n sites
BD ng L ⁻¹		0.24	0.06	0.01	2.09	30
TF ng L ⁻¹		0.32	0.15	0.03	1.10	19
LF mg kg ⁻¹		0.21	0.20	0.07	0.5	21
RW ng L ⁻¹		0.12	0.03	0.01	0.85	19
SC mg kg ⁻¹	0 – 10	0.30	0.082	0.01	1.54	19
	10 – 30	0.25	0.07	0.01	1.55	21
	30 – 200	0.19	0,07	0.015	1.44	17
	OH/litter	0.63	0.44	0.11	2.5	23

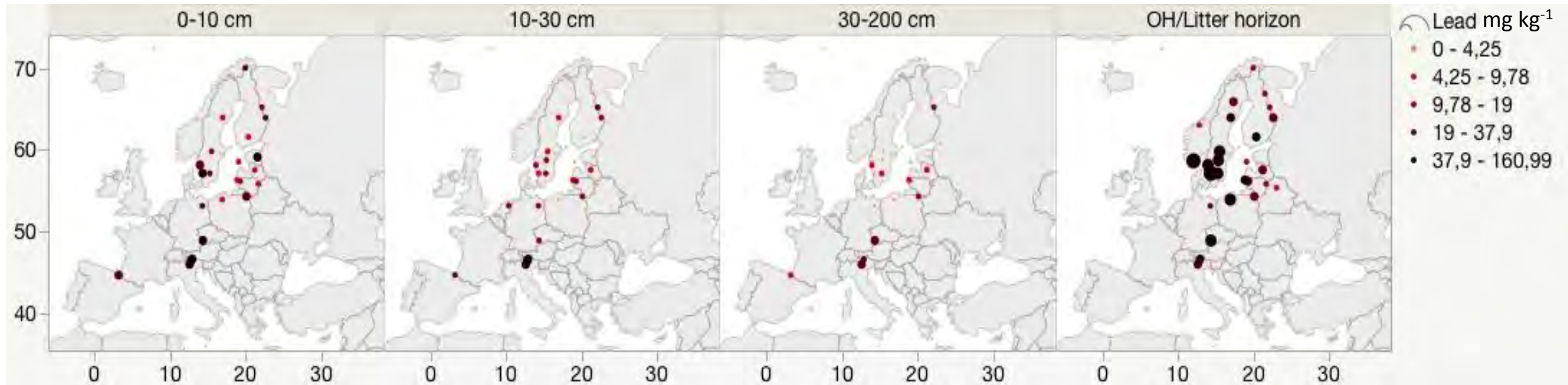
Lead



Pb Enrichment	mean	SD	n sites
RW/TF	0.57	0.67	13
TF/PC	5.7	9.9	20

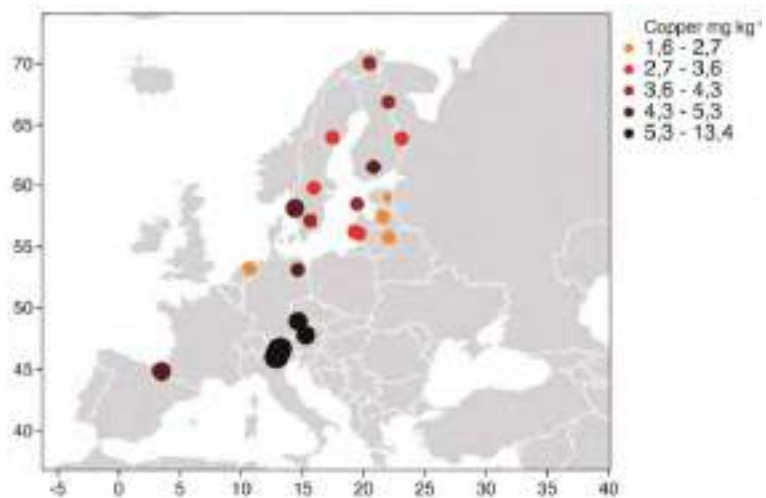
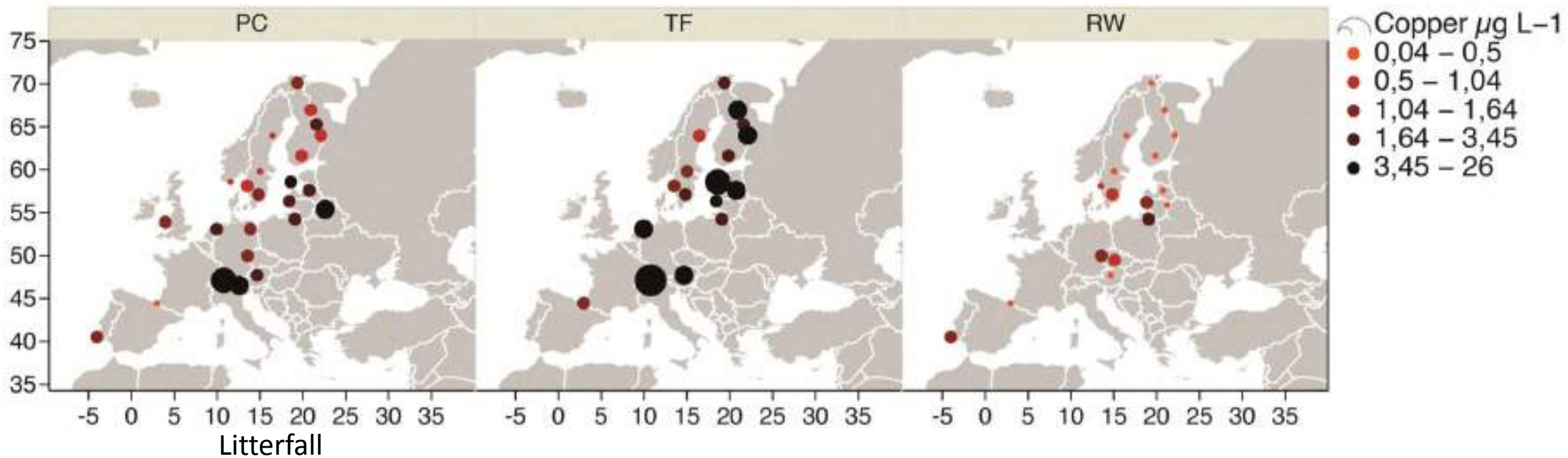
Lead

Soil chemistry



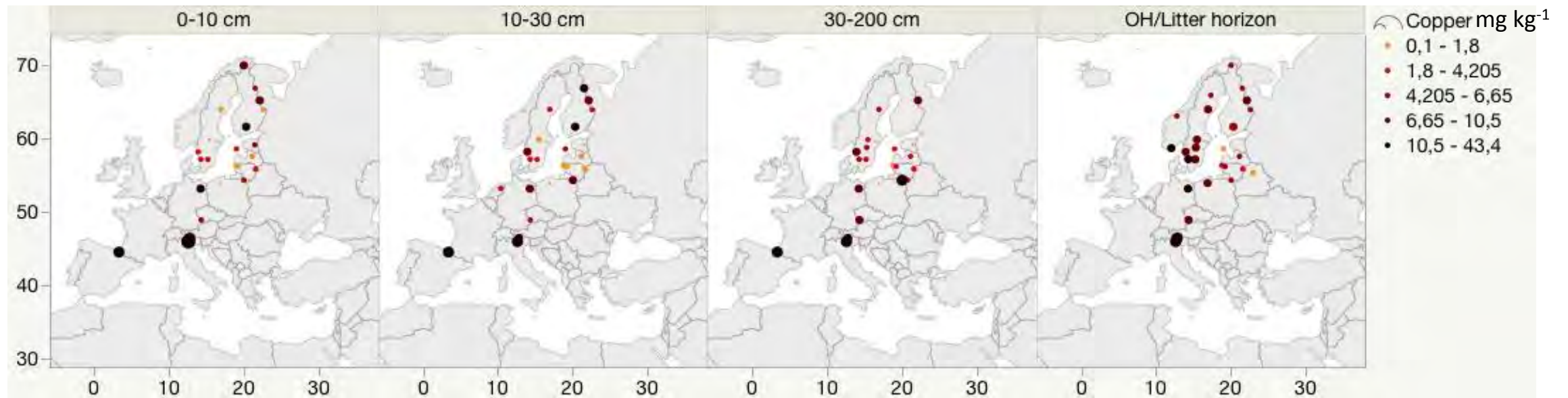
Lead	Depth (cm)	Mean	median	min	max	n sites
BD ng L ⁻¹		3.42	1	0.01	40	31
TF ng L ⁻¹		6.12	1.86	0.25	40.8	20
LF mg kg ⁻¹		3.38	3.31	0.83	11.1	21
RW ng L ⁻¹		0.68	0.49	0.04	3	18
SC mg kg ⁻¹	0 – 10	21.0	17.6	0.5	59	26
	10 – 30	13.7	11.2	1.0	46	25
	30 – 200	9.3	4.8	1.0	34	23
	OH/litter	45.7	33.6	5.6	161	28

Copper



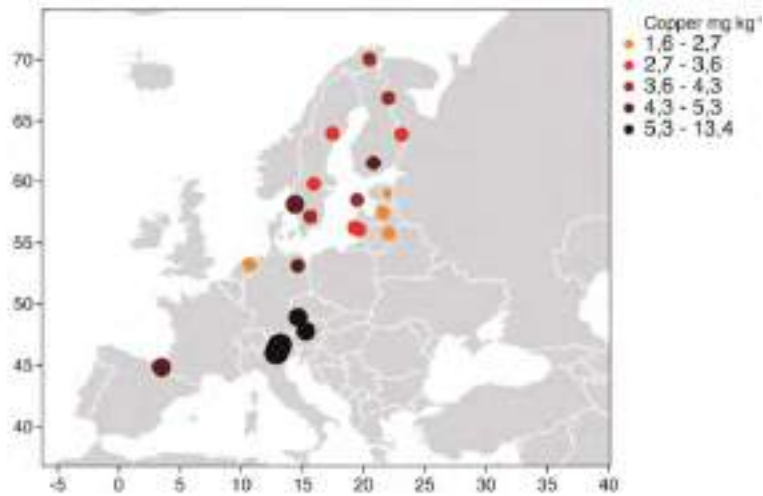
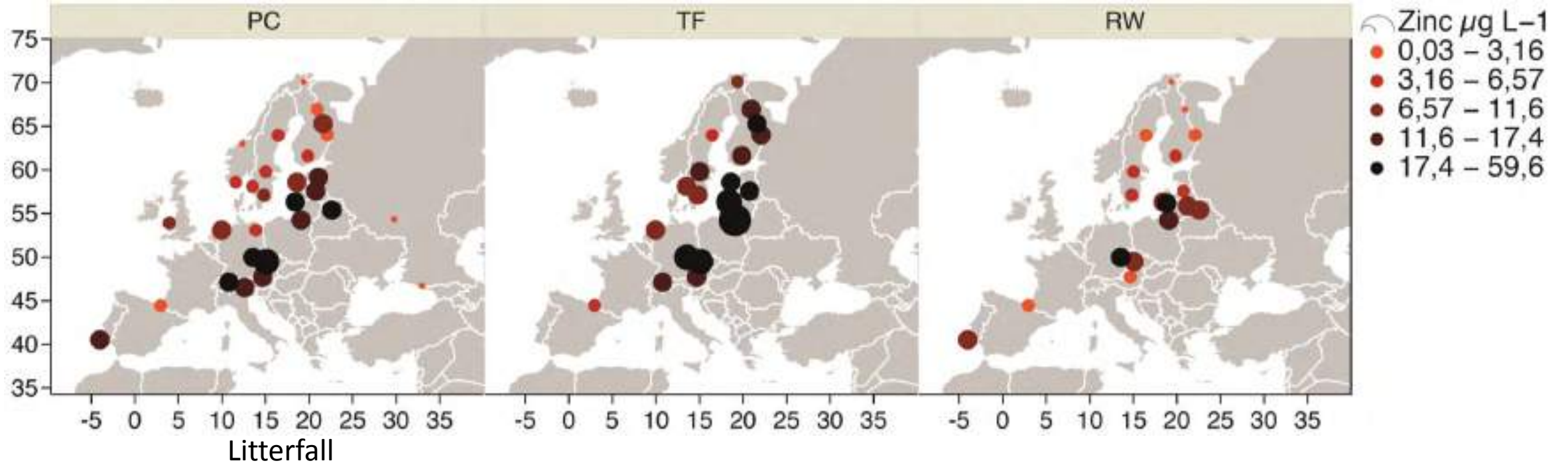
Cu Enrichment	mean	SD	n sites
RW/TF	0.24	0.15	13
TF/PC	2.7	1.7	17

Copper Soil chemistry



Copper	Depth (cm)	Mean	median	min	max	n sites
BD ng L ⁻¹		2.10	1.34	0.40	11.7	24
TF ng L ⁻¹		5.14	3.45	0.95	26.0	17
LF mg kg ⁻¹		5.02	4.27	1.6	13.4	23
RW ng L ⁻¹		1.13	0.66	0.1	3.5	19
SC mg kg ⁻¹	0 – 10	7.1	4.5	0.4	43.4	26
	10 – 30	6.9	4.5	0.1	19.4	25
	30 – 200	6.6	4.4	0.2	33.7	23
	OH/litter	8.2	7.0	1.5	23	28

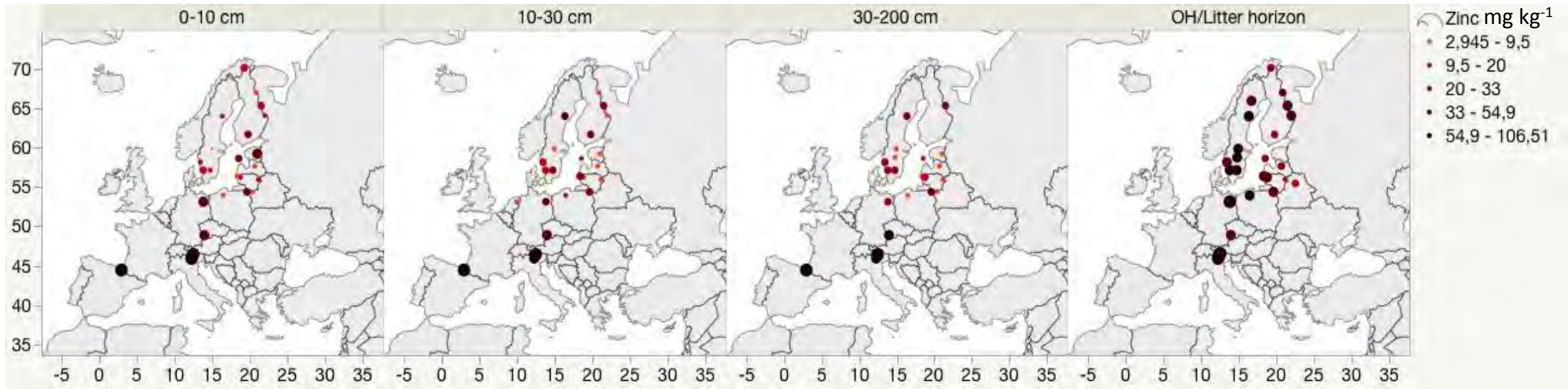
Zinc



Zn Enrichment	mean	SD	n sites
RW/TF	0.29	0.21	14
TF/PC	2.67	1.92	19

Zinc

Soil chemistry



Zinc	Depth (cm)	Mean	median	min	max	n sites
BD ng L ⁻¹		9.52	7.25	0.03	29.5	29
TF ng L ⁻¹		19.2	16.3	5.6	60	19
LF mg kg ⁻¹		56.0	53.2	21.6	132	24
RW ng L ⁻¹		7.2	5.4	0.62	21	19
SC mg kg ⁻¹	0 – 10	27.8	15.8	2.9	102	26
	10 – 30	26.8	20.5	4.0	93	25
	30 – 200	30.0	25	3.0	106	23
	OH/litter	48.9	46	13.2	101	26

Heavy metal concentrations in terrestrial compartments and runoff across European IM sites



Conclusions

- HM data from ICP IM sites important for evaluating responses from emissions and atmospheric transport of HM
- Uncertainty in data comparison between sites and regions
- Better resolution in spatial coverage of HM data; complement of data from compartments
- Others HM: Ni, Cr, Mo, V... (Sweden)
- More to come in ICP IM progress report autumn 2017...

Heavy metal concentrations in terrestrial compartments and runoff across European IM sites



What about POPs?



What about POPs?

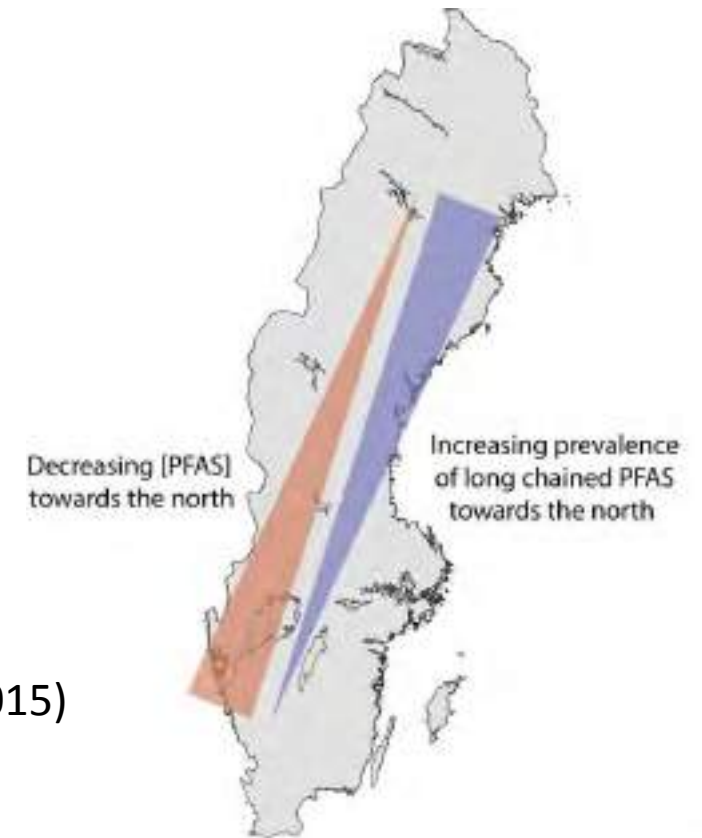
Variation and accumulation patterns of poly- and perfluoroalkyl substances (PFAS) in European perch (*Perca fluviatilis*) across a gradient of pristine Swedish lakes

Åkerblom S. et al. Accepted for publication in *Science of the Total Environment*
(Not ICP IM or ICP Waters, but related....)

- 7 PFASs detected: 6:2 FTSA, PFOS, PFDA, PFUnDA, PFDoDA, PFTriDA, and PFTeDA

—————→
Increasing length of carbon chain

- Σ PFAS : mean \pm SD: 0.99 ± 0.63 ng g⁻¹ ww (min / max : 0.31 / 3.4 ng g⁻¹ ww)
- Comparable background Σ PFAS levels from high mountain lakes in France ranging from 20.7 to 36.1 ng g⁻¹ ww (Ahrens et al. 2010)
- PFOS in Sweden: 0.01 / 0.93 ng g⁻¹ ww'
- PFOS in Resolute Bay (Nunavut, Canada): <0.001 – 2.0 ng g⁻¹ ww (Lescord et al. 2015)





Environmental drivers of leaf litter decomposition in streams

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University of Applied Sciences and Arts
of Southern Switzerland

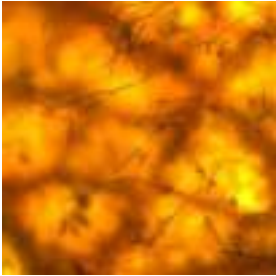
SUPSI

¹Institute of Earth Sciences

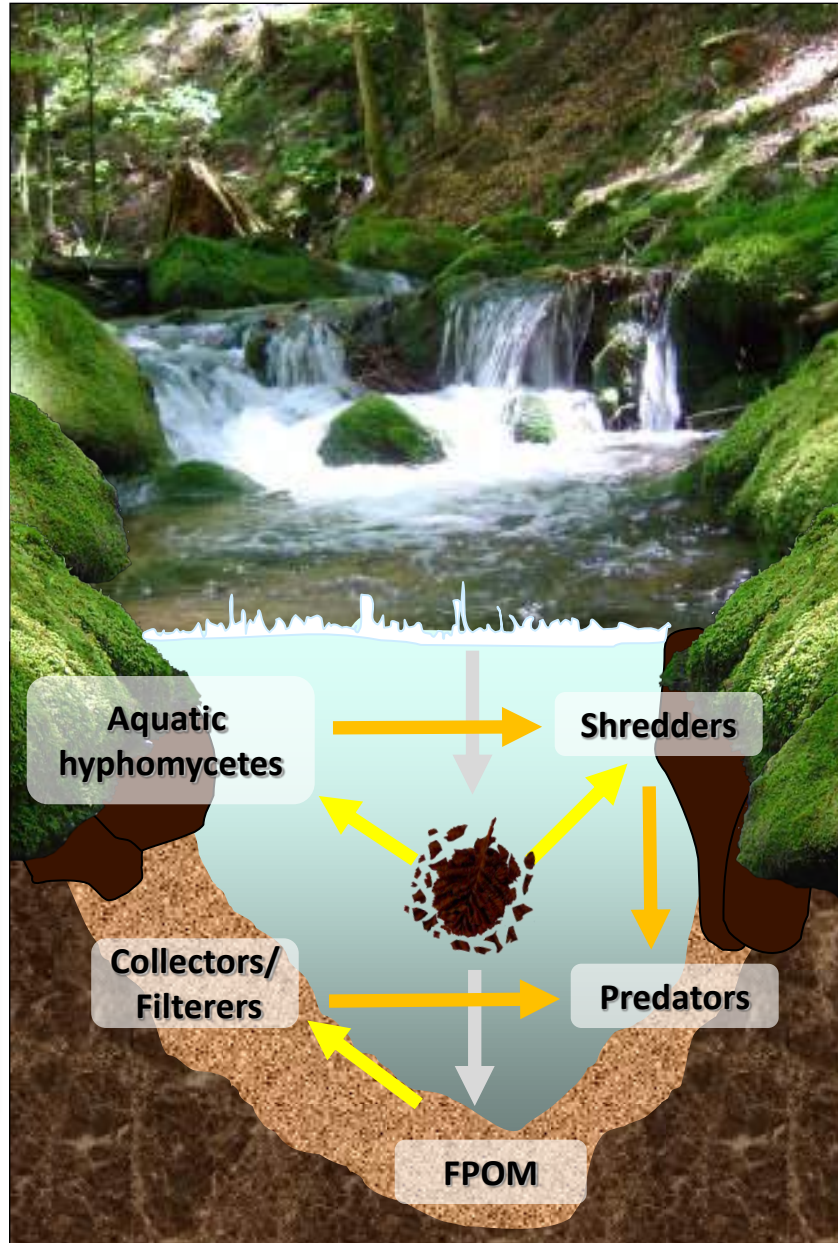
²Laboratory of Applied
Microbiology

food webs of forested streams

Aquatic hyphomycetes



Collectors/Filterers



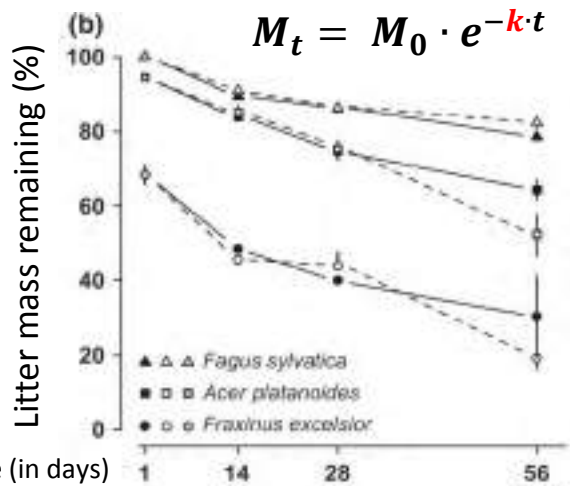
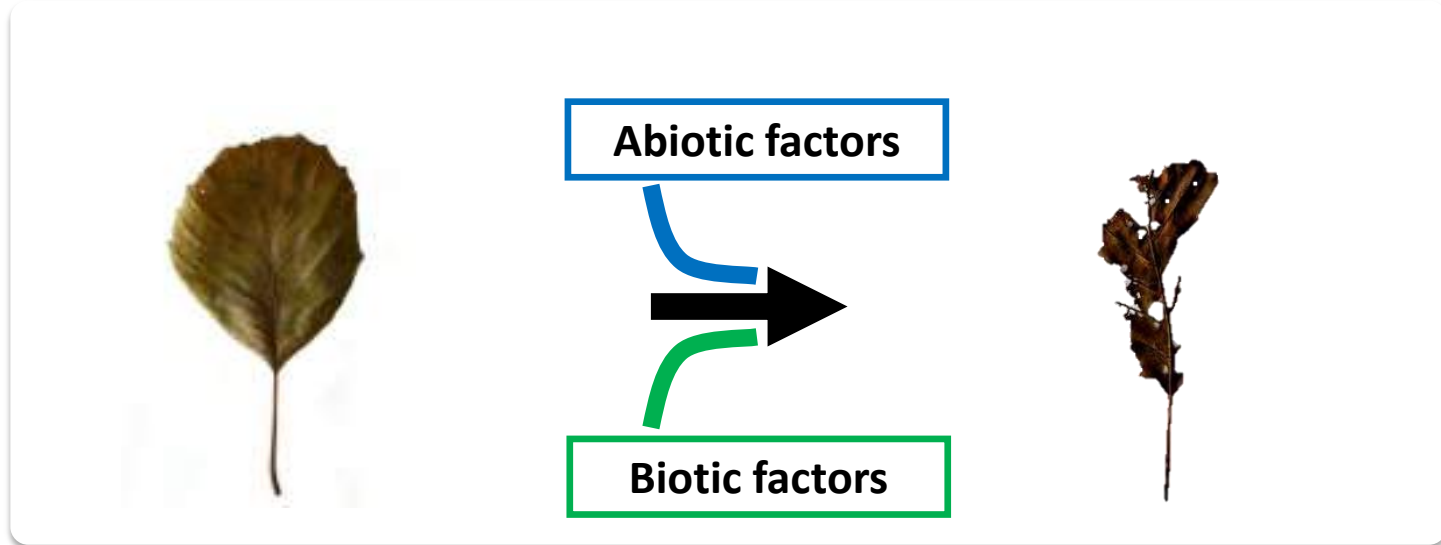
Shredders



Predators



leaf litter decomposition



Time (in days)

Bruder et al. (2014) Freshwater Biology

Abiotic factors

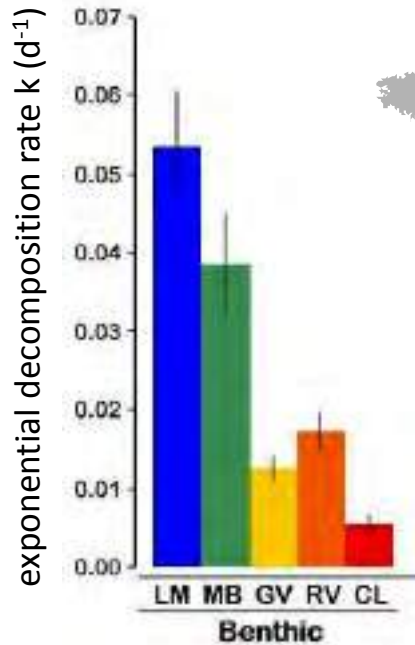
- > Temperature
- > Dissolved chemicals
- > Pollution
- > Flow velocity
- > ...

Biotic factors

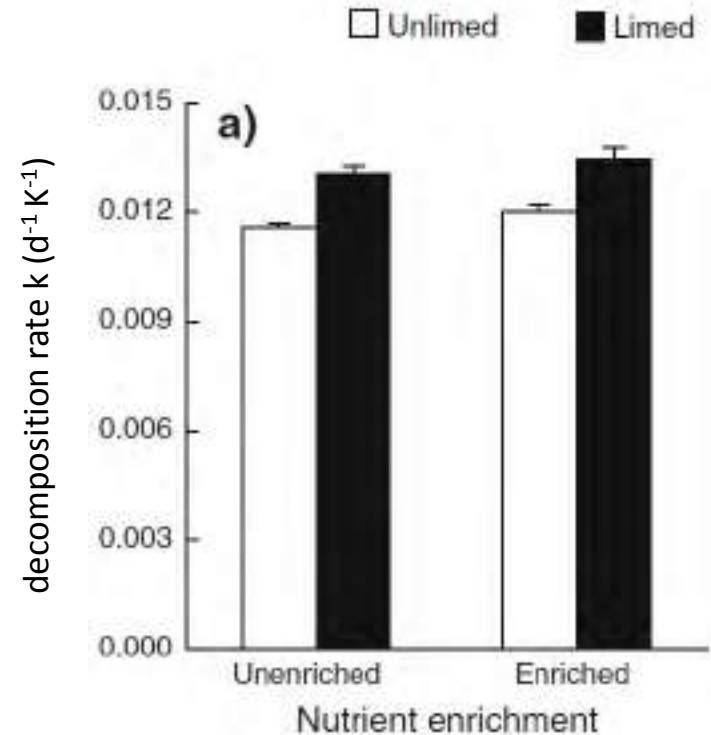
- > Litter characteristics
- > Biodiversity
- > Community composition
- > ...

role of abiotic conditions

acidification



Stream	LM	MB	GV	RV	CL
pH	7.4	7.0	6.4	6.1	4.6

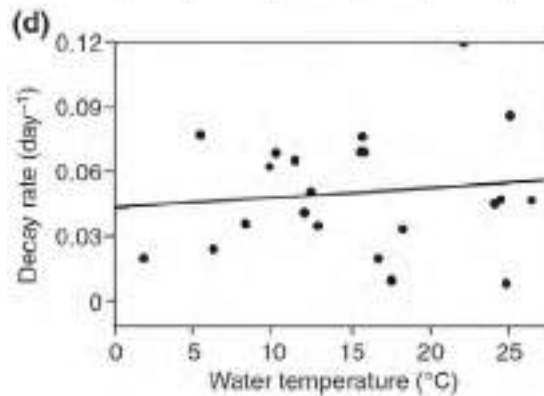
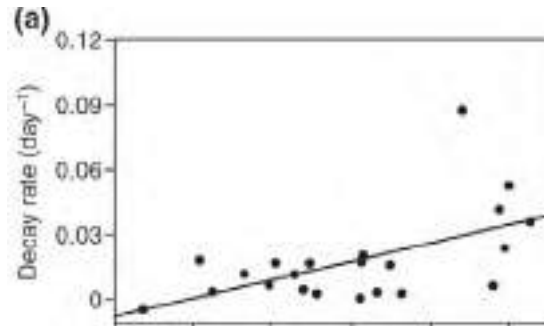
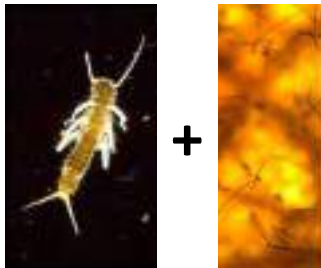
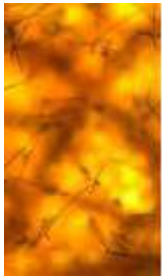
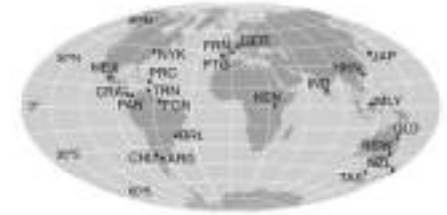


McKie et al. (2009) Oecologia

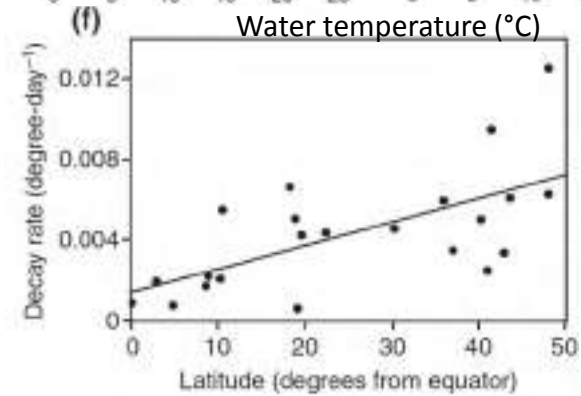
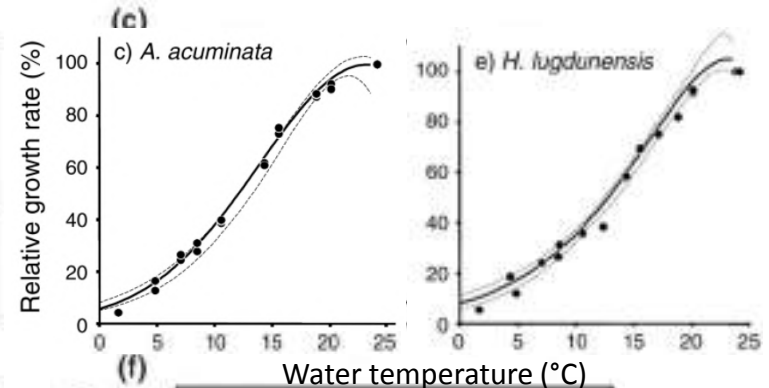
**In this study, effects of acidification on fungal and shredder biomass was small
> but negative effects on their activity and in turn on decomposition**

role of abiotic conditions

temperature and latitudinal gradient
(and decomposer groups)



Boyero et al. (2011) Ecology Letters



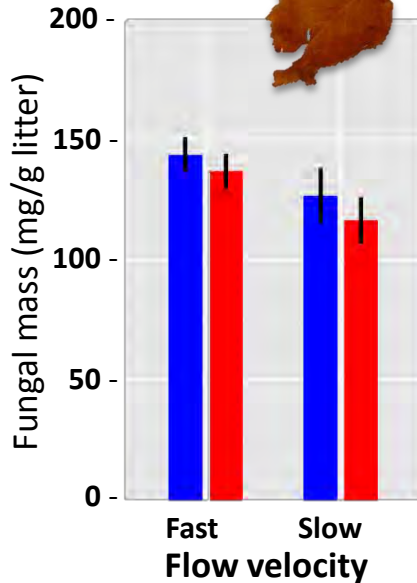
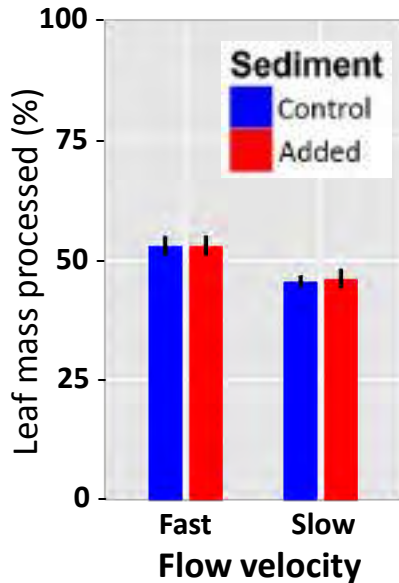
Dang et al. (2009) Ecology

Activity of fungi and shredders respond differently to temperature and latitude
> climate warming could change their relative importance for litter decomposition and affect the carbon cycle

role of abiotic conditions

flow velocity and fine sediment

Example from multiple stressor studies



Bruder et al. (2016) Functional Ecology



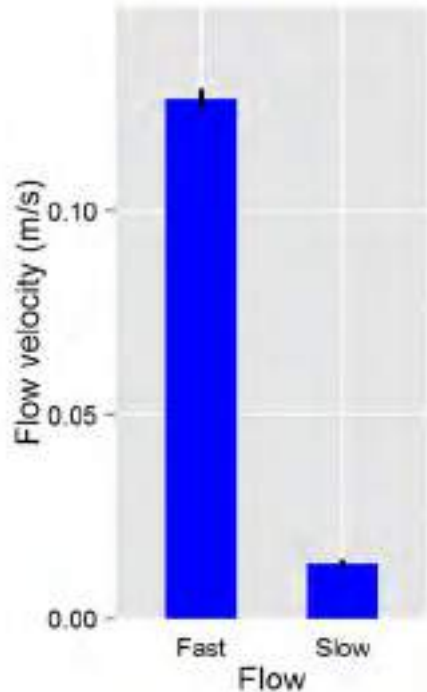
Outdoor/streamfed
N: 128
Area: 0.045 m²
Discharge: 2 L·min⁻¹
Volume: 3 L
Sediment: 0.5 L
Res. Time: 75s

Slow flow velocity reduces litter decomposition rate and fungal biomass
> situations comparable to littoral zones of forested lakes?

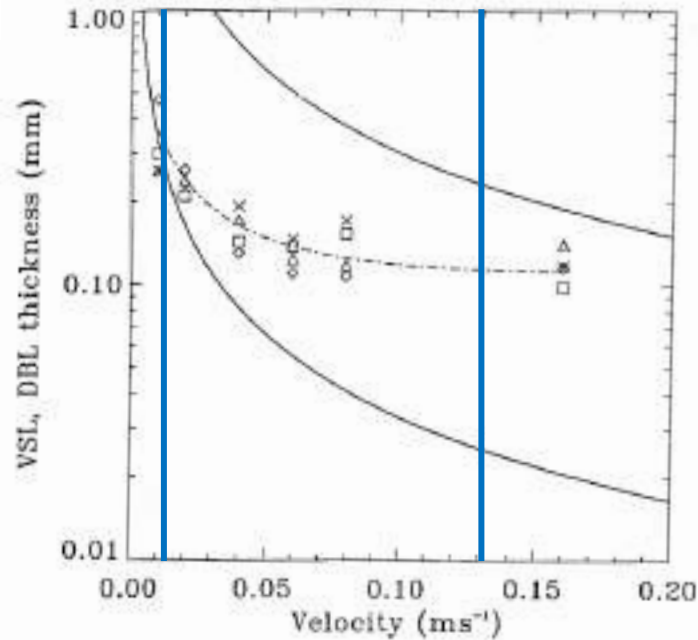
role of abiotic conditions

flow velocity and fine sediment

Measured flow velocity

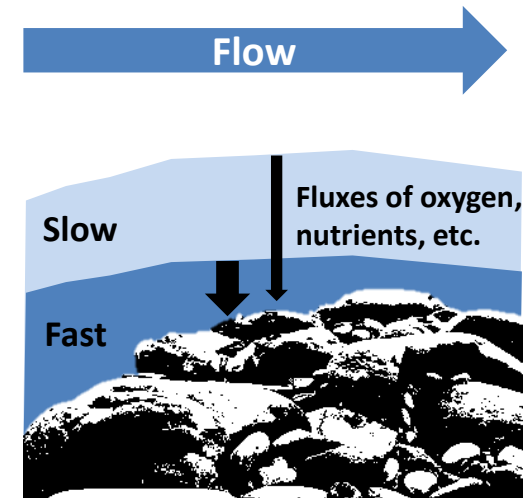


Flow velocity effects on boundary layer



mod. from Stevens & Hurd (1997) Hydrobiologia

Boundary layer effects on benthic habitat

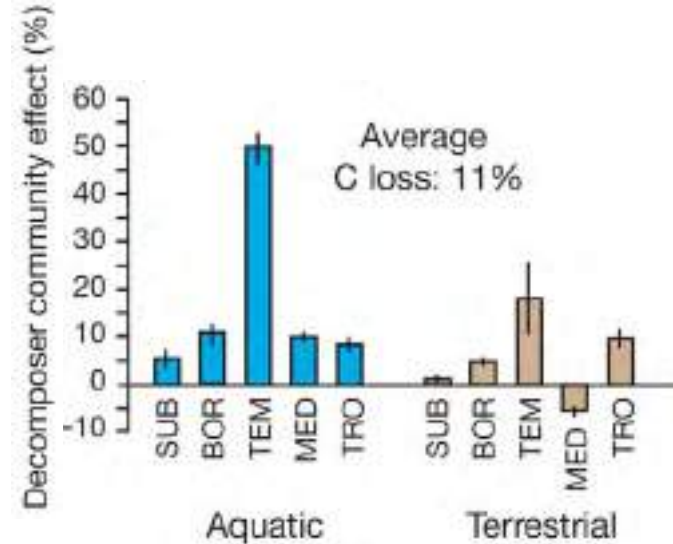
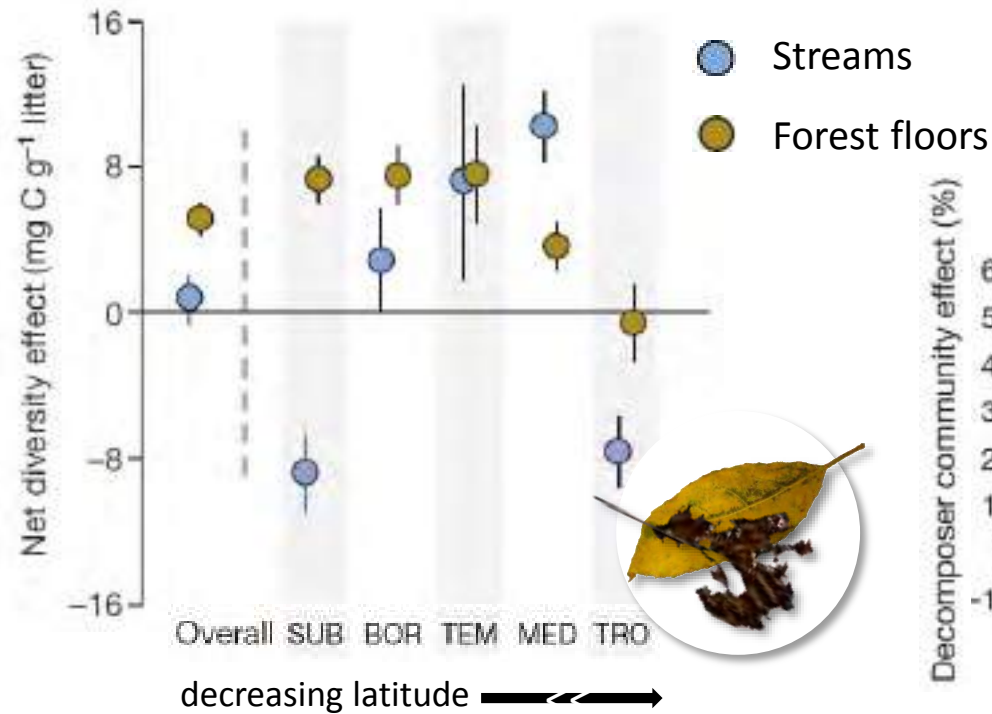


Fluxes = f (boundary layer thickness) = f (flow velocity)

Effects on physicochemical conditions in/on the sediment with consequences on organisms and biological processes

role of biodiversity

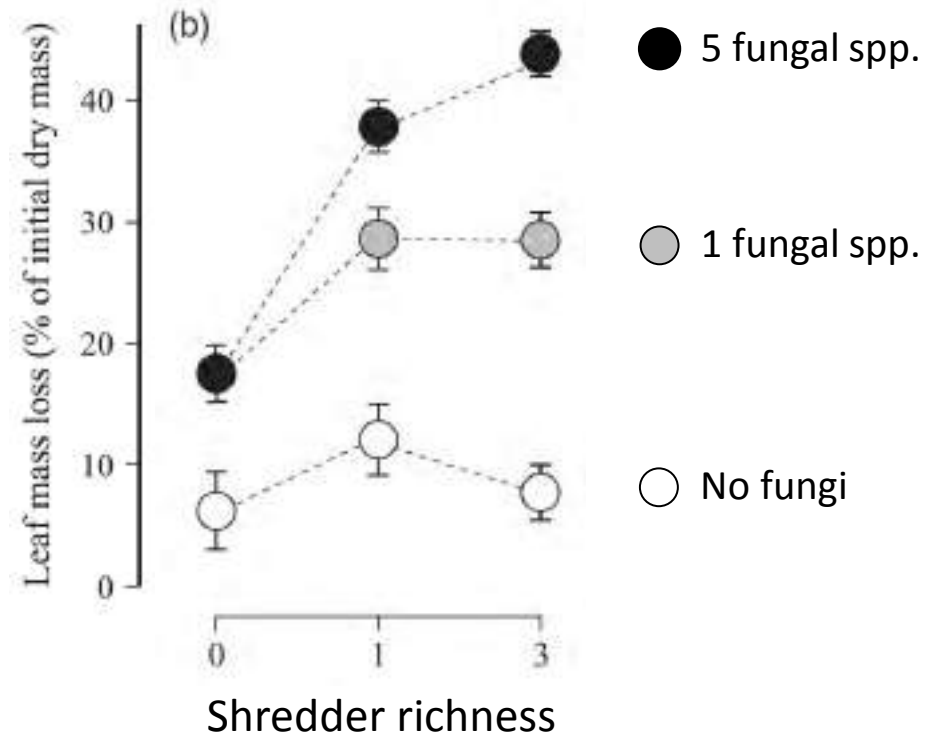
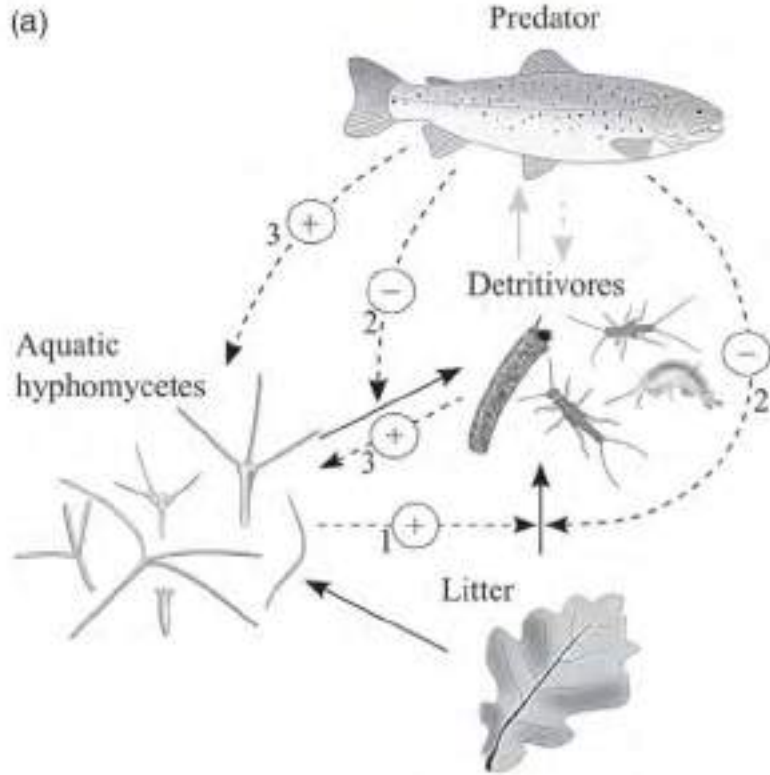
leaf litter diversity (and biome effects)



Species identity and environmental conditions dominate diversity effects

role of biodiversity

Decomposer diversity (and food-web complexity)



Both vertical and horizontal diversity influences litter decomposition rates

conclusions

Litter decomposition is sensitive to environmental conditions but:

- > response depends on specific ecological and environmental situation
- > can be different for the main decomposer groups (fungi vs shredders)
- > depends on biodiversity *but* the picture is not very clear for aquatic situation

Part of ecological variation could be reduced using standardized substrate:

- > for instance cotton strips (mainly cellulose)
- > this approach is used in the ICP IM indicator *Microbial Decomposition* but also in other global studies (e.g. CELLDEX – testing decomposition in streams and riparian vegetation)
- > results are transferable between studies using leaf litter and those using cotton

Thank you for your interest

and the Swiss Federal Office for the Environment (FOEN) for financial support to participate at the ICP IM TF Meeting



Species sensitivity to acidification in highly endemic regions of South Africa



Londiwe Khuzwayo, Chris Curtis, Helen James and Kari Austnes



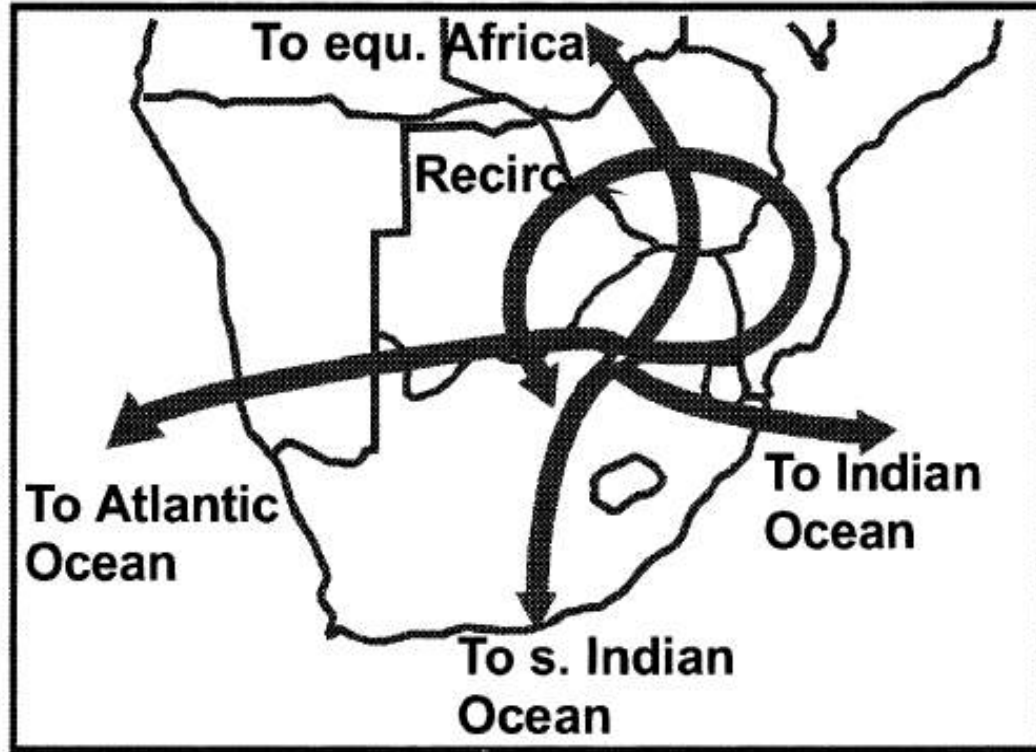
School of Geography, Archaeology and Environmental studies

Introduction

- South Africa is a major consumer of coal and is among the twenty most carbon intensive economies in the world
- About 74% of the countries energy demand is met by coal resources
- Pressure on the energy sector over the past 18 years
 - Increase energy demands
 - economic growth
 - Free Basic Energy Policy (1998)
- Two new coal power stations (4 800 MW)
- 16 coal power stations and 75% of these are concentrated in the Highveld (HV) region of South Africa



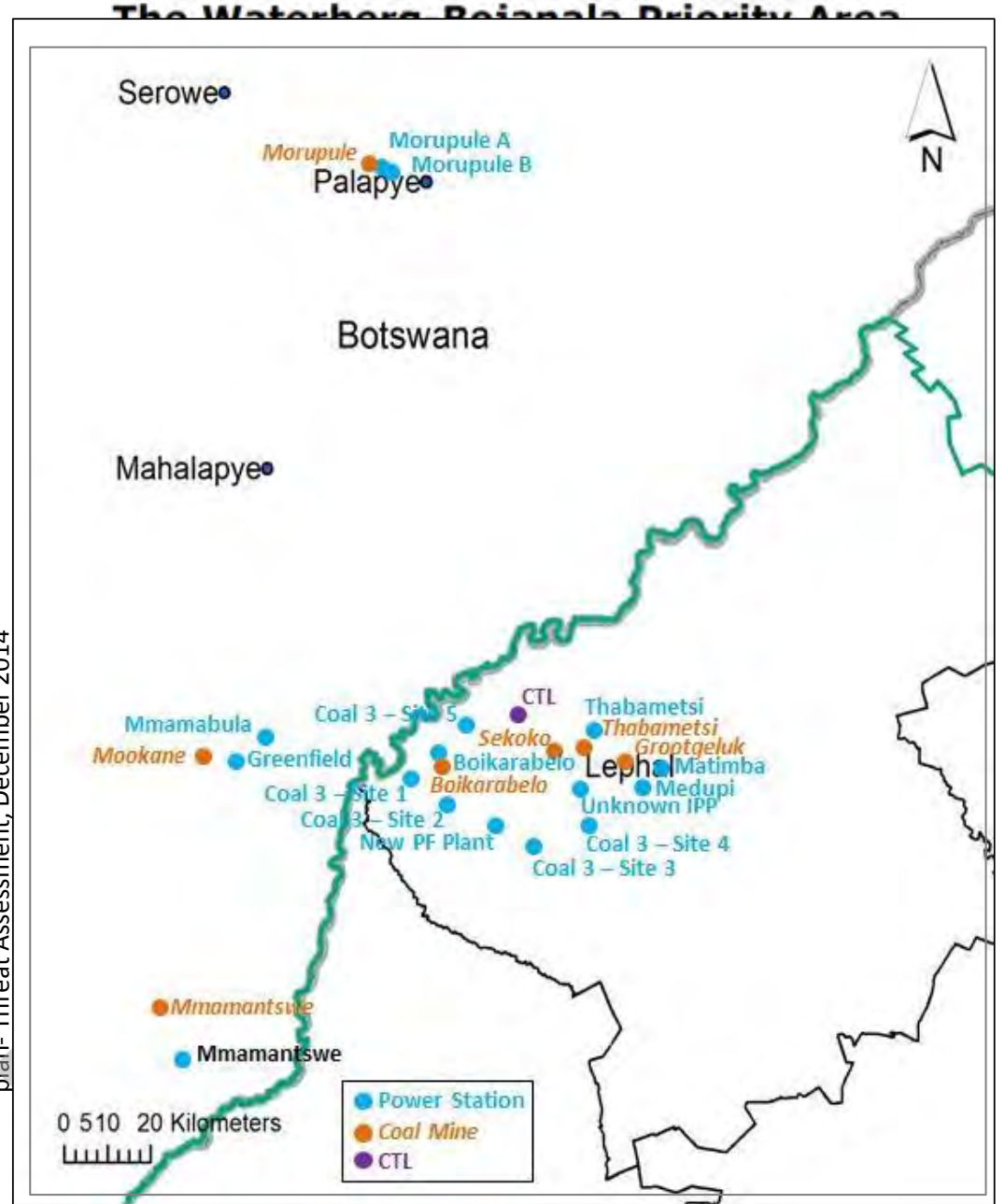
Main pathways transporting air in and out of the Highveld



Abbreviations: South (s.), Equatorial (equ.) and Recirculated (Recirc.)

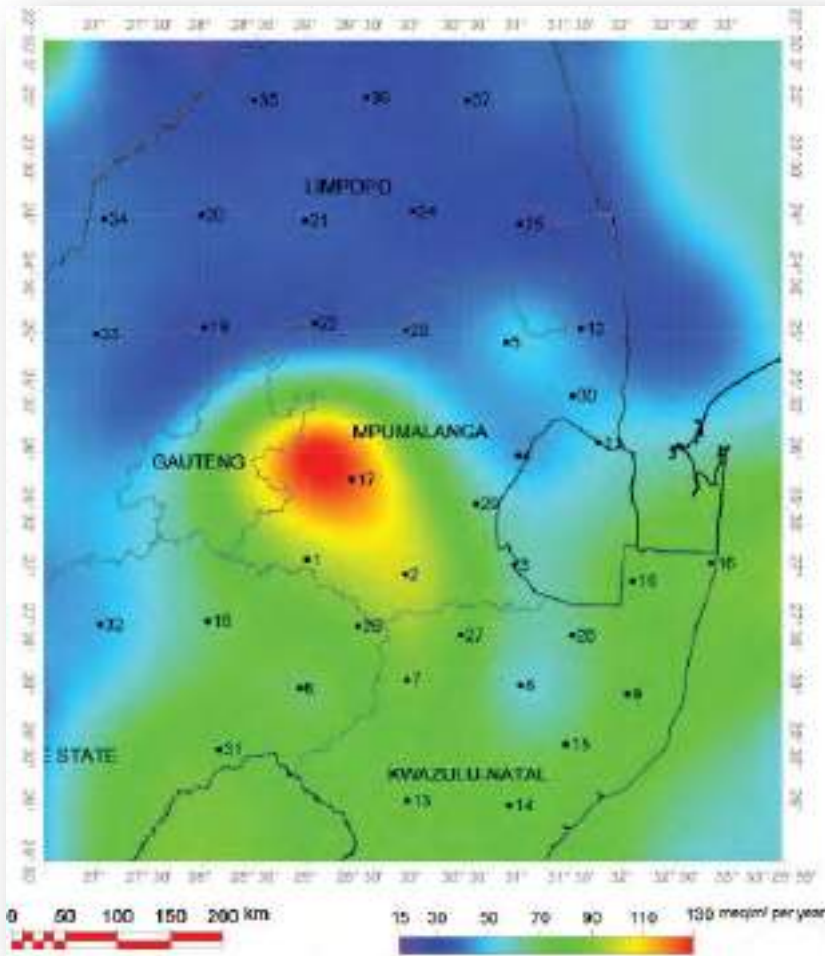
Source: Freiman, M. T. and Piketh, S.J. (2003) Air transportation into and out of the industrial Highveld Region of South Africa. American Meteorological Society 42.

Source: DEA, 2014. Waterberg-Bojanala Priority Area Air Quality Management plan- Threat Assessment, December 2014



Relative locations of the proposed energy-based and mining projects are indicated in blue and brown respectively while the proposed location for the coal-to-liquid (CTL) plant is shown in purple.

Total dry and wet acid deposition rate (meq/m²/year)



Source: Josipovic, M., Annegarn, H.J., Kneen, M.A., Pienaar, J.J. and Piketh, S.J. (2011) Atmospheric dry and wet deposition of sulphur and nitrogen species and assessment of critical loads of acidic deposition exceedance in South Africa. South African Journal of Science 107, 01-10.

Key Hypothesis

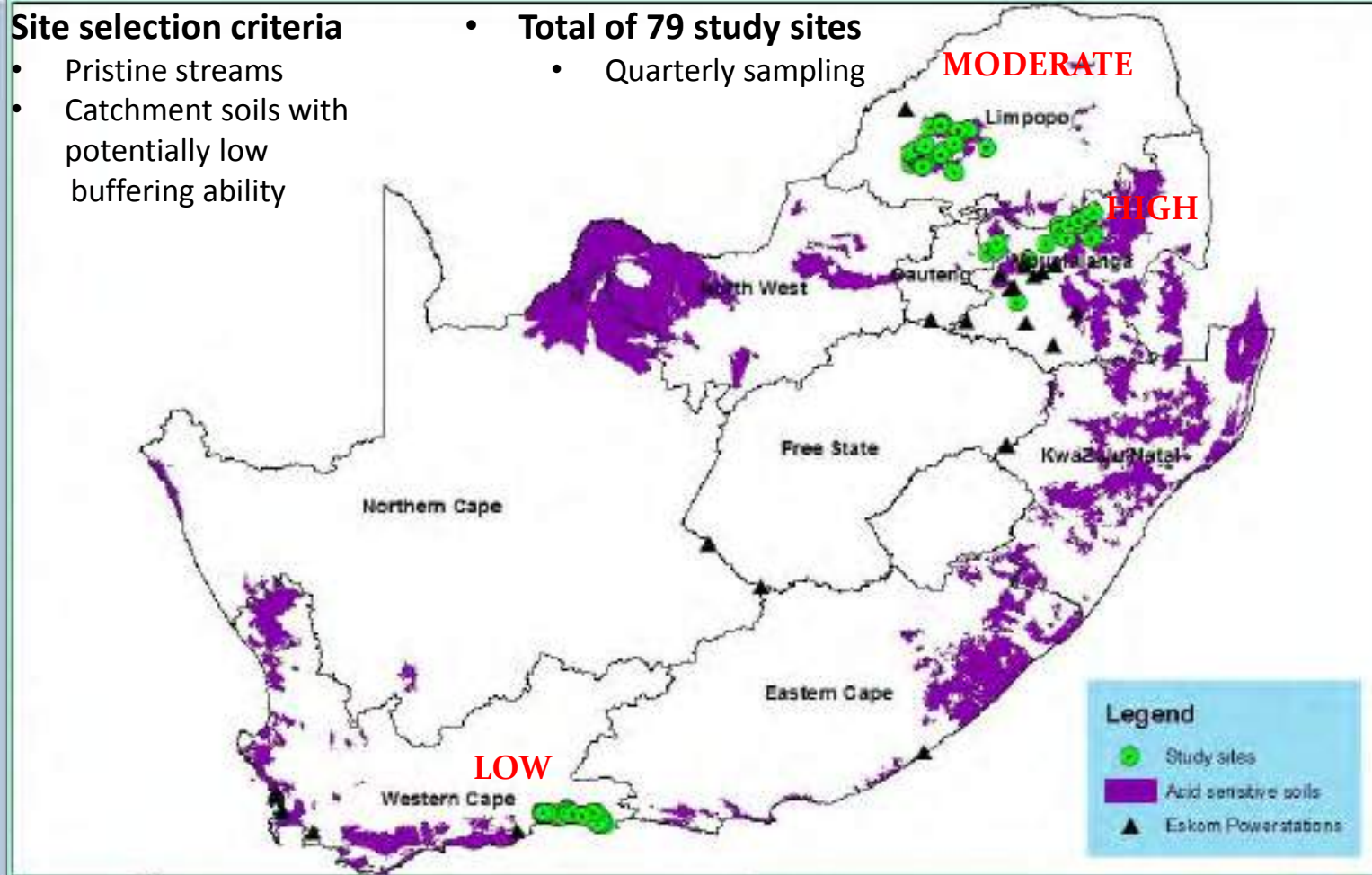
- Current levels of acid deposition exceed the critical load for acidification of acid sensitive streams
- Future point sources of S and N deposition pose a threat in acid loading to sensitive streams
- Anticipate climate change will exacerbate present-day acidification state

The aim of this study was to investigate differences in the aquatic ecosystem related to acidification by looking at water chemistry and macroinvertebrates.

Site selection criteria

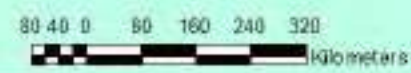
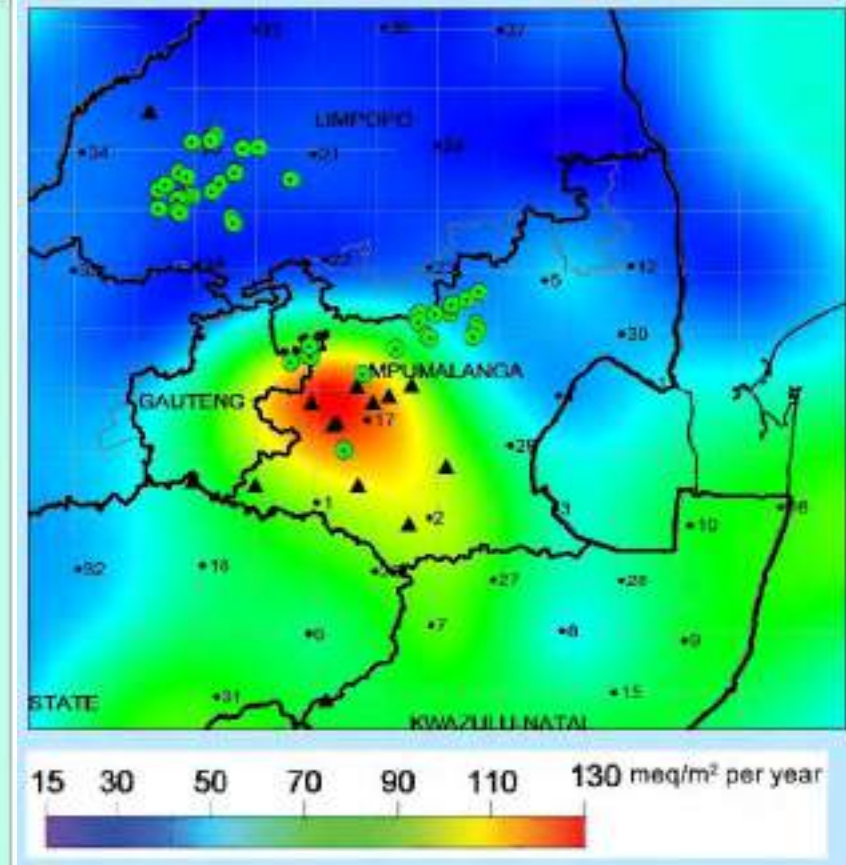
- Pristine streams
- Catchment soils with potentially low buffering ability

- Total of 79 study sites
 - Quarterly sampling



Legend

- Study sites
- Acid sensitive soils
- ▲ Eskom Power stations



Author: Londiwe Khuzwayo
 Date: June 2015
 Source: SOTER database
 Coordinate system: GCS_WGS_1984



South-western Cape
(SWC)



Waterberg (WB)



Highveld (HV)



Methods



- Biotope – Stoney riffle
- SASS5 net (1mm mesh with 30cm square frame)
- Kick sampling (2mins)

Preserved in 70% Ethanol

Sample emptied into a tray.
Individual species picked
Out and analysed under the
microscope

Key focus groups

- Order
 - Ephemeroptera (Mayflies)
 - Baetidae
 - Caenidae
 - Heptageniidae
 - Leptophlebiidae
 - Oligoneuriidae
 - Prosopistomatidae
 - Teloganodidae
 - Tricorythidae
 - Diptera (Trueflies)
 - Simuliidae (Blackflies)
 - Chironomids (Non-biting midges)
 - Plecoptera (Stoneflies) - One of the key indicator sp. for clean / non-polluted waters in South Africa



Trichoptera (Caddisflies)



Distribution of some Mayfly species across all three study regions

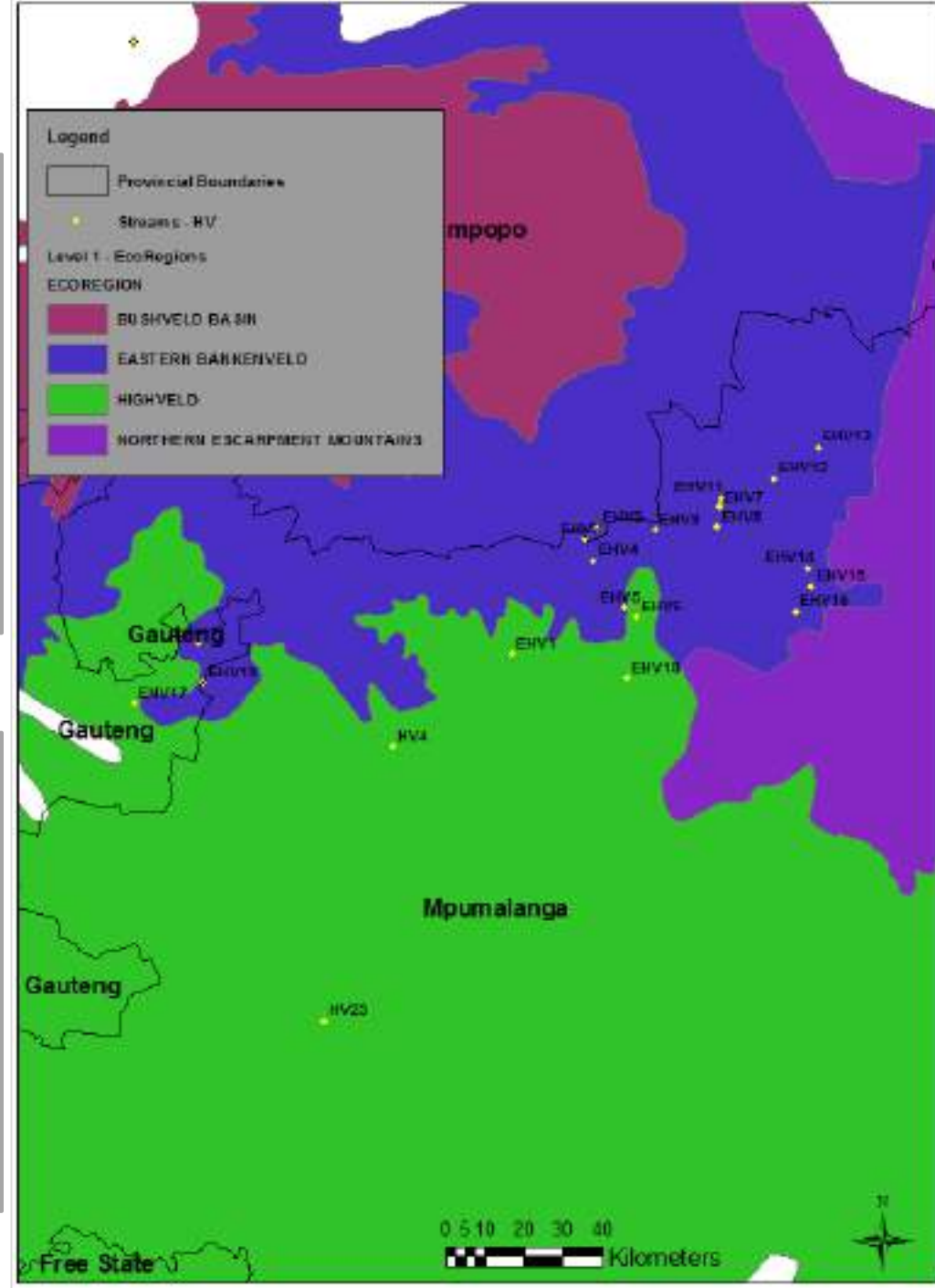
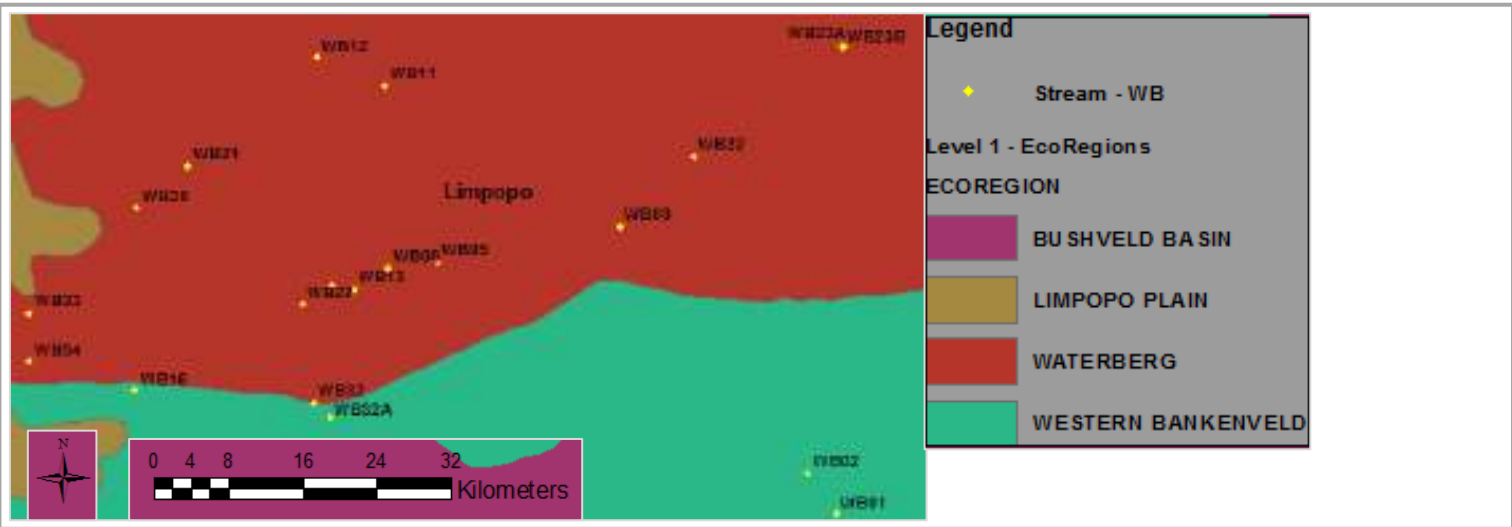
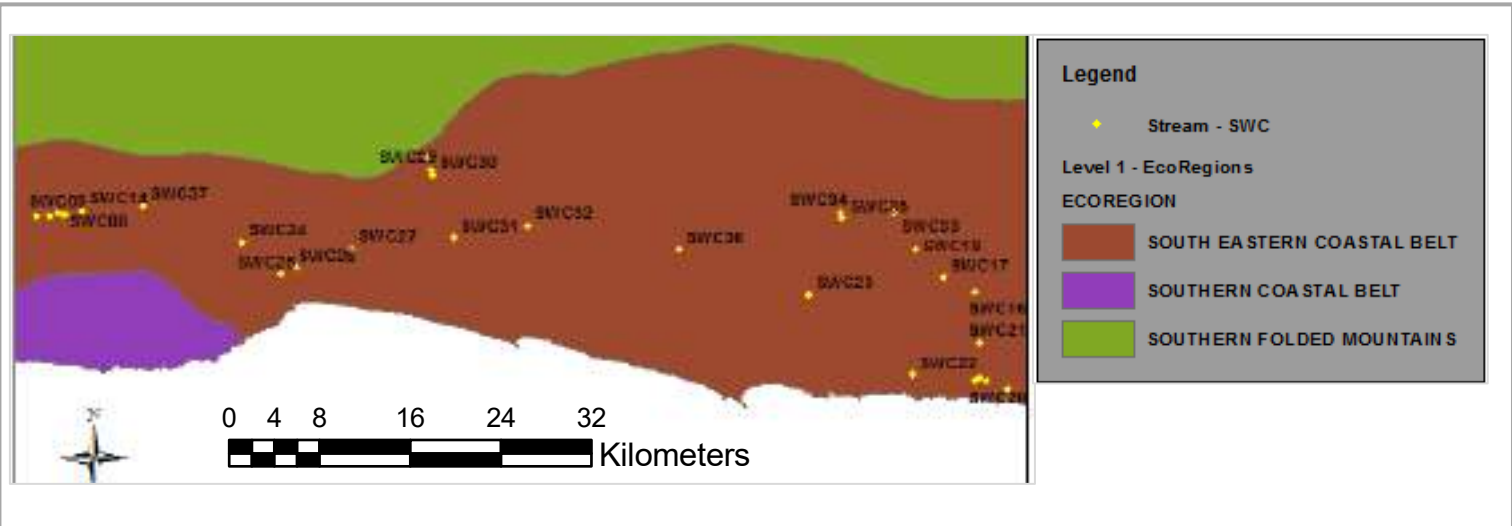
Order	Family	Genus/ species name	South-western Cape	Waterberg	Highveld
Ephemeroptera	Tricorythidae	<i>Tricorythus reticulatus</i>	Present	Absent	Present
	Leptophlebiidae	<i>Adenophlebia auriculata</i>	Absent	Absent	Absent
		<i>Adenophlebiodes bicolor</i>	Present	Absent	Absent
		<i>Aprionyx intermedius</i>	Absent	Absent	Absent
		<i>Castonophlebia calida</i>	Present	Absent	Absent
		<i>Choroerpes</i>	Absent	Absent	Absent
		<i>Euthraulus sp.</i>	Present	Absent	Present
		<i>Acanthiops varius</i>	Present	Absent	Absent
	Baetidae	<i>Afroptilum sudafricanum</i>	Absent	Absent	Absent
		<i>Baetis harrisoni</i>	Absent	Absent	Present
		<i>Baetis sp. (Unknown)</i>	Absent	Absent	Present
		<i>Bugilliesia margaretae</i>	Present	Absent	Present
		<i>Cheleocloeon excisum</i>	Absent	Absent	Absent
		<i>Cheleocloeon sp. (unknown)</i>	Absent	Absent	Absent
		<i>Cloeodes sp. (unknown)</i>	Absent	Present	Present
		<i>Cloeon sp. (unknown)</i>	Present	Absent	Absent
		<i>Crassabwa flava</i>	Absent	Absent	Absent
		<i>Dabulamanzia indusii</i>	Absent	Present	Present
		<i>Demoreptus capensis</i>	Absent	Absent	Present
		<i>Demoreptus monticola</i>	Absent	Present	Present
		<i>Demoreptus sp. (unknown)</i>	Present	Absent	Present
		<i>Demoulinia crassi</i>	Present	Absent	Present
		<i>Labiobaetis vinosus</i>	Absent	Present	Present
		<i>Pseudocloeon glaucum</i>	Absent	Absent	Absent
		<i>Pseudocloeon latum</i>	Absent	Absent	Absent
		<i>Pseudocloeon piscis</i>	Absent	Absent	Absent
		<i>Pseudocloeon sp. (unknown)</i>	Absent	Absent	Present
		<i>Pseudopannata maculosa</i>	Present	Absent	Absent
	<i>Susua sp. (unknown)</i>	Present	Present	Absent	

Present Found in 70% or more sites sampled in that region / season

Absent Not found in any of the sites sampled in that region

Sparse Found in 40% or less sites in that region / season

EcoRegions



Highveld region

- The Highveld region had the least number of overall sampling site and Macroinvertebrate sites
- High pH values
- Historical data (1954 -1965) obtained from Albany museum
 - 10 Simuliidae spp. were described during that time of which four were identified in this project
 - Ephemeroptera was largely dominated by Baetidae spp. While samples collected during this project were mainly dominated by *Tricorythus reticulatus* and *Euthraulus* spp.
 - *Neoperla* is still the only genus found in this region with just one described sp., *Neoperia spio*
- These observations suggest that spp. composition has changed over the years and there has been a huge decline in Baetidae spp. for this region, which may be accounted to changes in land-use
- The high pH values in the mist of well documented deterioration of air quality suggests that the soils in this region have a high buffering ability

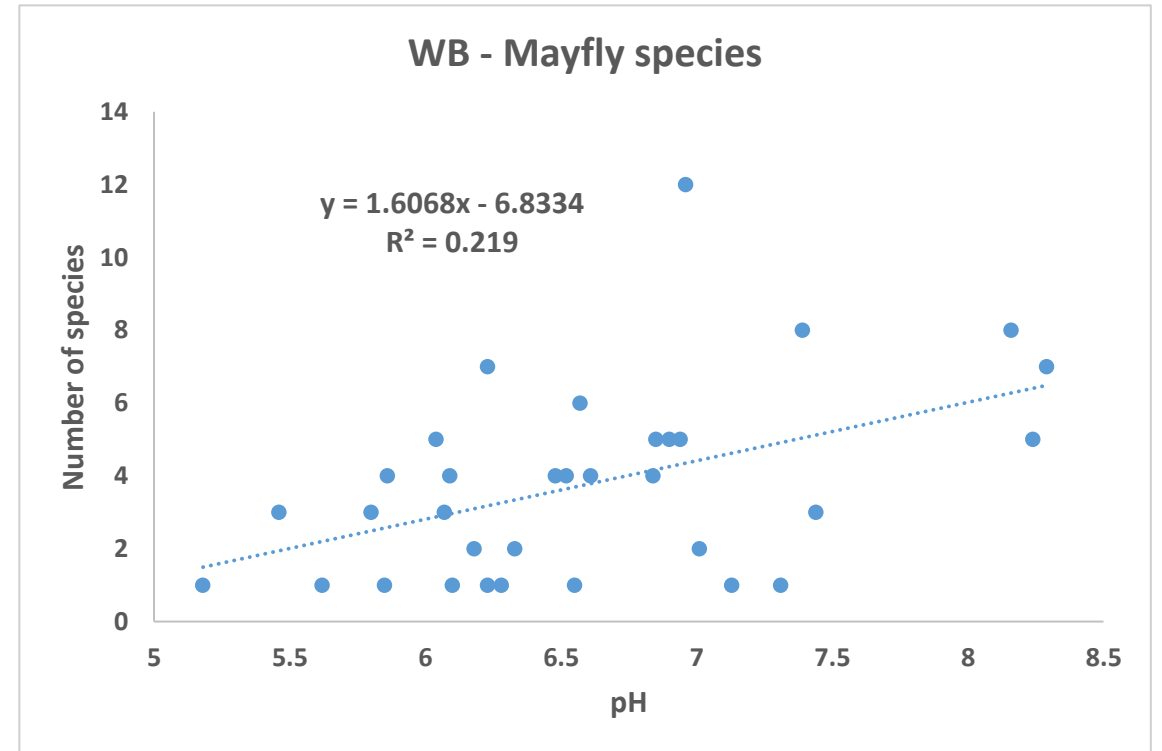
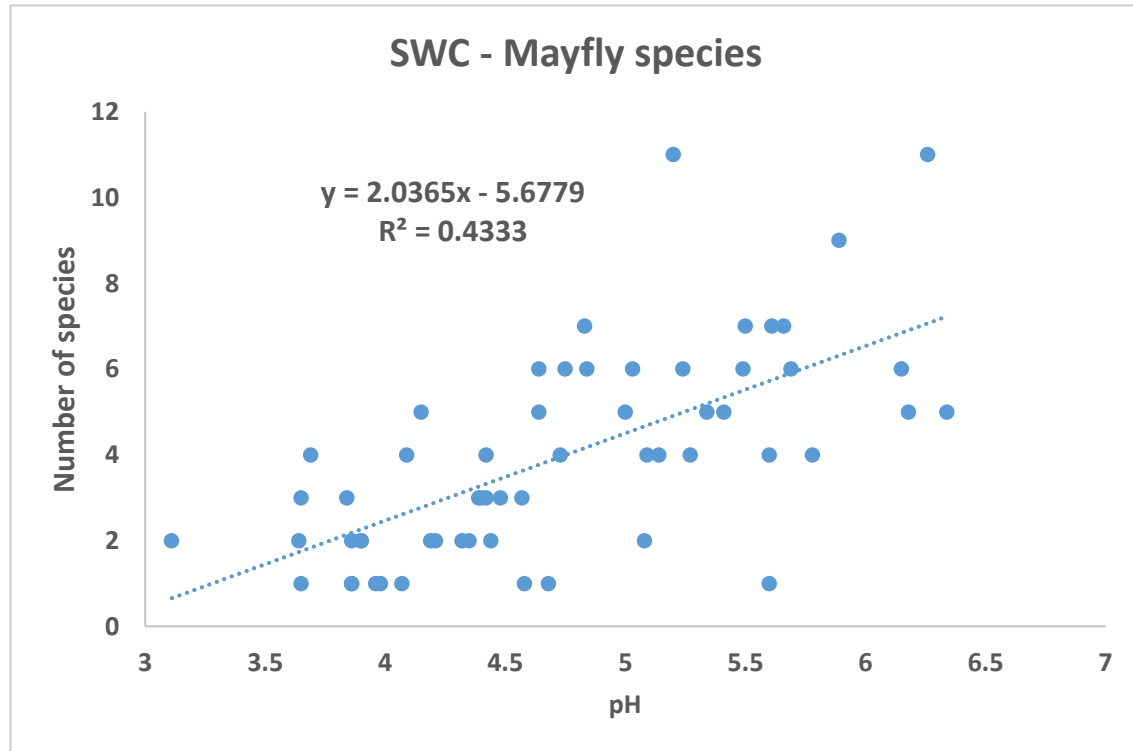
Waterberg region

- Least studied area of the three regions
- Consists of acidic and basic streams
- Largely dominated by uncommon Mayfly species that are sparsely distributed

South-western Cape region

- Most studied area in aquatic sciences
- Naturally acidic streams
- High degree of diversity and endemism
 - World's 200 significant Freshwater Ecoregions

The effects of pH concentration on species numbers



Visit to Norway

- Chemistry
 - Q & A
 - Summarising and analyzing data
- Relationship between chemistry and biology
 - Individual species
 - Potential indices
- Critical loads

THANK YOU



NORTH-WEST UNIVERSITY
YUNIBESITI YA BOKONE-BOPHIRIMA
NOORDWES-UNIVERSITEIT



The Research Council
of Norway



National
Research
Foundation



The overall project addresses the theme “Environment” with a specific interest in the effects of air pollution on the aquatic ecosystem as well as predicting future scenarios of climate change and changes in sulphur and nitrogen emissions into the atmosphere.

South African indices

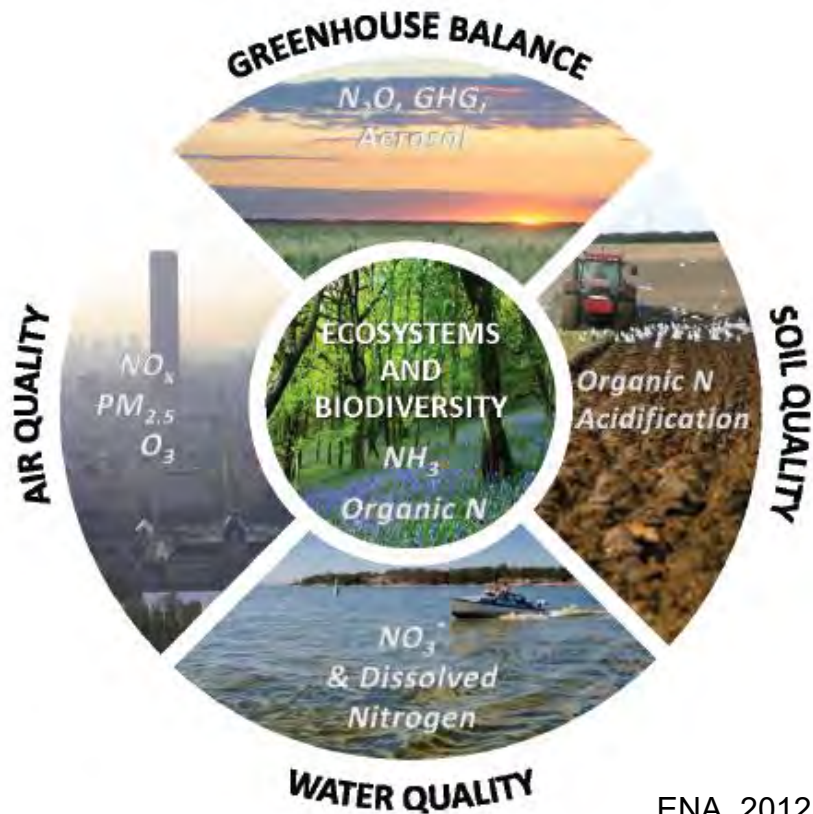
- There are several indices that are available in South Africa and are used to assess the integrity of the aquatic ecosystem
 - Index of habitat integrity (IHI)
 - Fish assemblage integrity index (FAII)
 - South African scoring system (SASS5)
 - Macro-invertebrate response assessment index (MIRAI)
 - Fish response assessment index (FRAI)
 - Riparian vegetation index (RVI)
- SASS5 and MIRAI are specific to macroinvertebrates and they assess the overall health of the aquatic ecosystem and identification is done at family level
- In this study we tried to investigate the feasibility of establishing a South African scoring system specific to acidification using macroinvertebrates



LONG-TERM NITROGEN DEPOSITION IMPACTS IN A TEMPERATE FOREST ECOSYSTEM IN AUSTRIA

F. Rokop

THOMAS DIRNBÖCK **IKA DJUKIC** JOHANNES KOBLER
MICHAEL MIRTL

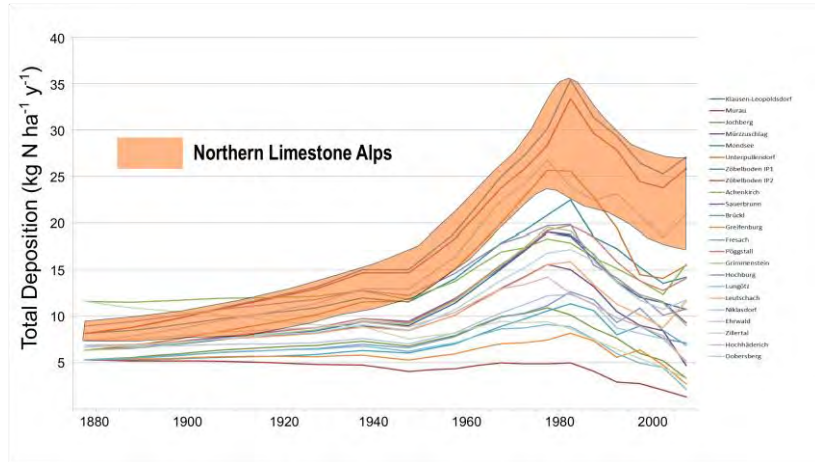


ENA, 2012

REACTIVE NITROGEN HARMS THE ENVIRONMENT

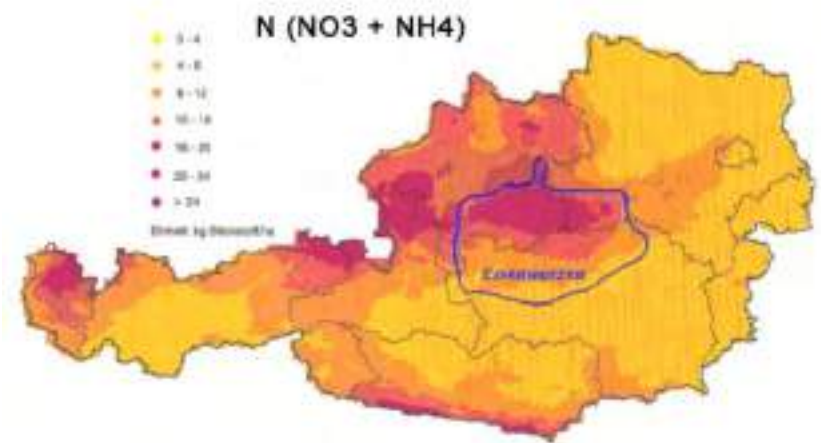
- **Nitrogen is an essential nutrient** but today emissions from agriculture and fossil fuel burning provide $\frac{3}{4}$ of the N input into the biosphere (in Europe)
- **Ecosystem effects** include tree nutritional imbalances, NO₃ loss to the groundwater, N₂O emissions, soil acidification, and biodiversity loss
- The annual costs of reactive N in the environment in the EU amount to 70 - 380 billion € (ENA, 2012)

HIGHEST ATMOSPHERIC NITROGEN LOADS IN THE NORTHERN LIMESTONE ALPS



Total (dry and wet) N deposition since 1880

3-5 fold the preindustrial N deposition

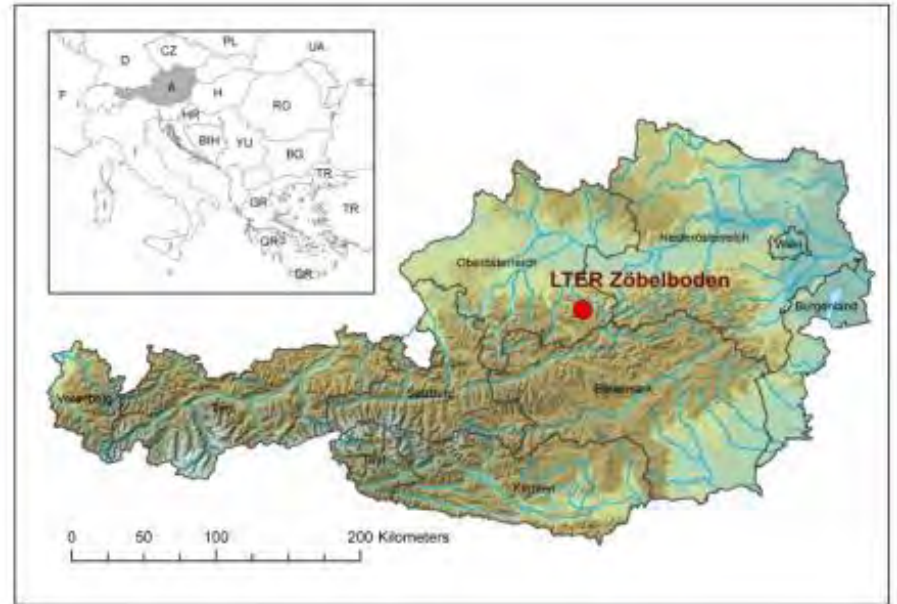


Distribution of wet N deposition in Austria (Schneider 1998)

Highest N deposition in the humid Northern Limestone Alps due to precipitation and emission sources

25 YEARS ECOSYSTEM MONITORING AT ZÖBELBODEN

- Forested, 90 ha Karst catchment (550 - 950 m a.s.l) in the Northern Limestone Alps
- Established in the year 1992 as the Austrian's contribution to ICP Integrated Monitoring (UN-ECE) of air pollution effects in Europe
- Today LTER Zöbelboden serves as a highly instrumented ecosystem monitoring and research site for the effects of air pollution and climate change including biodiversity

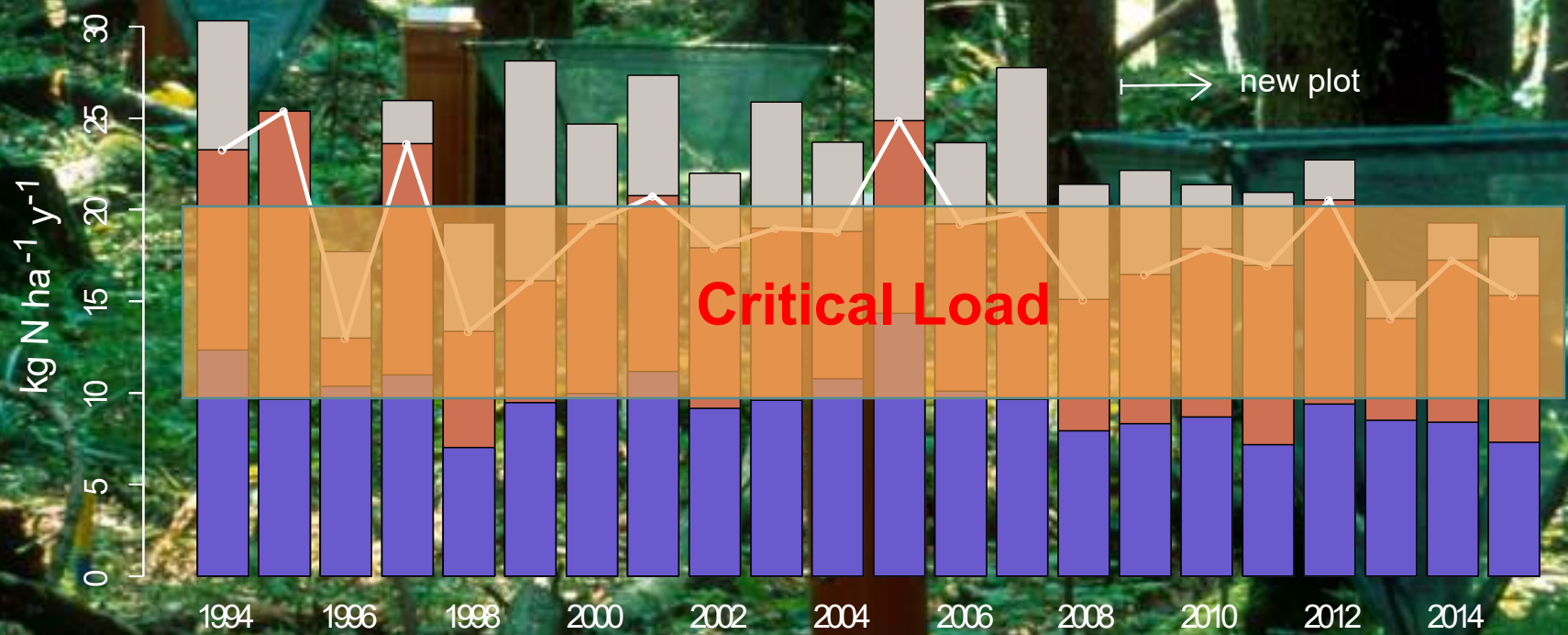


Norway Spruce forest

NO_3^- -N

NH_4^+ -N

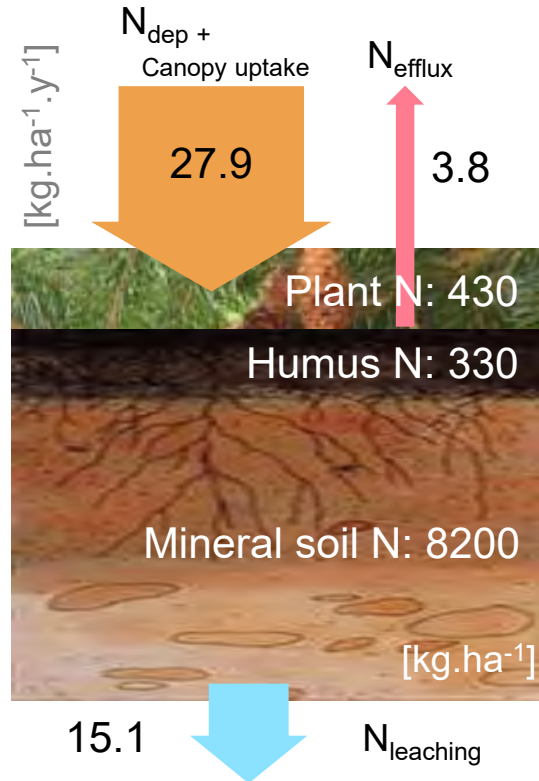
N_{org}



Canopy throughfall deposition

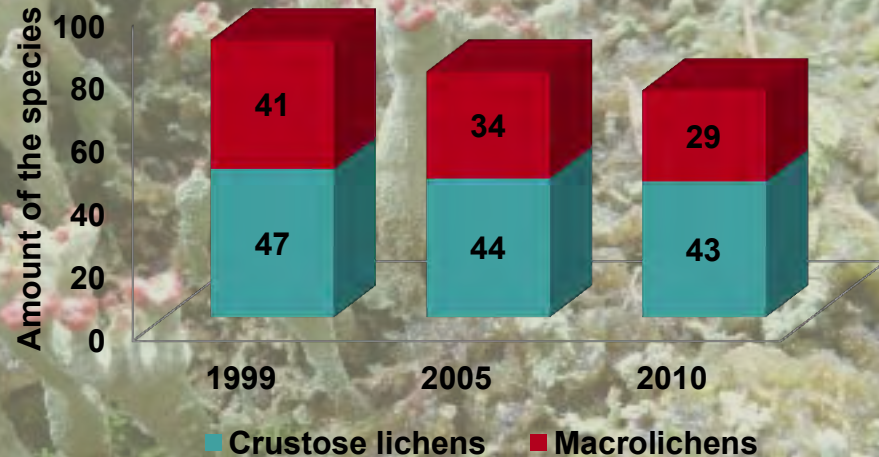
Additional Canopy Uptake: $8\text{-}10 \text{ kg N ha}^{-1} \text{ y}^{-1}$

OPEN N CYCLE DUE TO CHRONIC N DEPOSITION



- Mull humus with low C:N ratio (17) and low microbial N immobilization
- From the total N deposition, 68% is lost to the groundwater (mostly NO_3^-) or as gaseous efflux (NO , N_2 , N_2O)

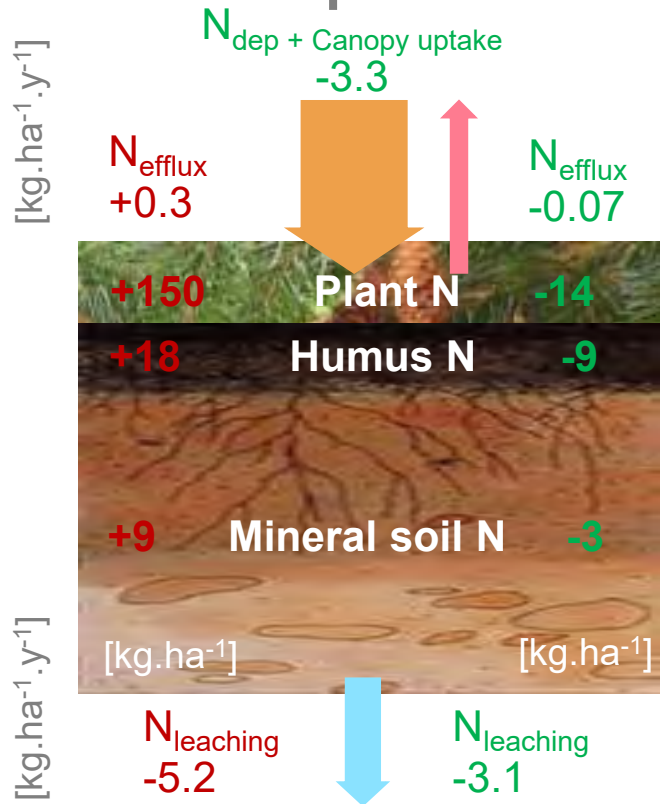
HIGH N DEPOSITION NEGATIVELY AFFECTS BIODIVERSITY



2100 scenario

+ 3 °C warming

N Reduction
Current Legislation



FUTURE REDUCTION IN N DEPOSITION AND CLIMATE CHANGE MAY INCREASE ECOSYSTEM N RETENTION

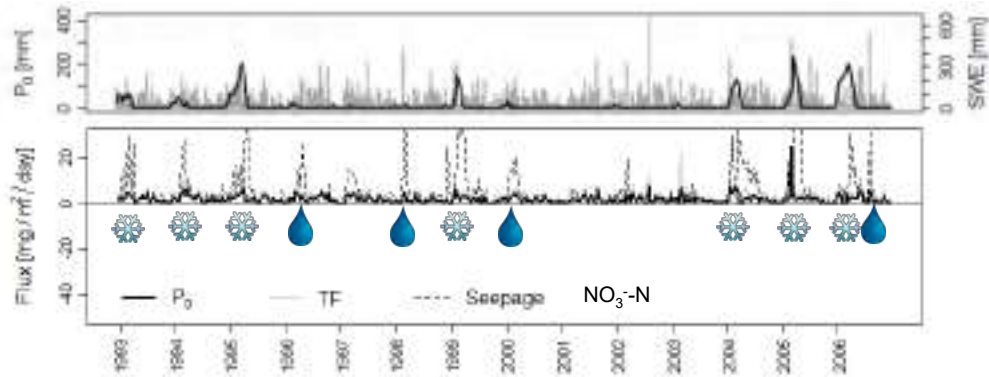
- According to current legislation **N deposition will decrease by $\sim 3 \text{ kg N ha}^{-1} \text{ y}^{-1}$** causing **less N loss**
- Expected **climate warming** will particularly **increase immobilization of N in trees**
- Overall N retention is expected to increase rendering the system **significantly less leaky**

CLIMATE AND DISTURBANCE EVENTS MAY BE MORE IMPORTANT IN FUTURE

- Scenarios rarely capture changes in seasonal variation and extreme climate events
- Droughts, stormflows and snow dynamics may change in an unexpected way
- Forest disturbances are predicted to increase
- What can we learn from the measured N dynamics?

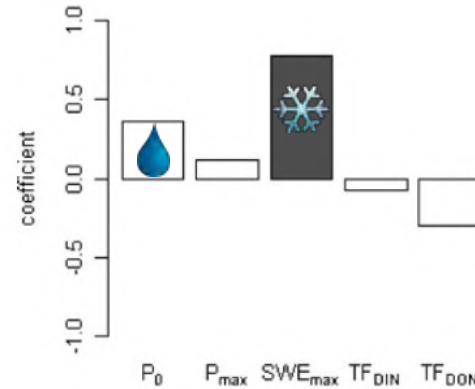


CLIMATE EVENTS DRIVE NITRATE LOSS TO THE GROUNDWATER AND NOT THE AVERAGE CLIMATE



Long-Term climate and Nitrate fluxes

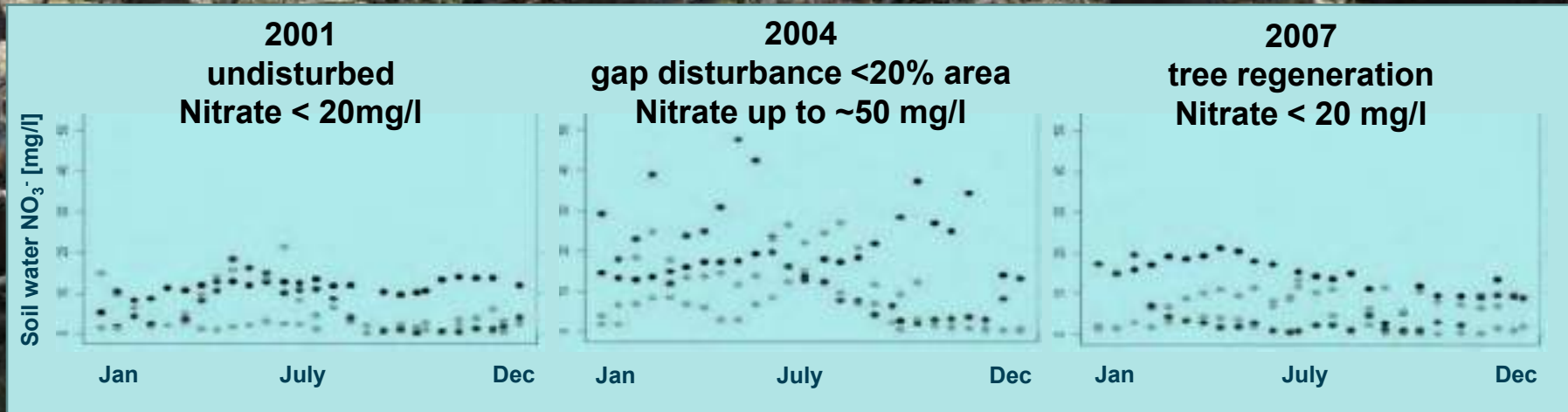
Very high variation in $\text{NO}_3\text{-N}$ seepage fluxes



Coefficients of determination of NO_3 leaching

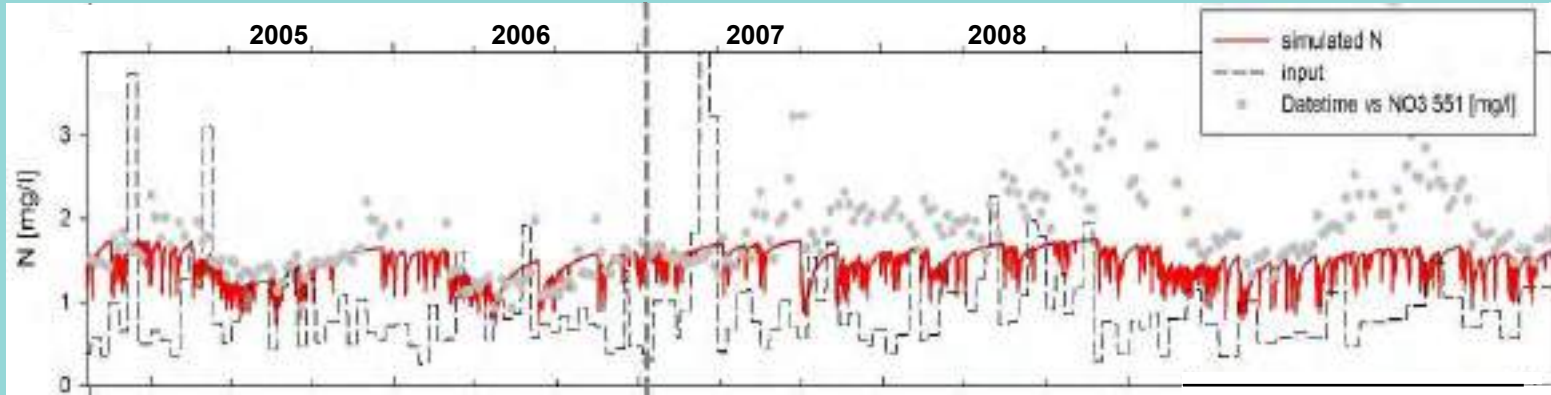
Nitrate leaching is predominately controlled by snow melt and heavy rain events

FOREST DISTURBANCE STRONGLY REDUCES N RETENTION CAUSING ELEVATED NITRATE LOSS



undisturbed

disturbed



Catchment response to forest disturbance

Windthrow at approx. 5-10% of the catchment

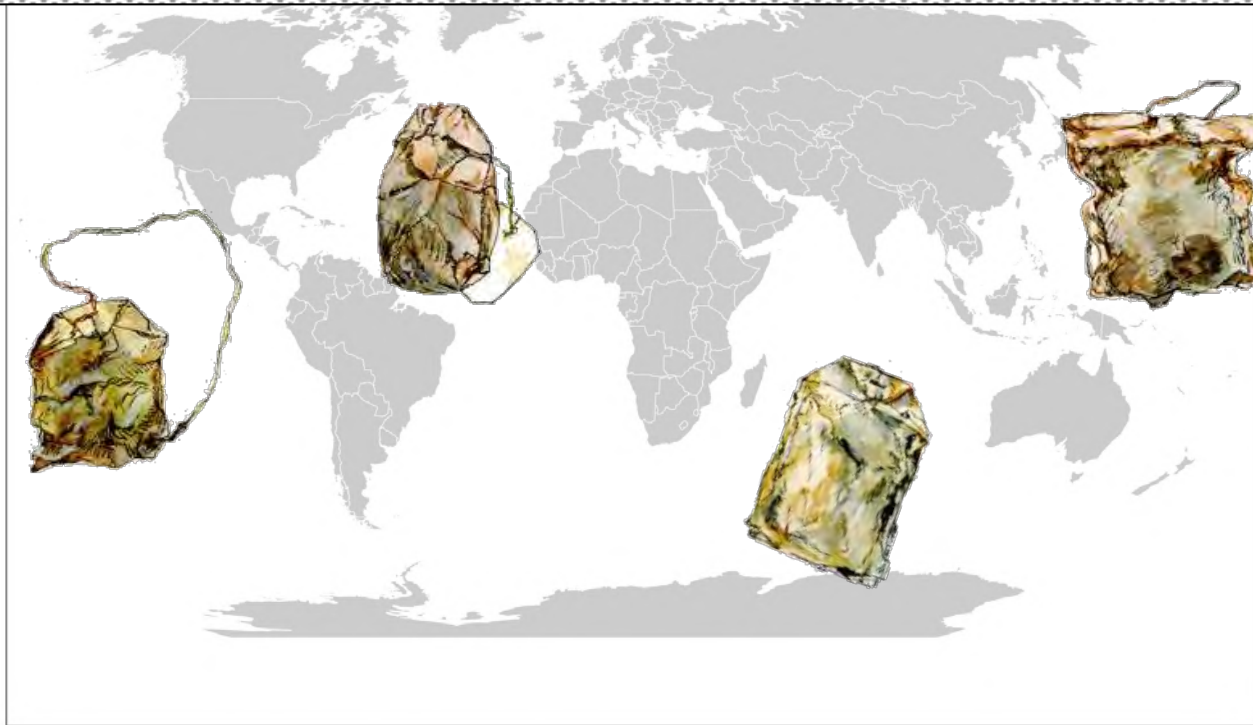


CONCLUSIONS AND OUTLOOK

- Long-term high N deposition in the Northern Limestone Alps has resulted in forest ecosystems with particularly **open Nitrogen cycles**
- Future **reduction in N deposition** according to current legislation **and climate change will cause higher N retention**
- **Future changes in climate events** may have a significant impact on how the ecosystem and the karst catchment will be able to retain Nitrogen
- The expected increase in **forest disturbances may additionally cause N pollution pulses** with long-lasting consequences on drinking water quality
- The ICP IM and LTER site **Zöbelboden will be part of LTER-CWN**, a network for the investigation of extreme climate events on C, N and water cycles in Austrian ecosystems

GLOBAL LITTER DECOMPOSITION STUDY

TEA
COMPOSITION



TeaComposition H₂O – Global Aquatic Litter Decomposition Initiative



Blue Carbon Lab
A DEAKIN IDEA



Contact: Stacey Trevathan-Tackett
s.trevathantackett@deakin.edu.au

Common Metric: Phytometer

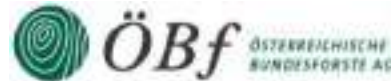


Contact: Wilfahrt Peter
Peter.Wilfahrt@uni-bayreuth.de

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Umweltbundesamt
www.umweltbundesamt.at

ILTER Austria Konferenz
Illmitz ● 3.5.2017

***Gas supersaturation may cause effects
on the biota comparable to acidification***

Gaute Velle, Ulrich Pulg, Sebastian Stranzl,
Sondre S. Kvalsheim, Godtfred A. Halvorsen

GAS SUPERSATURATION

BACKGROUND

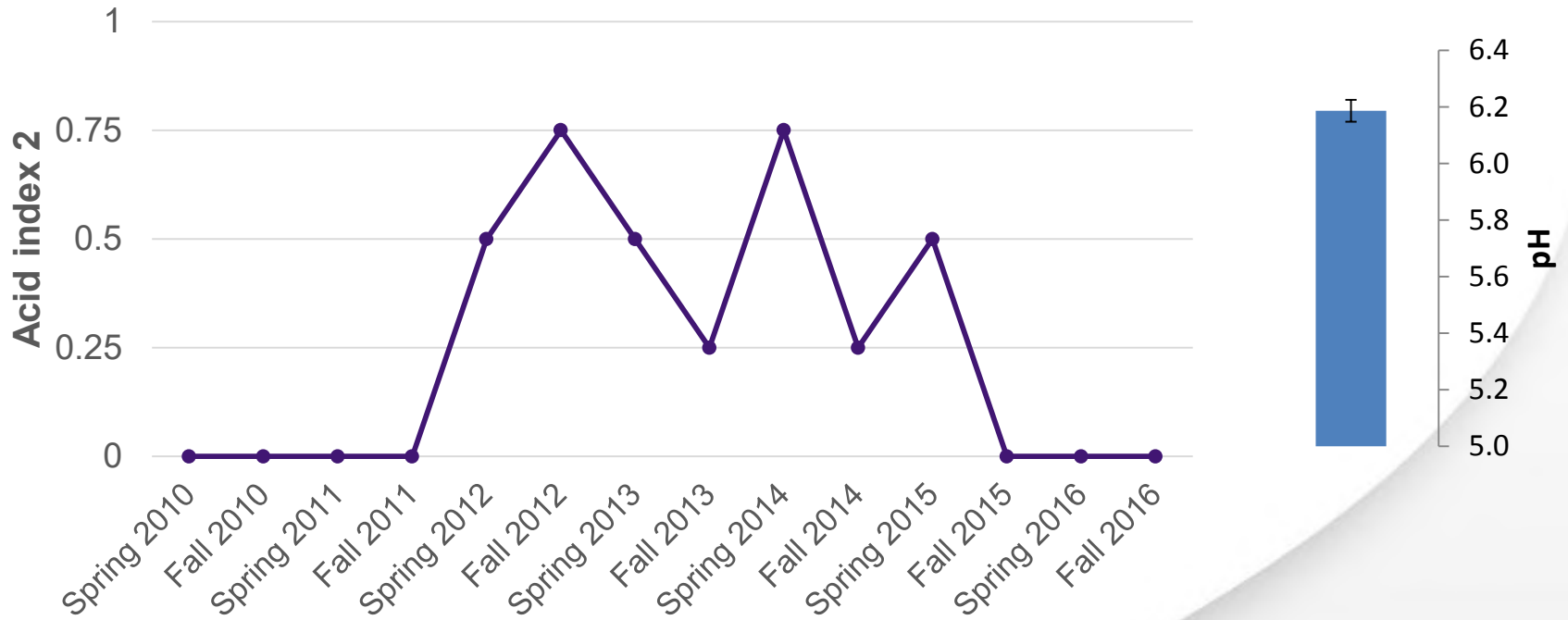
MEASUREMENTS

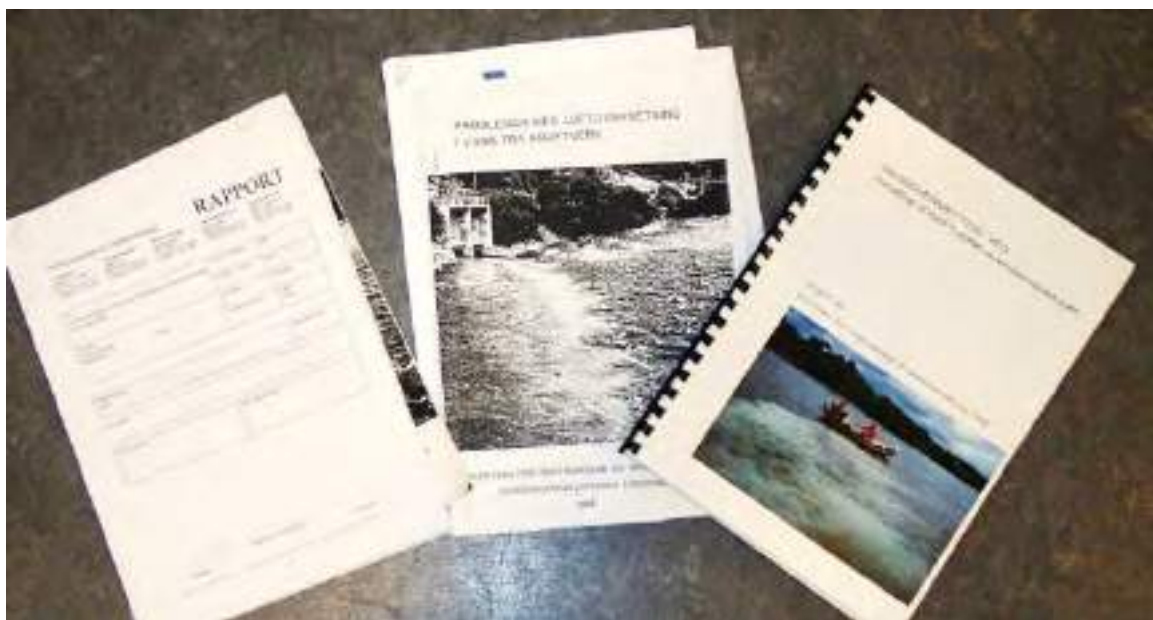
CAUSES

BIOLOGICAL EFFECTS

CONCLUSIONS AND RELEVANCE

ACID INDEX 2 IN RIVER OTRA





Gas supersaturation

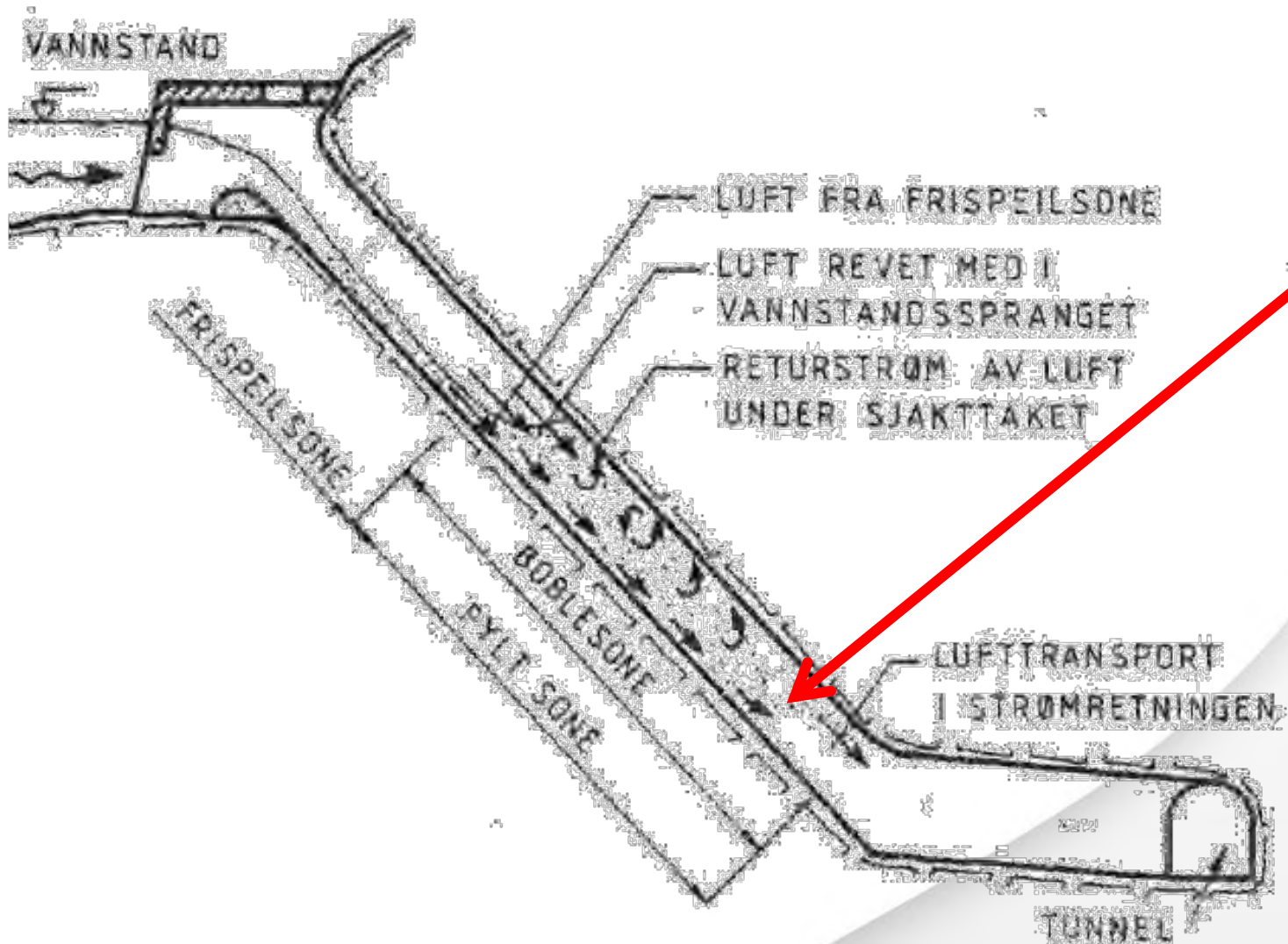
Normal river water: 100 % Total Gas Pressure (TGP)

Carbonated bottle water: 120%-130% TGP

TGP may occur naturally: Water falls; temperature changes, photosynthesis



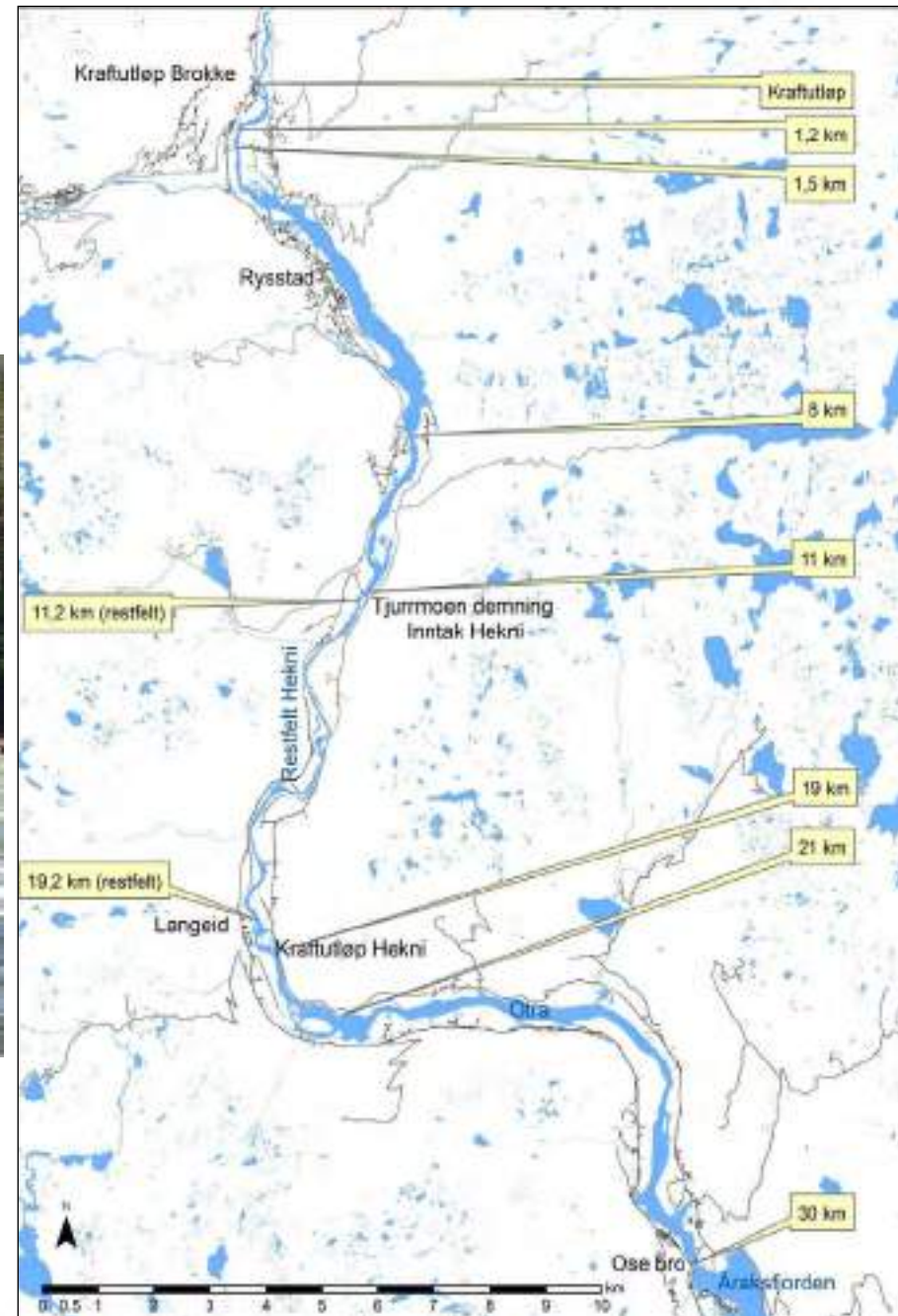
Man made: Gas + fluids + pressure + reduced pressure



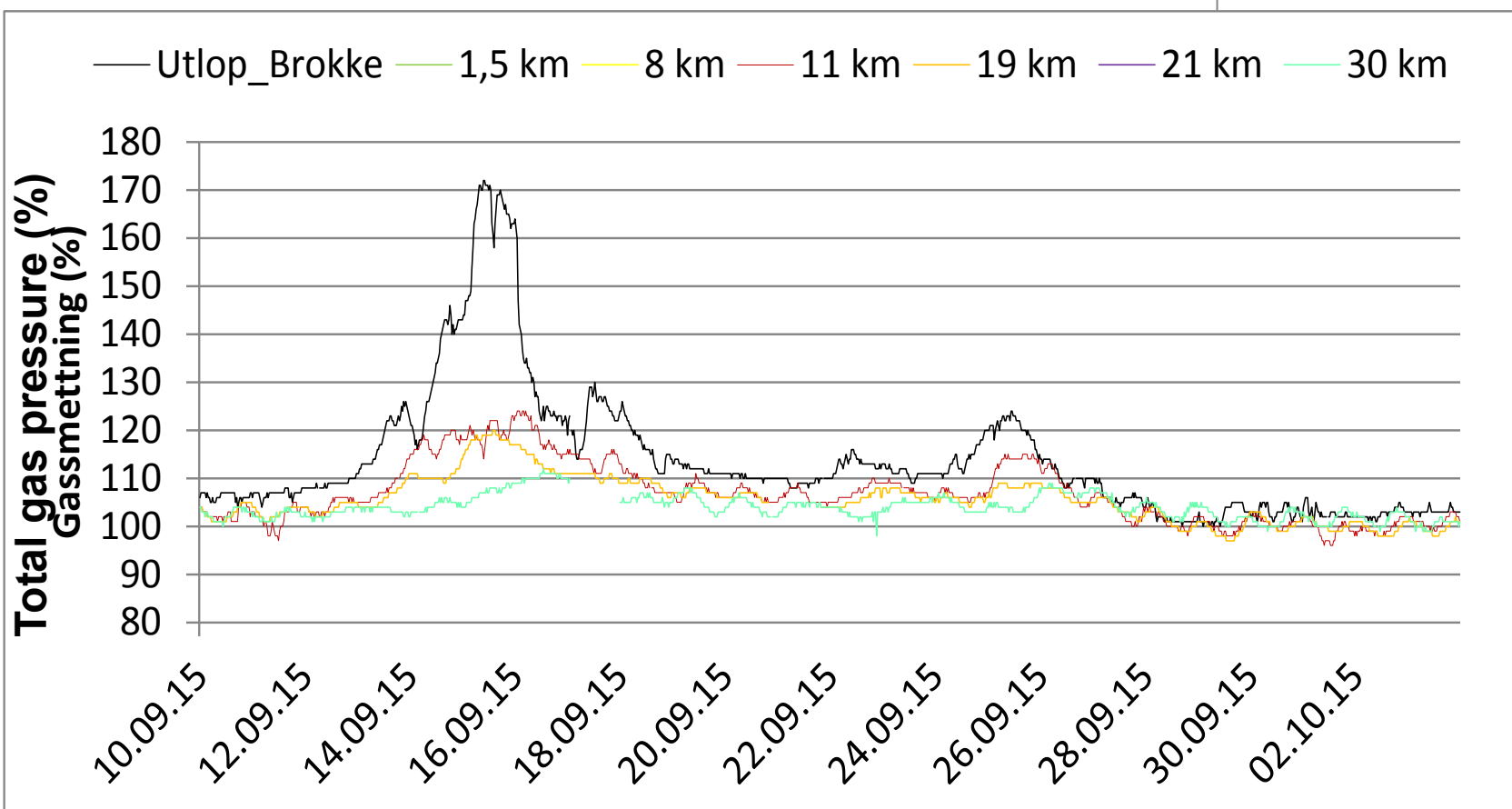
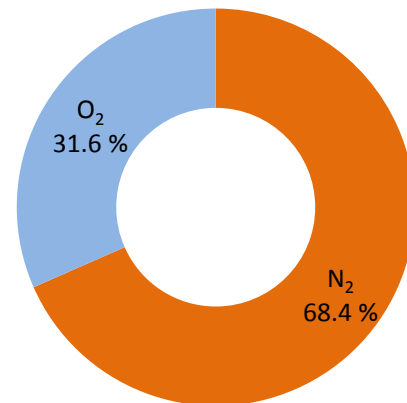
Lack of measuring equipment



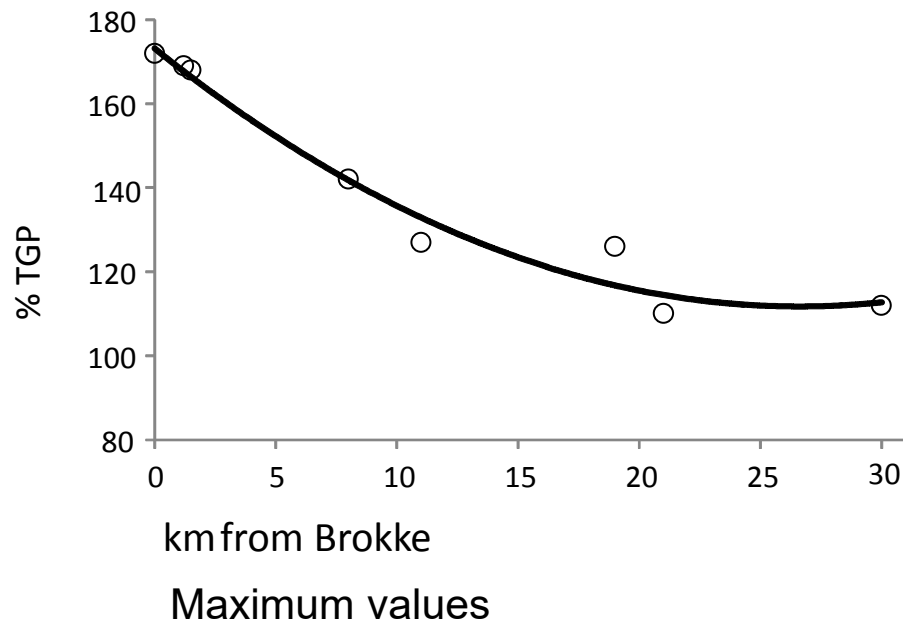
Gas saturation downstream from Brokke hydropower plant 2011-2016



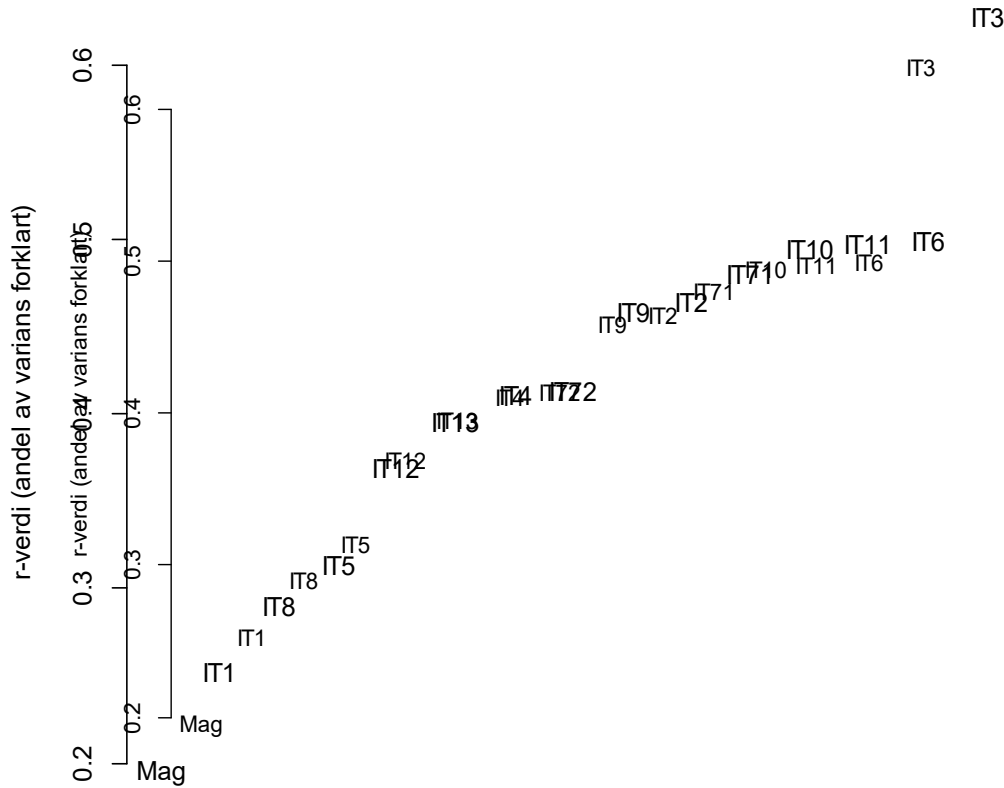
River Otra



Wave characteristics, aeration, dilution



Cause



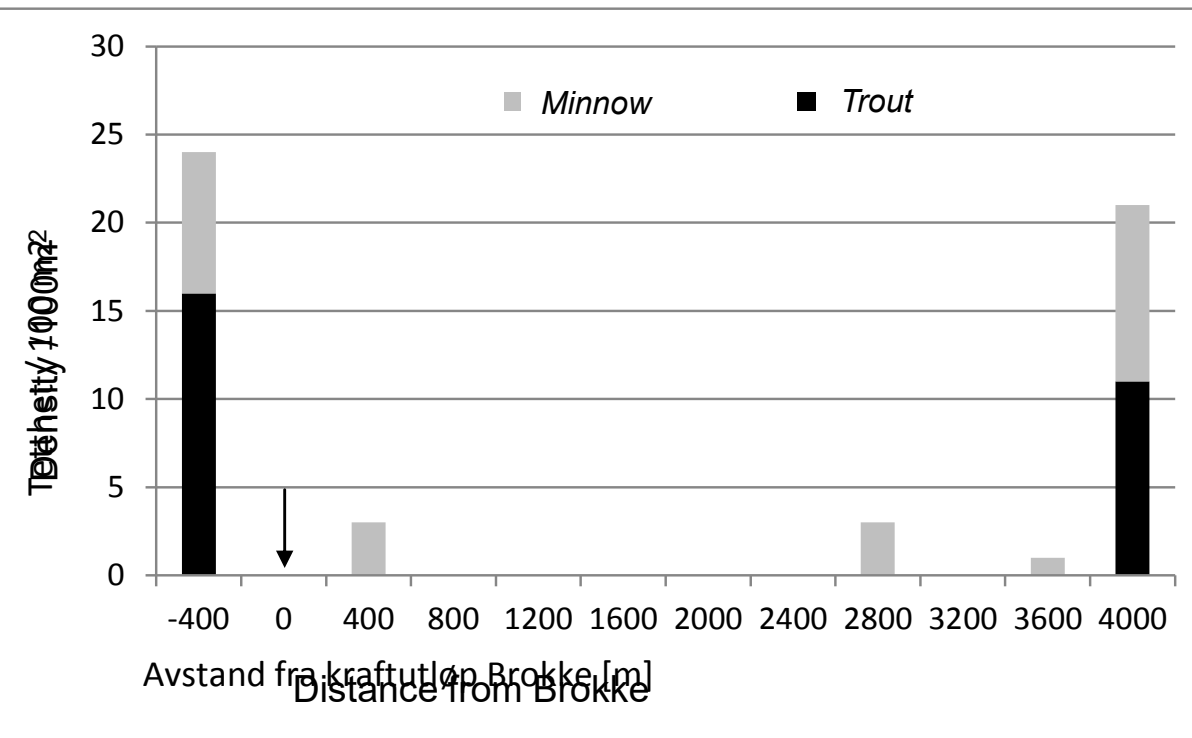
Nr			
3	6	Ljosåni	Ljosåni
6	11	Lisleå	Lisleå
11	10	Faråna	Faråna
10	10	Havestøylani	Havestøylani
71	71	Flossi	Flossi
2	2	Hisa	Hisa
9	9	Hylesdalsar	Hylesdalsar
72	72	Flossi liten	Flossi liten
4	4	Gjesløy	Gjesløy
13	13	Myklevatnet	Myklevatnet
12	12	Kvernani	Kvernani
5	12	Skiptesbekken	Kvernani
8	5	Kvinnani	Skiptesbekke
1	8	Holsbekken	Kvinnani
	1		Holsbekke

Cause

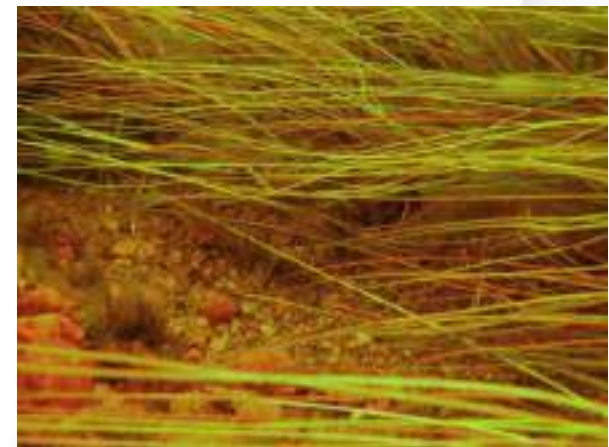


Effects on the biota

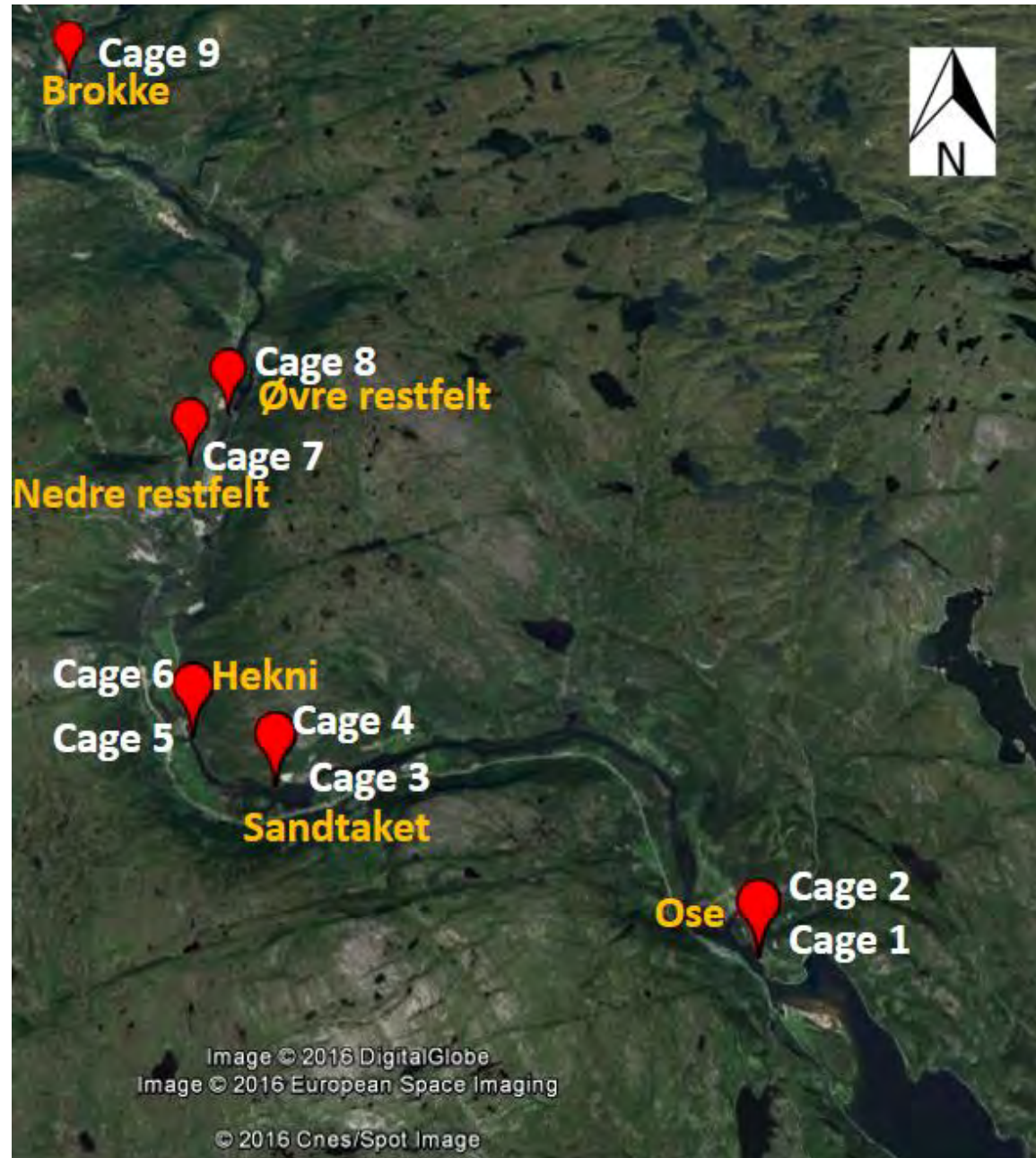
Electrofishing, Otra



Data fra el-fiske 30.-31. 10. 2012. Ørret ble bare funnet ovenfor kraftutløpet og i Rystadbassenget. Ørekyte fantes også innimellom men i svært lave tettheter og samtlige ble funnet ved bekkemunninger.



Cage experiments







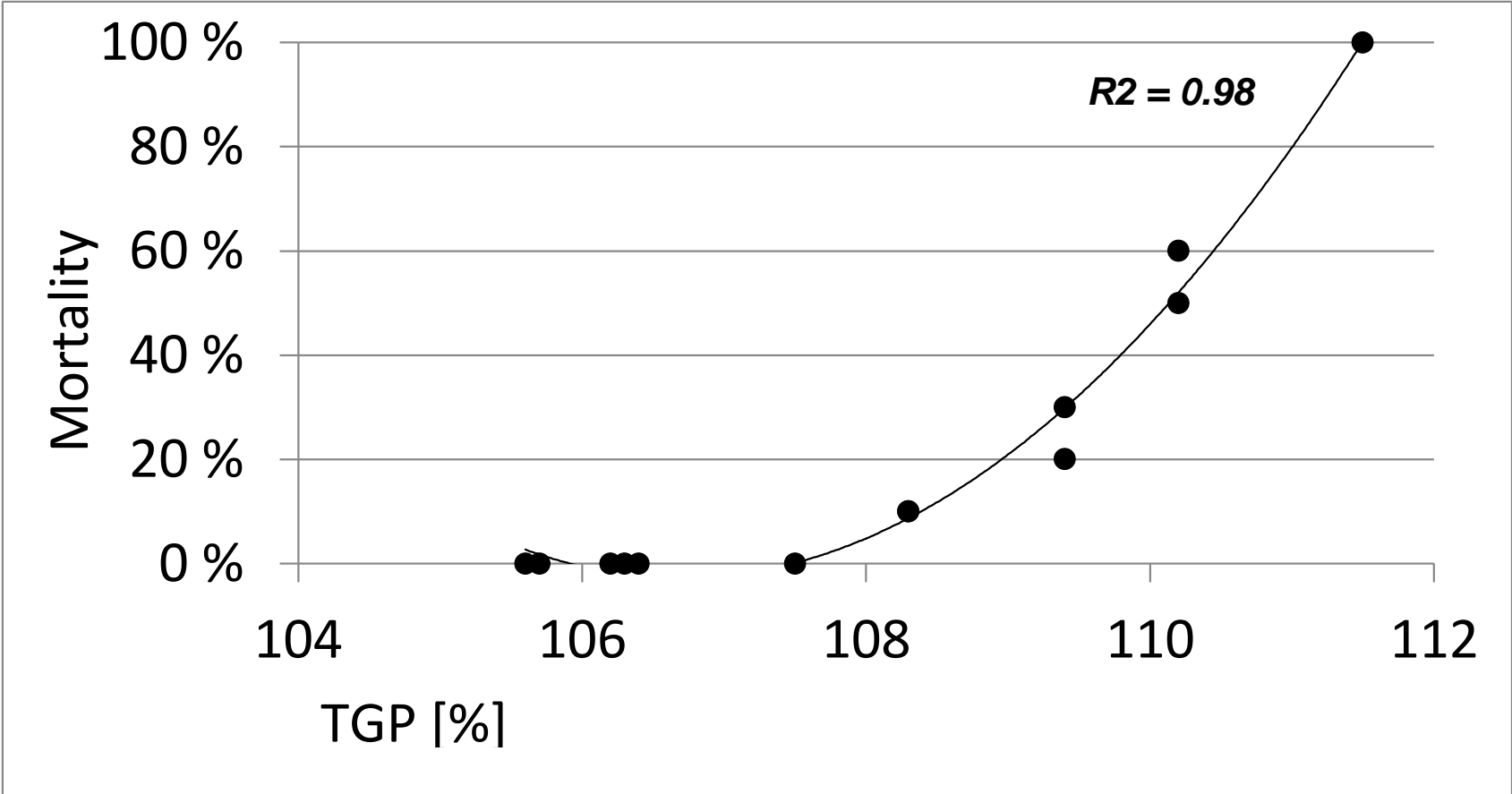
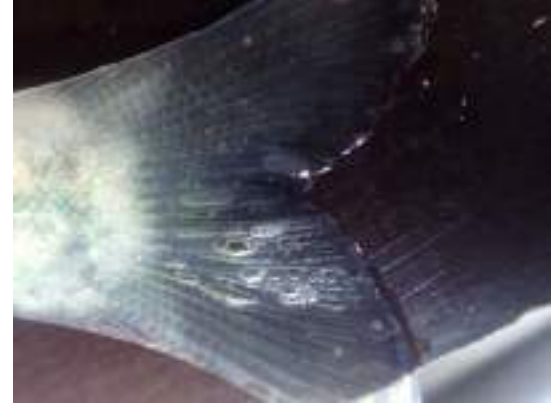
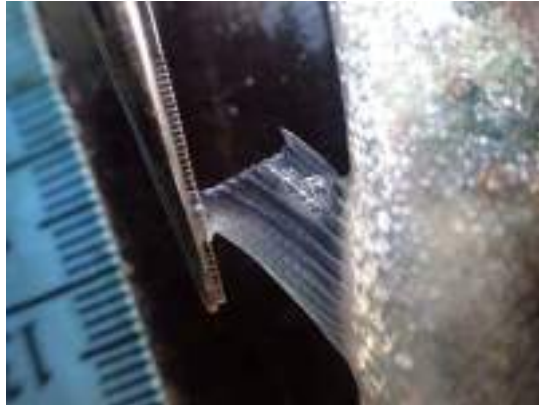




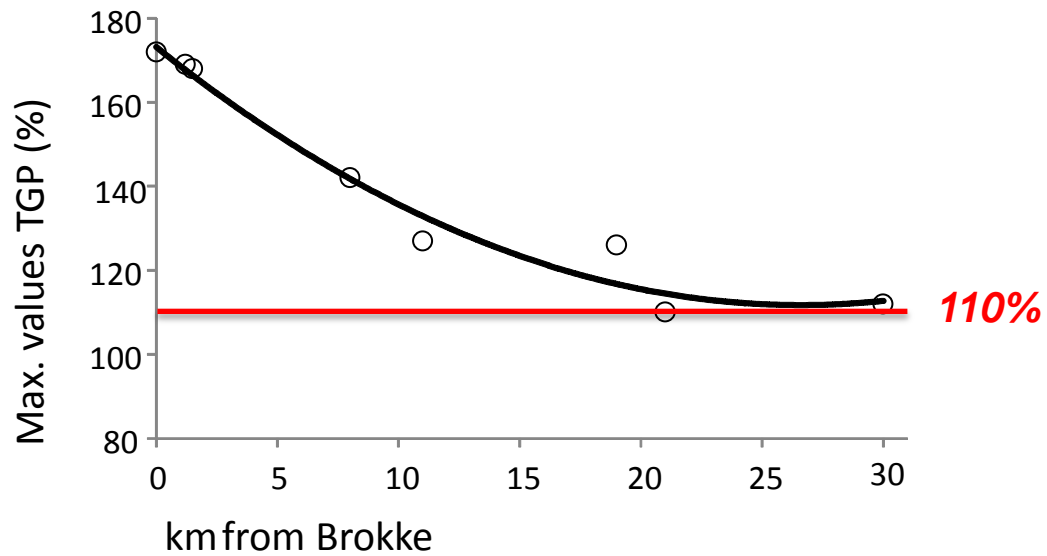


Right gill arch 1 = metal analysis
Left gill arch 2 = histology
Right gills = check for gas bubbles
Fins = bubbles
Blood = Plasma cortisol





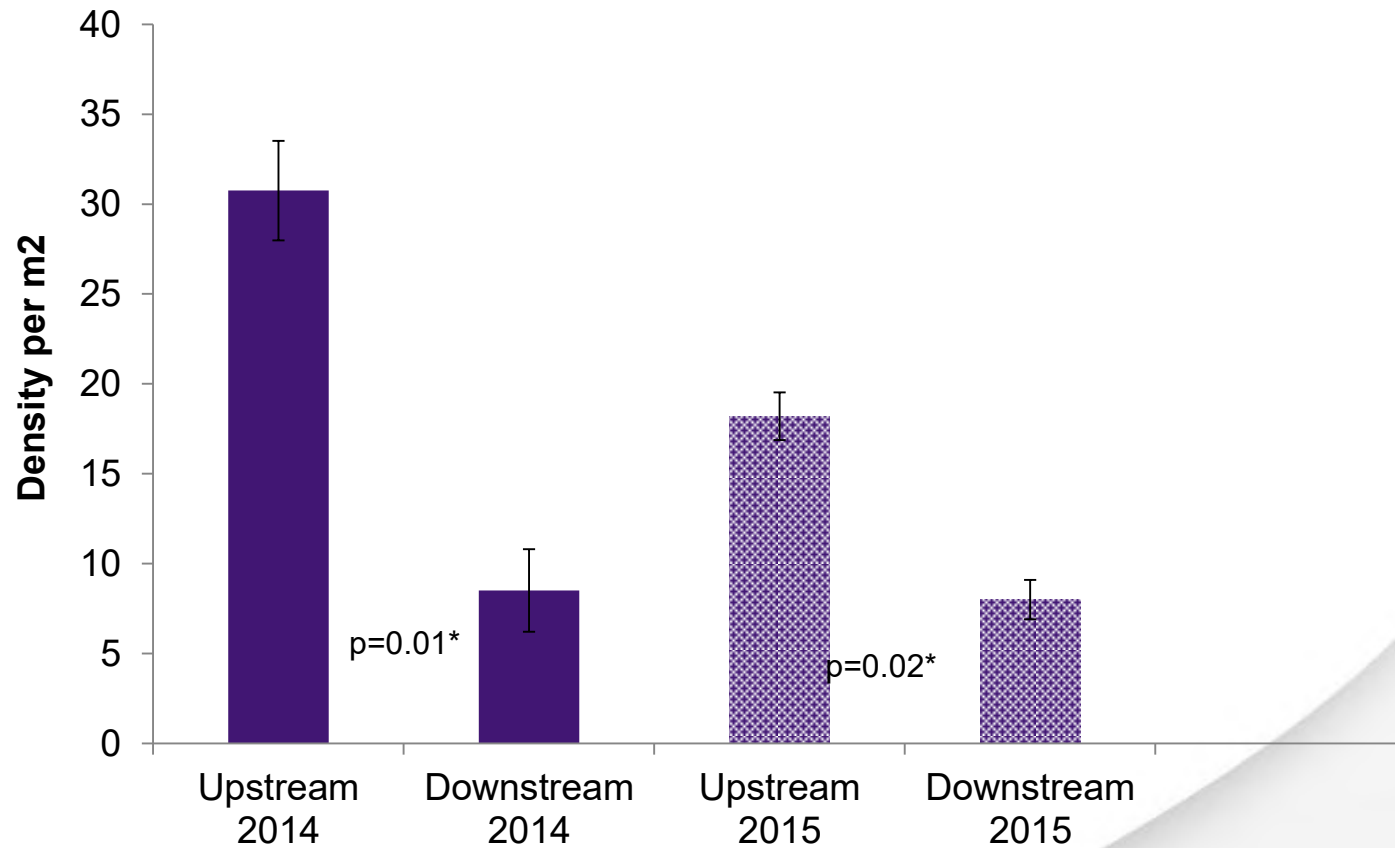
LD50 = 110% gas supersaturation



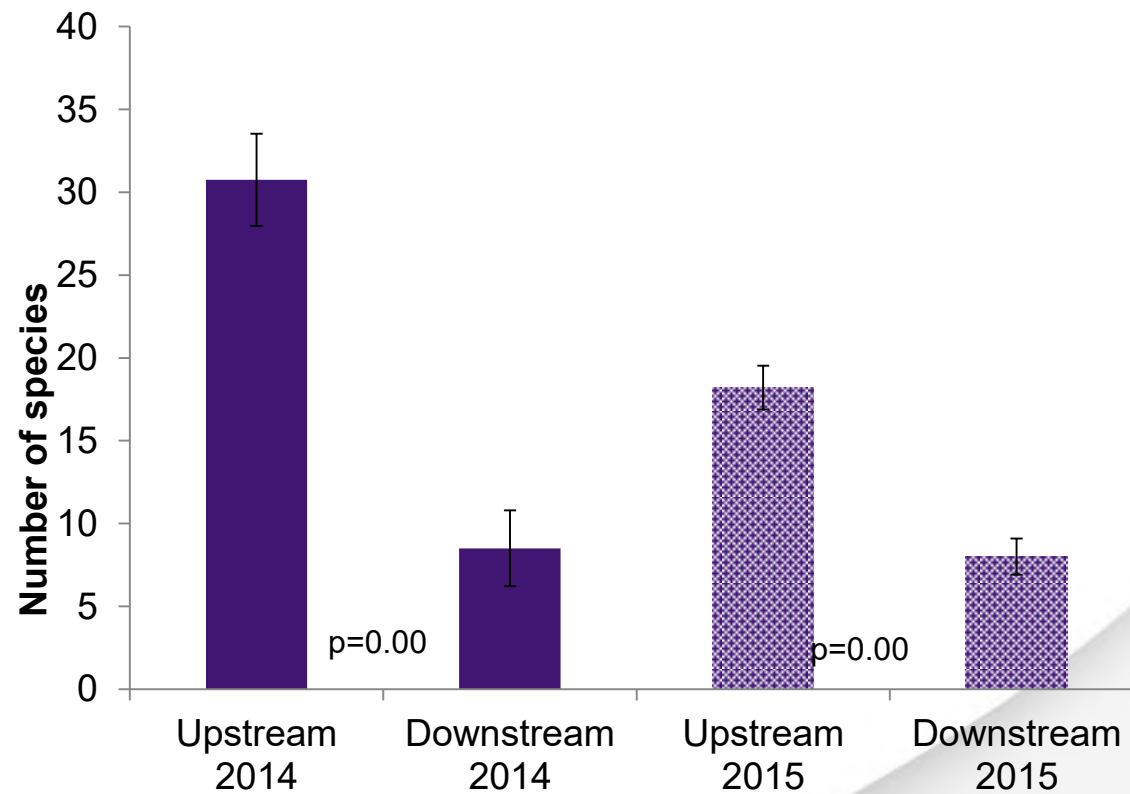
Effects on benthic invertebrates



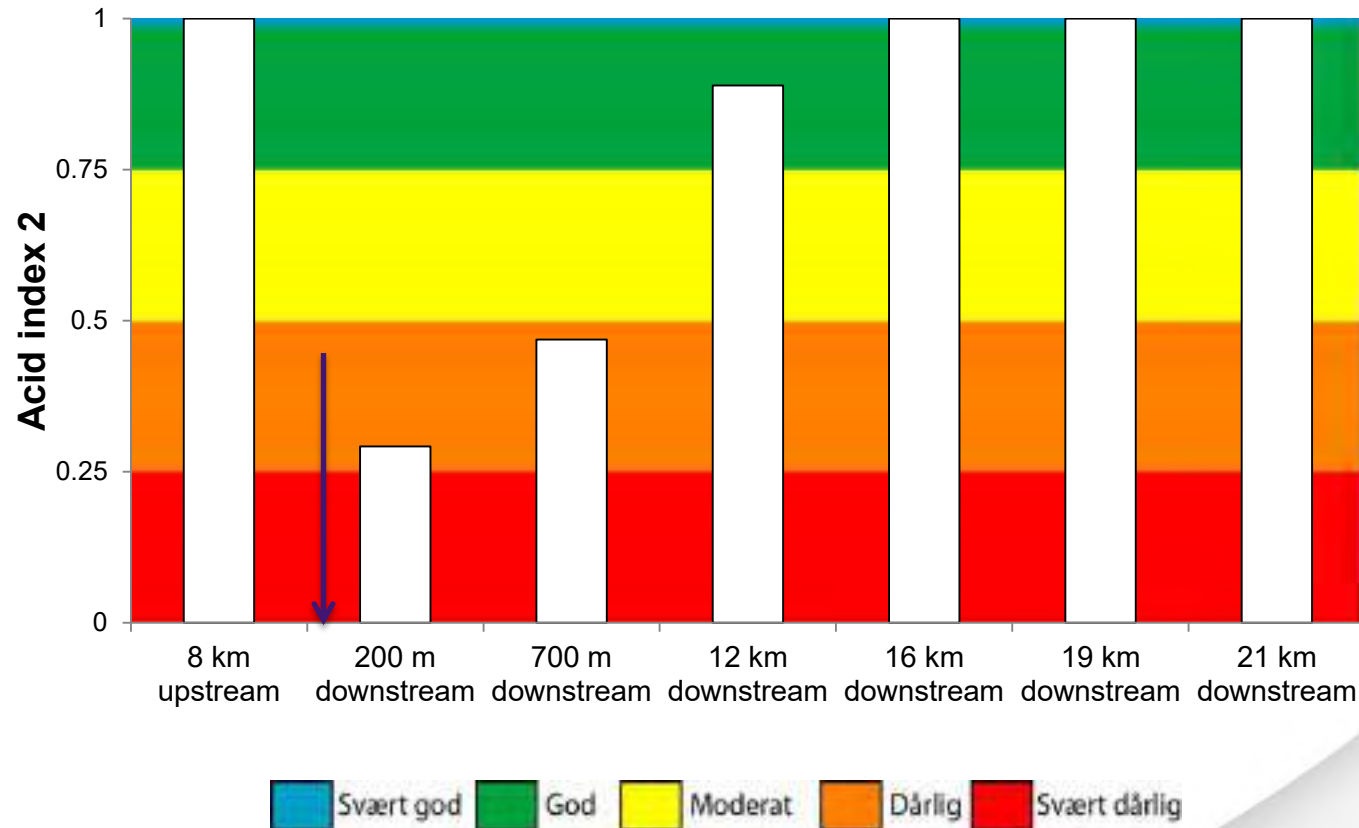
Density of invertebrates



Diversity of invertebrates



Average acidification index 2010-2015



Fauna dominated by digging taxa

200 m upstream		700 m downstream		11 km downstream	
Taxon	No./m ²	Taxon	No./m ²	Taxon	No./m ²
Chironomidae	6896	Nematoda	1513	Chironomidae	6299
Oligochaeta	1407	Oligochaeta	1299	Oligochaeta	842
Acari	1103	Acari	235	Nematoda	188
Oxyethira sp.	776	Chironomidae	145	Acari	168
Nematoda	458	Apatania sp.	51	Oxyethira sp.	84
Simuliidae	318	A. borealis	23	Ostracoda	64
Oecetis testacea	305	Simuliidae	13	Leptophlebia marginata	23
Lepidostoma hirtum	302	Tipula sp.	8	Pisidium sp.	22
Leuctra fusca/digitata	283	Nemoura cinerea	7	Empididae indet.	19
Pisidium sp.	240	Collembola	7	Simuliidae	15
Radix balthica	220	Empididae indet.	6	Apatania sp.	14
Amphinemura borealis	199	Pisidium sp.	6	Leuctra hippopus	13



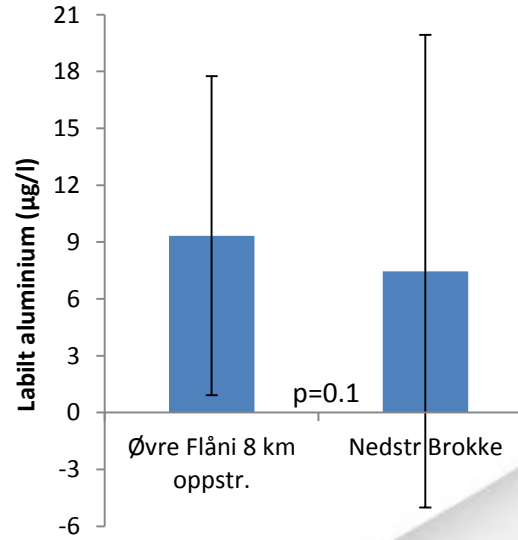
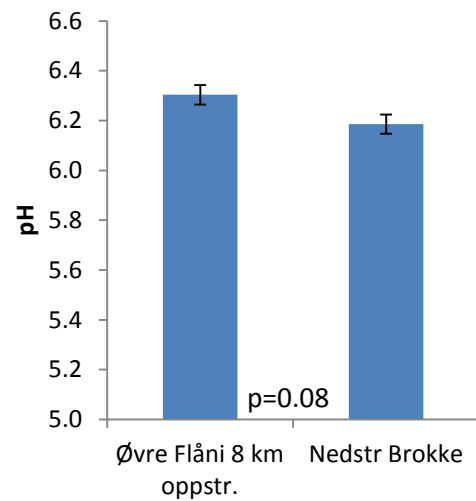
Possible causes

- Substrat



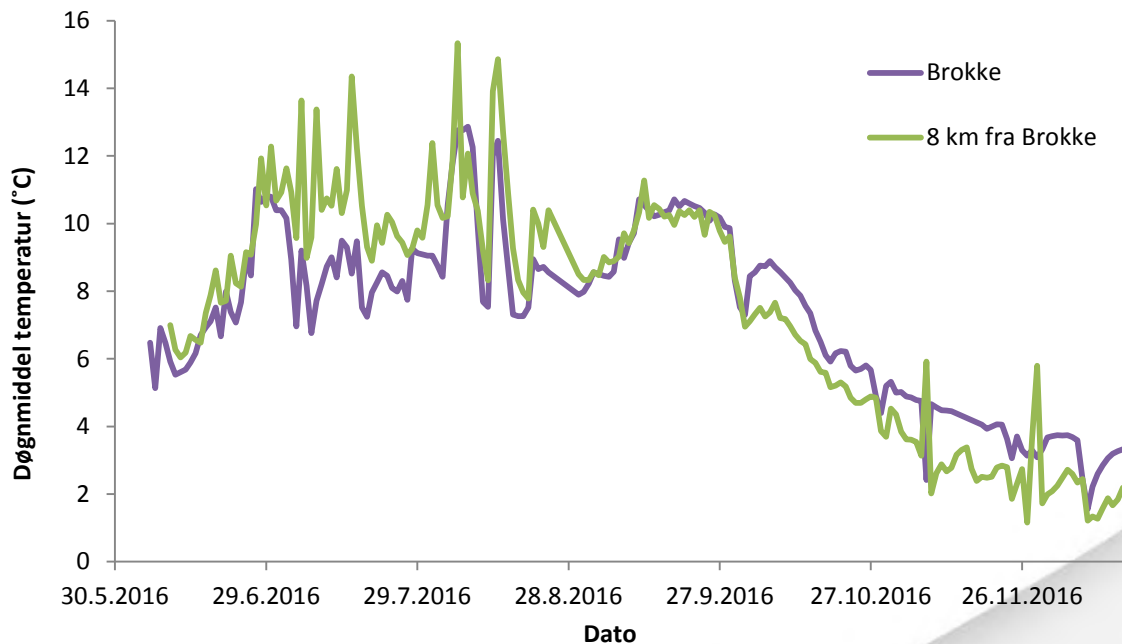
Possible causes

- Substrat
- Water chemistry/ organic load



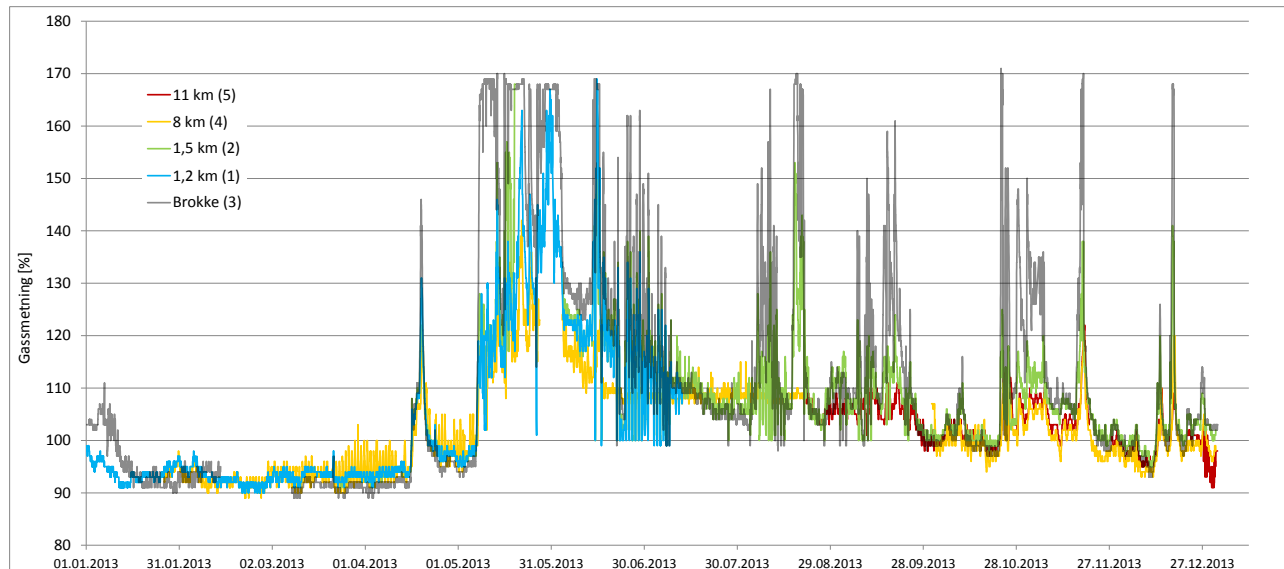
Possible causes

- Substrat
- Water chemistry/ organic load
- Temperature



Possible causes

- Substrat
- Water chemistry/ organic load
- Temperature
- Gas supersaturation



Summary Otra (2011-2016)

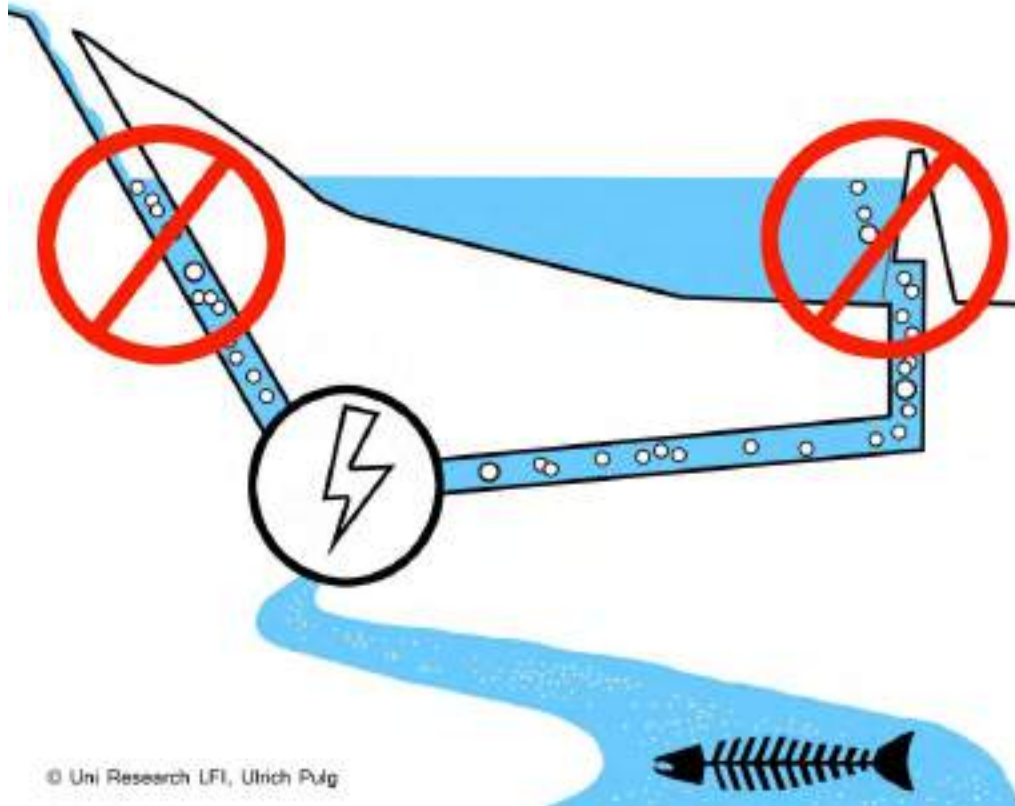
110% for > 20 km

100-110% for >30 km

4 km without fish

12 km increased fish mortality

8 km influence on invertebrates



Conclusions

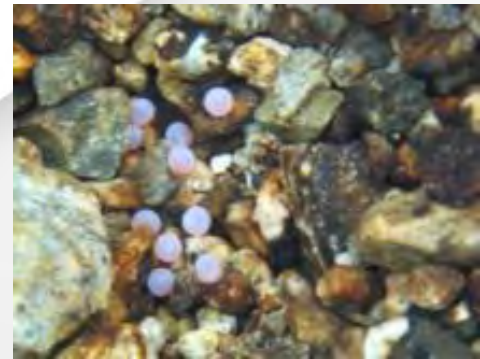
- **Supersaturation may severely influence the the biota**

Lack of knowledge

- Occurrence of supersaturation elsewhere
- Effects on benthic animals and zooplankton
- Sub-lethal doses and avoidance behaviour of fish
- *No regulations or awareness exists*

Relevance for ICP waters

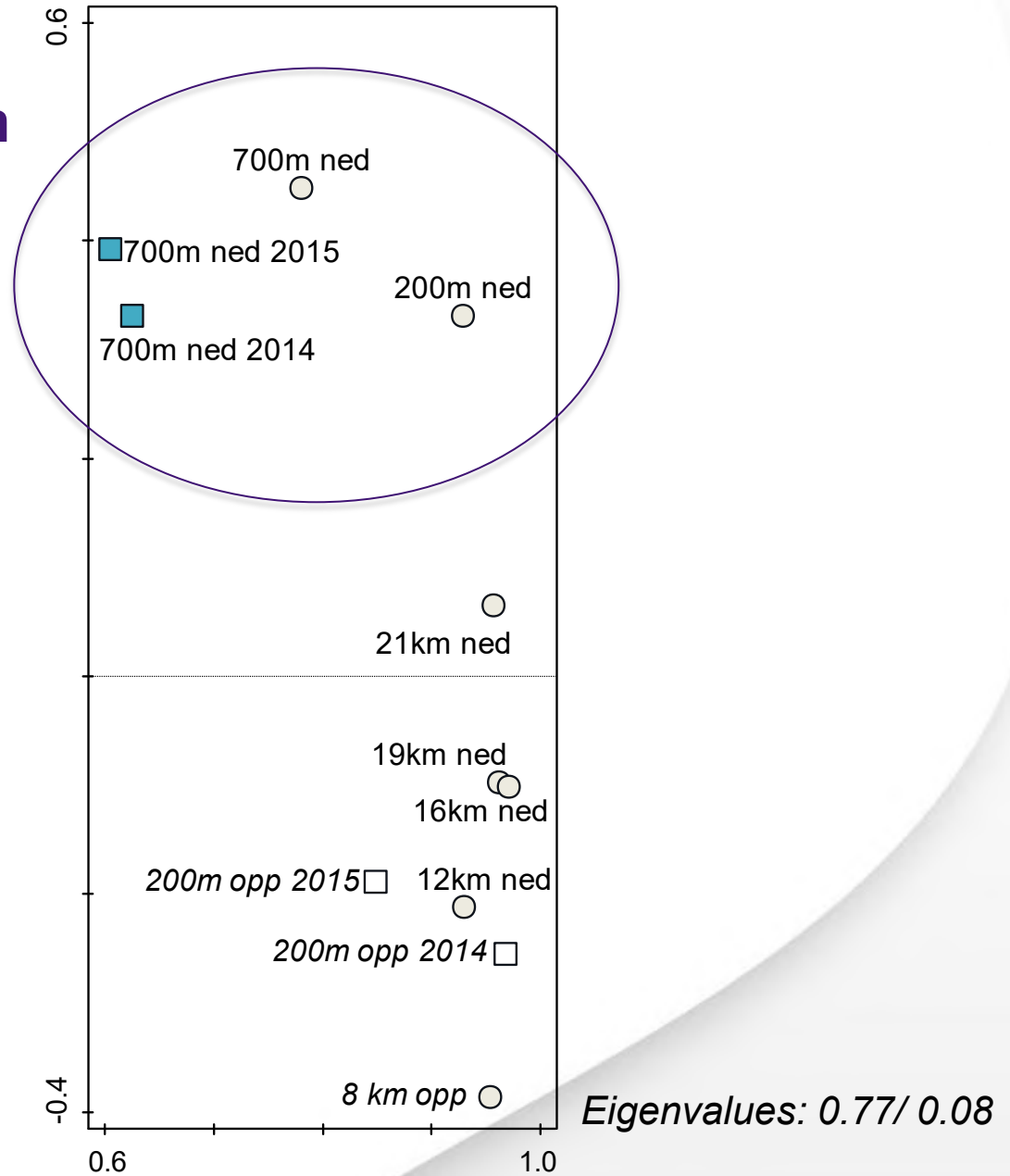
- Biota may indicate acidification when the cause is supersaturation
- May occur downstream from dams and power plant outlets, also downstream of falls?

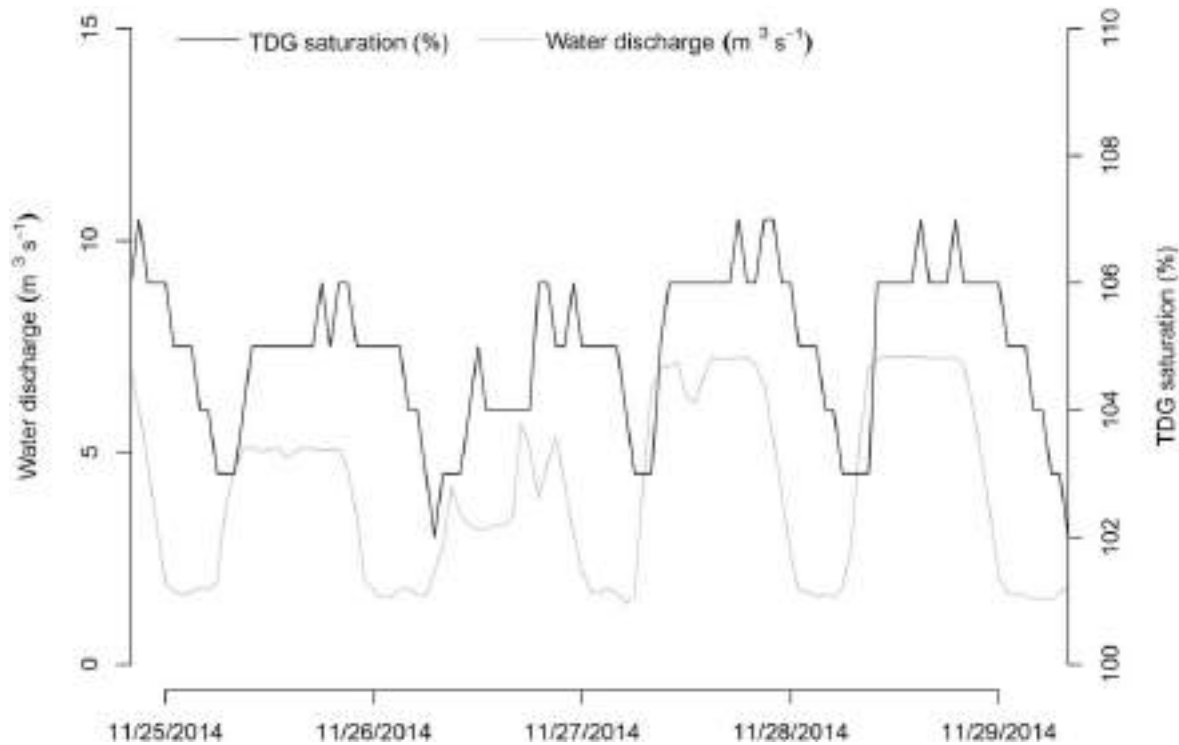


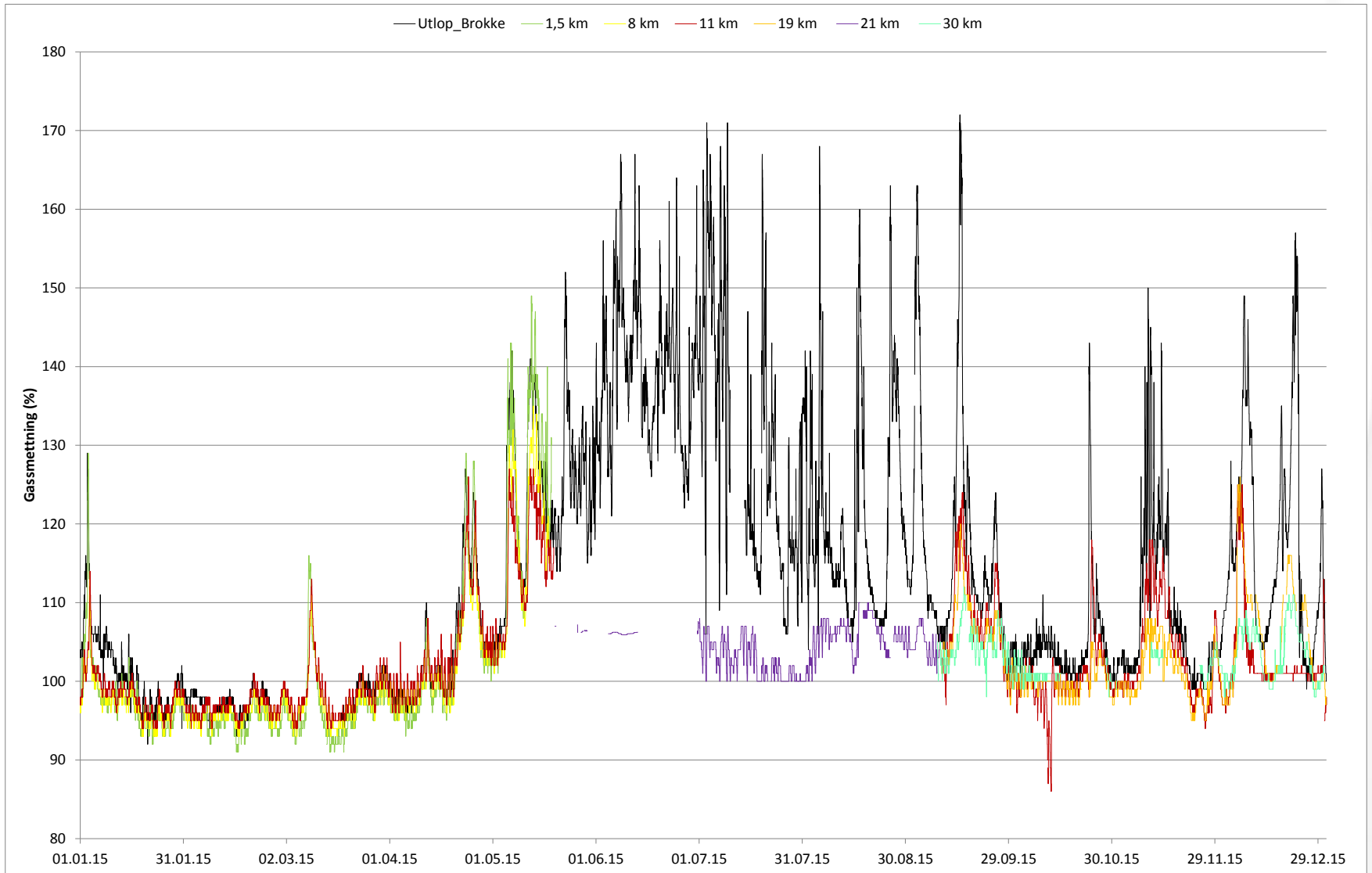
Thanks for your attention!



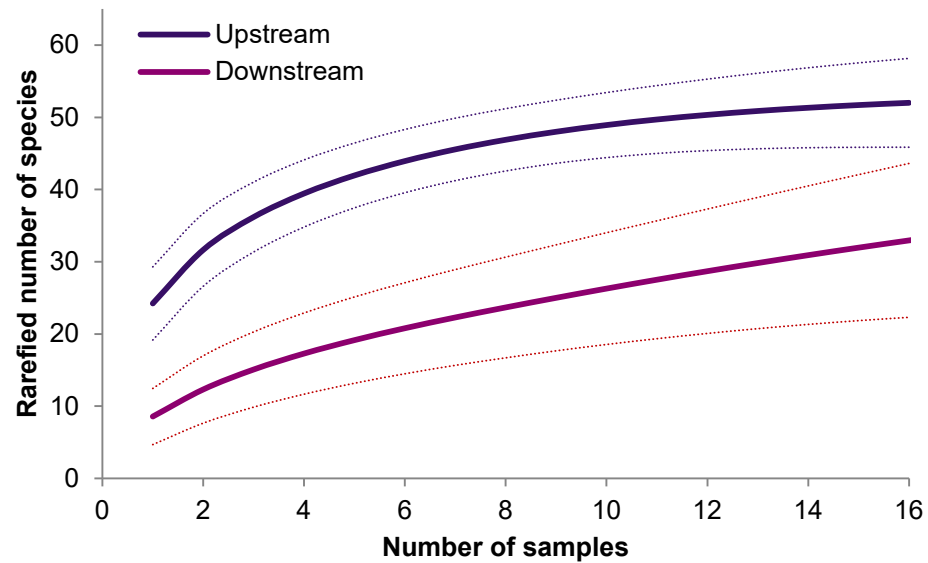
Species composition







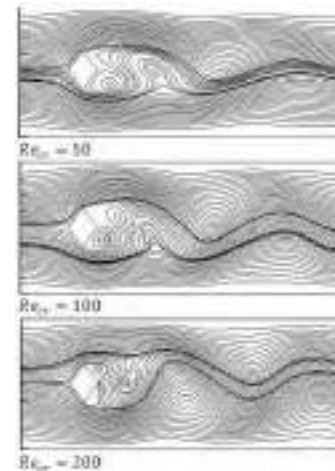
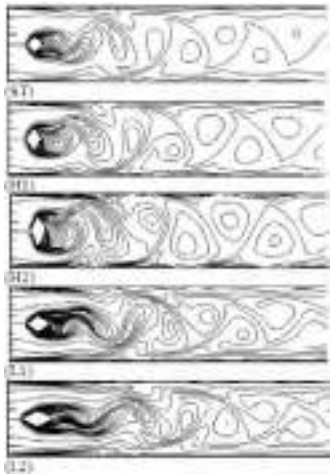
Accumulation curves for species richness



Low Pressure / Variable Pressure Zones in Rivers

Mitigation measure: Pressure Variability due to form drag?

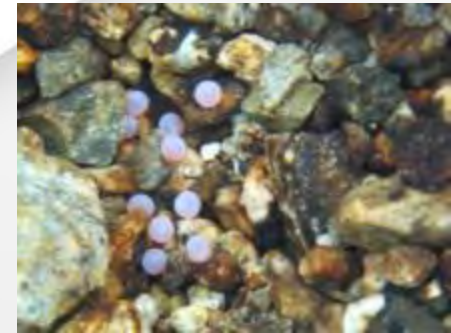
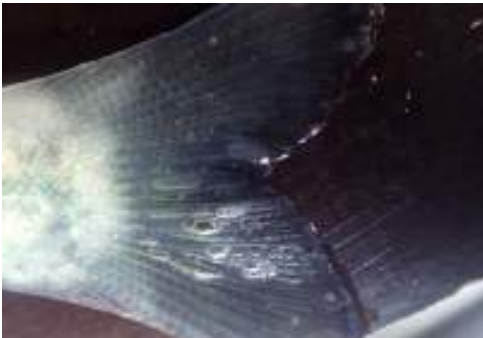
Form drag: Form drag occurs because localized flow separation can create a high pressure upstream from an object and a low pressure downstream in the objects wake (Wohl, 2015). The resulting pressure-gradient force opposes flow and creates viscous energy losses downstream of the object (Tritton, 1988; Roberson & Crowe, 1993).



Biological effects from gas supersaturation

Fish

- Gas bubble disease
- Depending on species, life stage and depth
- Fish in hatcheries have increased mortality from 102-103 %
- Fish in rivers experience acute death from 110 %
- Immune system and behaviour influenced below 110%

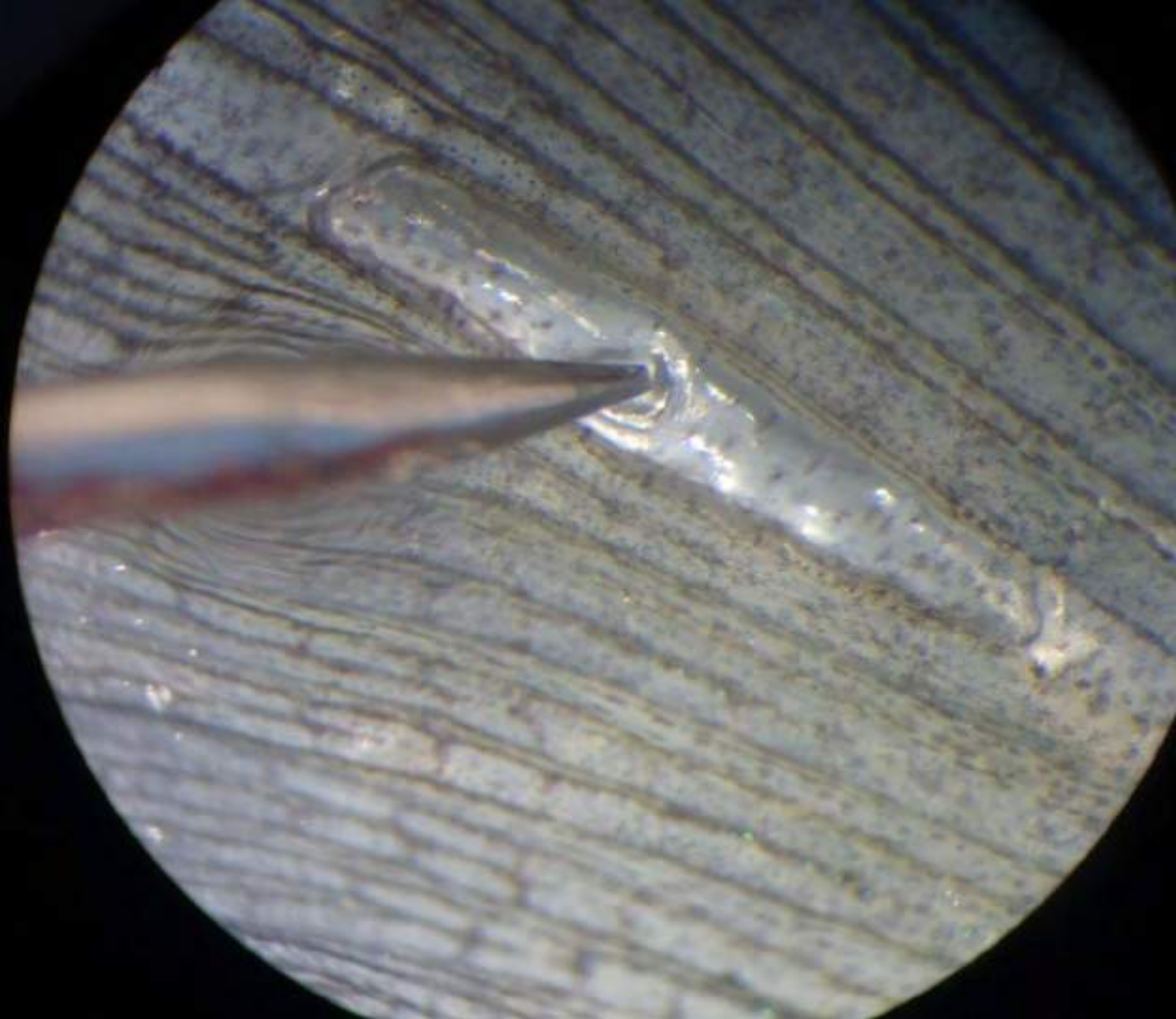


Biological effects from gas supersaturation

Other groups

- Bullfrog 120% (*Rana catesbeiana*, Colt et al. 1984)
- Zooplankton 115% (*Daphnia magna*, (Nebeker et al. 1975).
- Crayfish: 120%-127% (*Orconectes limosus*, Nebeker et al. 1976)
- Stoneflies: 130% (*Acroneuria californica*, *A. pacifica*, *Pteronarcys californica*)
- Aquatic plants: no data– buoyancy effects? CO₂- boost?

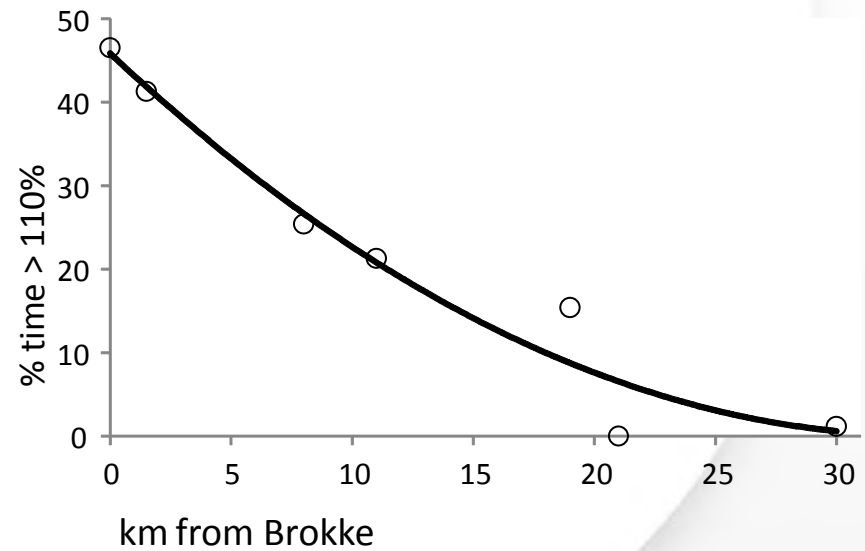
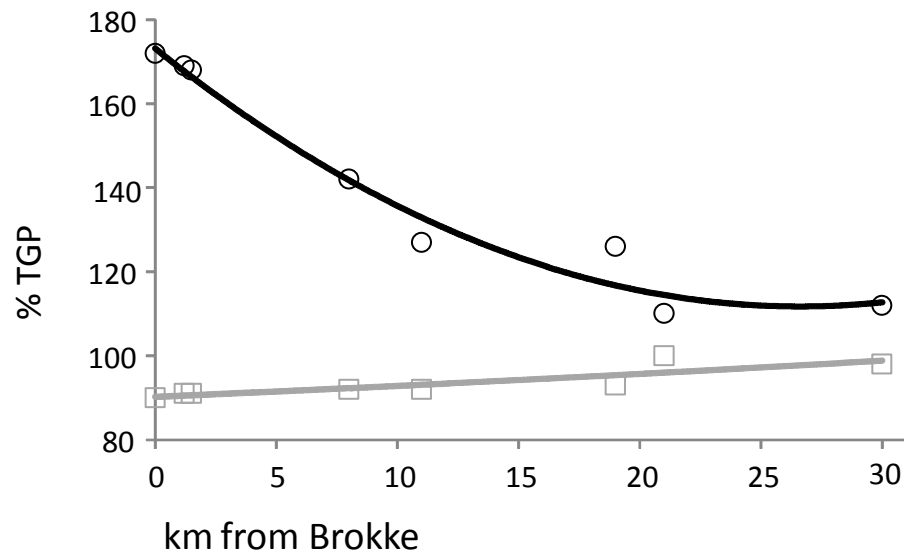


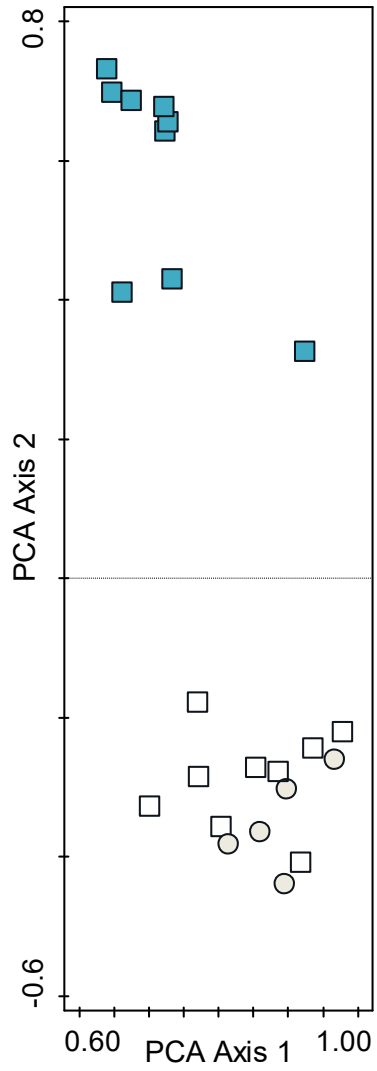


Tiltak mot gassovermetning

- Gjøre noe med kilden; Vanskelig å isolere hvilke inntak til Brokke som bidrar mest til overmetningen
- Opprettholde loggestasjoner som beskriver situasjonen med fravær/forekomst av gassovermetning – bestemme influensområde (rom & tid)
- Hvordan responderer fisk og bunndyr på eksponeringen – unnviklsesatferd?
- Kartlegge influensområdet nedstrøms Tjurrmodammen dvs. hvilke deler av blekas nåværende utbredelsesområde er påvirket – 2 km - 12 km
- Miljøbasert kjøring av Hekni dvs. planlegg stans i Hekni i perioder uten gassovermetning slik at restfeltet ikke får overmettet vann (Uni kan utrede når det er gassovermetning og anbefale konkrete kjøreregler)- allerede innført?
- Tiltak med å lufte ut gassovermettet. Mulighet for utgassing på fallstrekningen nedstrøms Brokke. Uni anbefaler dette og kan sammenfatte eksisterende data fra mange vassdrag og fra litteratur for å bestemme potensialet.

Wave characteristics, aeration, dilution





**Co-analysis of coniferous forest state
parameters and atmospheric deposition data
series obtained by ICP IM and EMEP at
European part of Russia**



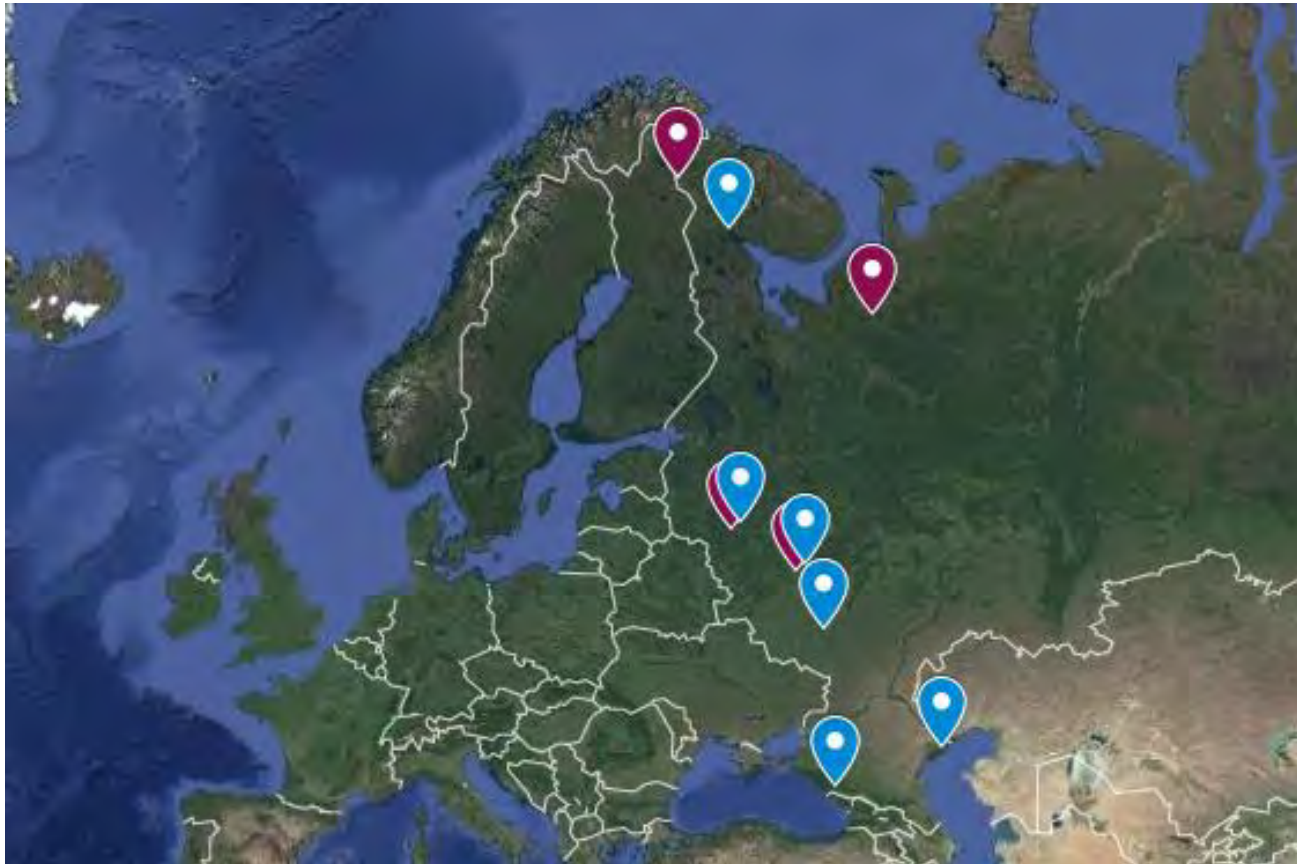
Pozdnyakova E.

Institute of Global Climate and
Ecology of Roshydromet and RAS



katemukudori@mail.ru

ICP IM and EMEP Russian network



**ICP IM
station**



**EMEP
station**

ICP IM and EMEP research parameters

ICP IM

Coniferous stands	Defoliation (%)
Annual sampling	
Analysis were carried out by visual assessment	
	Depigmentation (%)

EMEP

Concentration in precipitation	K⁺
	Na⁺
Daily sampling	Ca²⁺
	Mg²⁺
Analysis were carried out by methods of ion and liquid chromatography	NH₄⁺
	NO₃⁻
	NO₃⁻
	SO₄²⁻
	Cl⁻

Stands

White Sea Biological Station

fir-wood



pinery



Stands

Oka-Terrace Biosphere Reserve

fir-wood



pinery



Stands

Central-Forest Biosphere Reserve

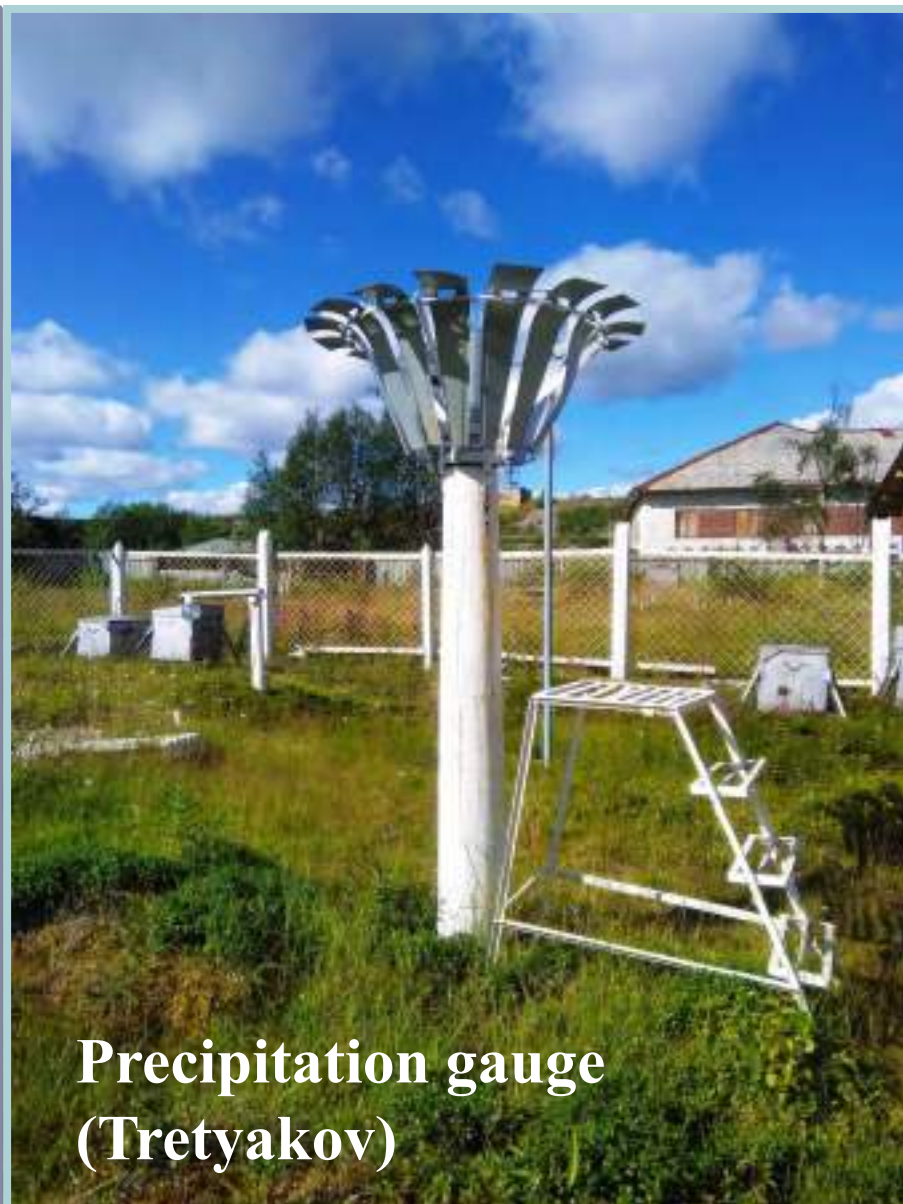
fir-wood



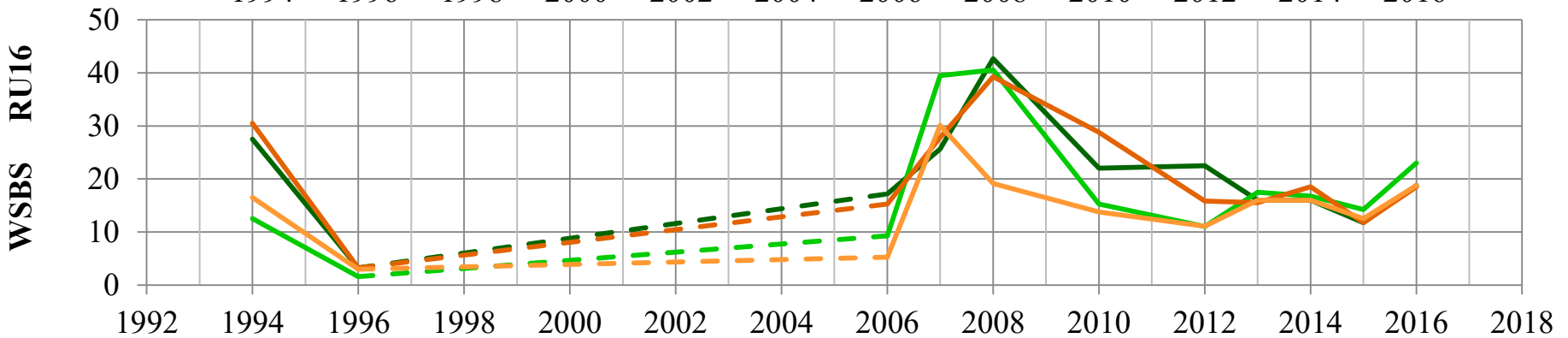
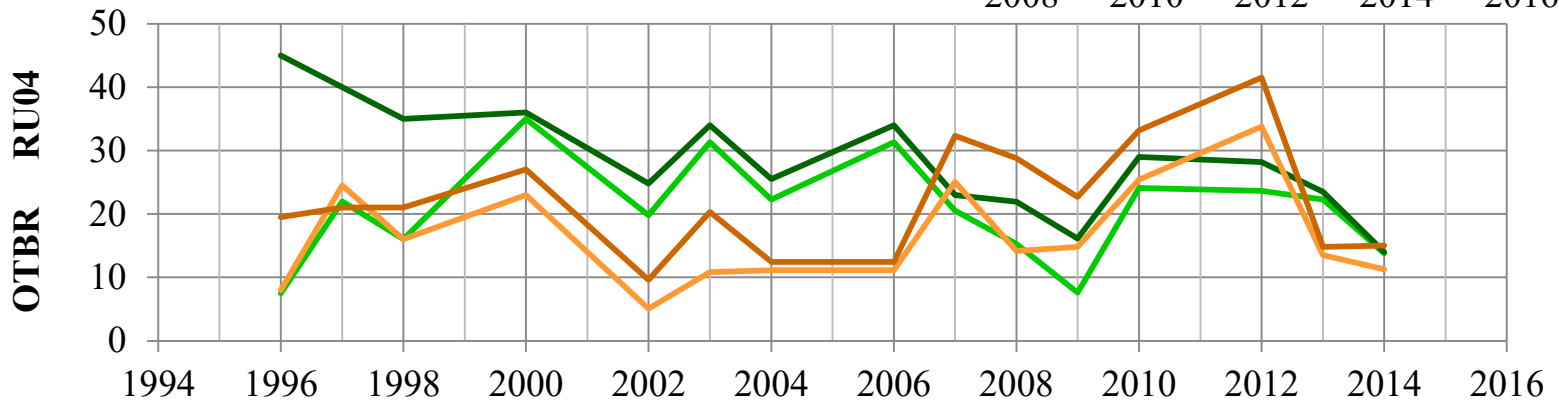
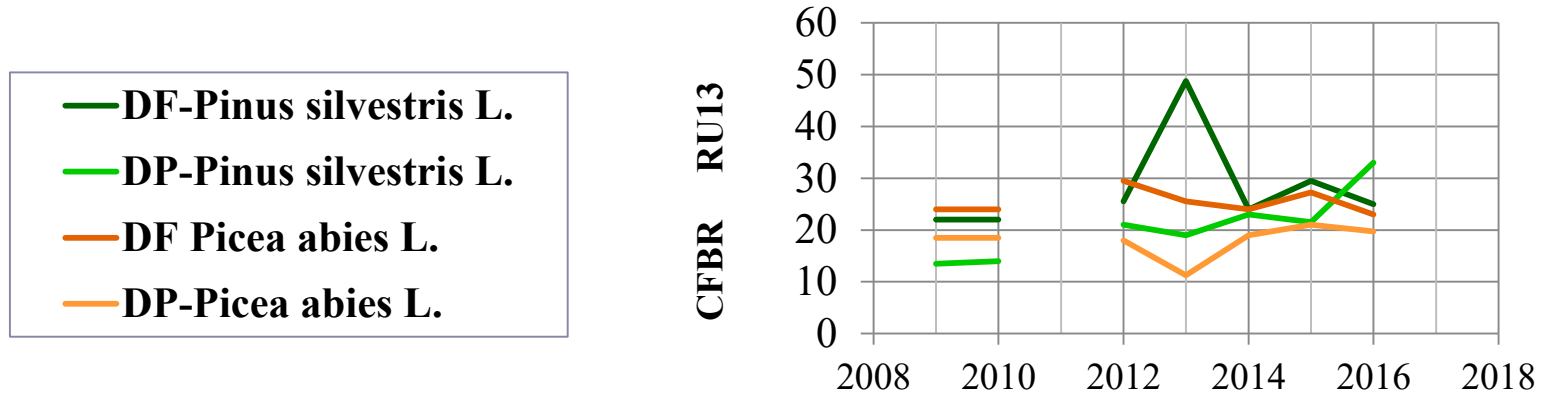
pinery



EMEP station



Interannual fluctuations of defoliation(%) and depigmentation(%) in stands of Scots pine (*Pinus sylvestris*) and spruce (*Picea abies*) .



Defoliation and depigmentation in pine and spruce stands changed unidirectionally for RU16 and RU04

Correlation coefficients ($p = 0.05$)			
RU04		RU16	
R(DF&DP)	R(DF&DP)	R(DF&DP)	R(DF&DP)
Picea abies	Pinus sylvestris	Picea abies	Pinus sylvestris
0,861	0,698	0,649	0,748

There is no the same correlations for RU13

Tab. 1 Coefficients of correlations between the parameters of coniferous stands and total wet deposition of pollutants for current(C) and previous(P) year. (p = 0.05)

Stands	Parameter	SO ₄ (S)	NO ₃ (N)	NH ₄ (N)	Na	Mg	Ca	Cl	K
Pinus sylvestris L.	DF			-0.63C; 0.63P;					-0.63C;
Picea abies L.	DF						0.58C;		
Pinus sylvestris L.	DP	-0.94C; -0.73P;			-0.68C; -0.60P;	- 0.73C;		-0.64C; -0.66P;	-0.73C; -0.63P;
Picea abies L.	DP			-0.59P;					

Tab. 2 Coefficients of correlations between the parameters of coniferous stands and concentration pollutant in precipitation for current(C) and previous(P) year. (p = 0.05)

Stands	Parameter	SO ₄ (S)	NO ₃ (N)	NH ₄ (N)	Na	Mg	Ca	Cl	K
Pinus sylvestris L.	DF	-0.69P;	0.67C;						-0.61C;
Picea abies L.	DF	-0.65C;							
Pinus sylvestris L.	DP		0.55P;	0.73P;			0.59C;	-0.58P;	
Picea abies L.	DP	0.58P;		0.74C;					

Tab. 5 Coefficients of correlations between the parameters of coniferous stands and total wet deposition of pollutants for current(C) and previous(P) year. (p = 0.05)

Stands	Parameter	SO4(S)	NO ₃ (N)	NH4(N)	Na	Mg	Ca	Cl	K
Pinus sylvestris L.	DF				0,87C; 0,55P;	0,87P;	0,52P;	0,74C; 0,55P;	0,73P;
Picea abies L.	DF			-0,59P;	0,79C; 0,69P;	0,72P;		0,70C; 0,63P;	0,59P;
Pinus sylvestris L.	DP	0,71C;	0,52C; -0,51P;	-0,79P;	0,92C;	0,68C; 0,53P;	0,52P;	0,92C;	0,65C;
Picea abies L.	DP	0,72C; -0,59P;	0,48C; - 0,70P;		0,71C;	0,86C;	0,53C;	0,87C;	0,84C;

Tab. 6 Coefficients of correlations between the parameters of coniferous stands and concentration pollutant in precipitation for current(C) and previous(P) year. (p = 0.05)

Stands	Parameter	SO4(S)	NO ₃ (N)	NH4(N)	Na	Mg	Ca	Cl	K
Pinus sylvestris L.	DF	0,51C;			0,73C; 0,64P;	0,93P;		0,89C; 0,52P;	0,75 P;
Picea abies L.	DF	0,61C;		-0,61P;	0,86P;	0,76P;	-0,55C;	0,83C; 0,64P;	0,62 P;
Pinus sylvestris L.	DP		-0,48P;	-0,67P;	0,65C; 0,55 P;	0,74P;	0,75 P;	0,94C; 0	0,73 P;
Picea abies L.	DP			-0,54P;			0,88P;	0,75C;	

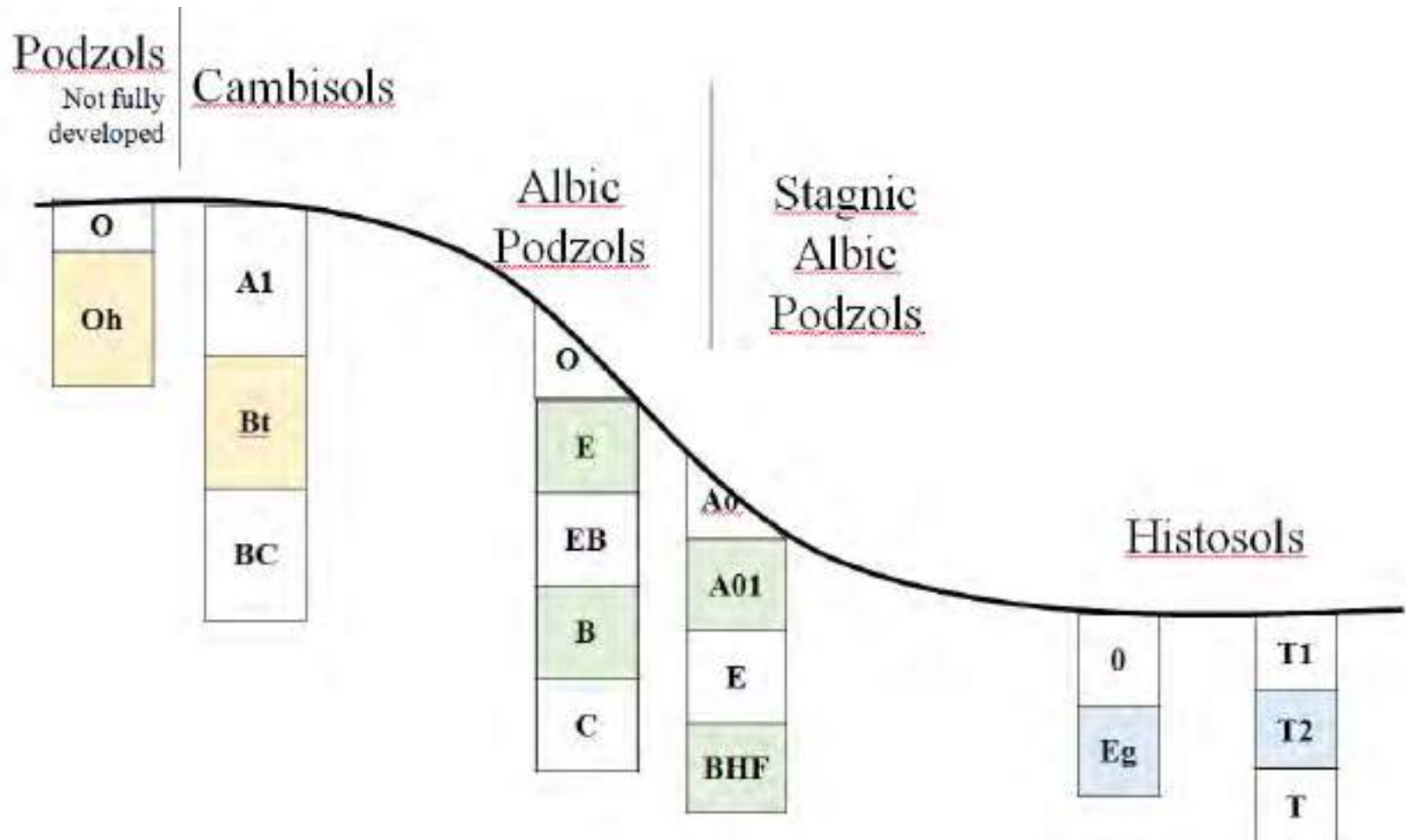
WSBS**&****Table. 7 Coefficients of correlations between the parameters of coniferous stands and total Janiskoski wet deposition of pollutants for current(C) and previous(P) year. (p = 0.05)**

Stands	Parameter	SO4(S)	NO ₃ (N)	NH4(N)	Na	K	Ca
Pinus sylvestris L.	DF	0,59 C;		0,61C;	0,79C;	0,69C;	0,69C;
Picea abies L.	DF		-0,65 C; - 0,48 P;		0,63C; 0,72 P;	0,46C; 0,57P;	0,50C; 0,59P;
Pinus sylvestris L.	DP			0,62 C;	0,83C;	0,77C;	0,80C;
Picea abies L.	DP				0,48C;	0,53C;	0,58C;

Table. 8 Coefficients of correlations between the parameters of coniferous stands and concentration pollutant in precipitation for current(C) and previous(P) year. (p = 0.05)

Stands	Parameter	SO4(S)	NO ₃ (N)	NH4(N)	Na	K	Ca
Pinus sylvestris L.	DF		-0,63C; -0,48 P;	0,62C;	0,57C; 0,66 P;	0,65C;	0,62C; 0,48 P;
Picea abies L.	DF		-0,79C;		0,91 P;	0,75 P;	0,76 P;
Pinus sylvestris L.	DP		-0,58C;	0,63C;	0,77C; 0,62 P;	0,74C;	0,71C;
Picea abies L.	DP	0,47 P;	-0,59C;		0,53 C;		

The distribution of soils at White Sea Biological Station



Conclusions

- Defoliation and depigmentation of pine and spruce changed unidirectionally for RU16 and RU04
- For all territories, discovered the significant correlations between the state of the forest and annual fluctuations of pollutants
- Coniferous forests of the North are more sensitive to fluctuations in pollutants
- Nitrogen compounds act as fertilizers for the territories of the far North
- We assume that the results of this research, could be used for verification models of the atmospheric pollutants transport for Northern area



Thank you

Recovery of benthic algal assemblages from acidification: how long does it take, and is there a link to eutrophication?

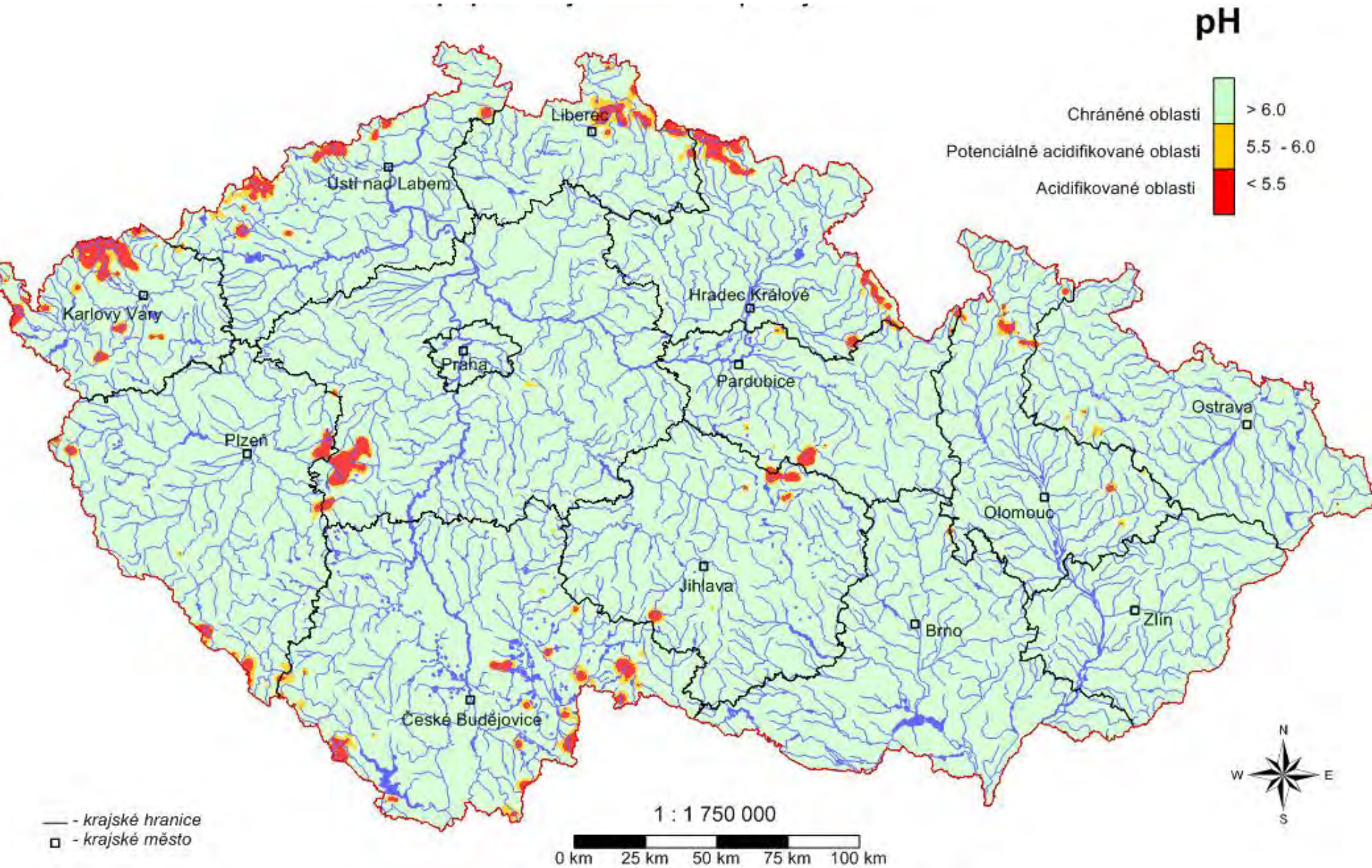
Susi Schneider¹, Filip Oulehle², Pavel Krám², Jakub Hruška²

¹Norwegian Institute for Water Research, Norway

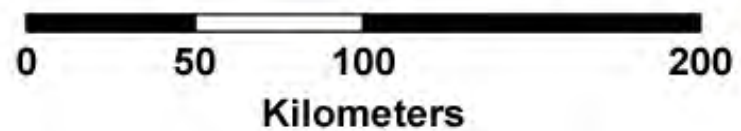
²Czech Geological Survey, Praha, Czech Republic



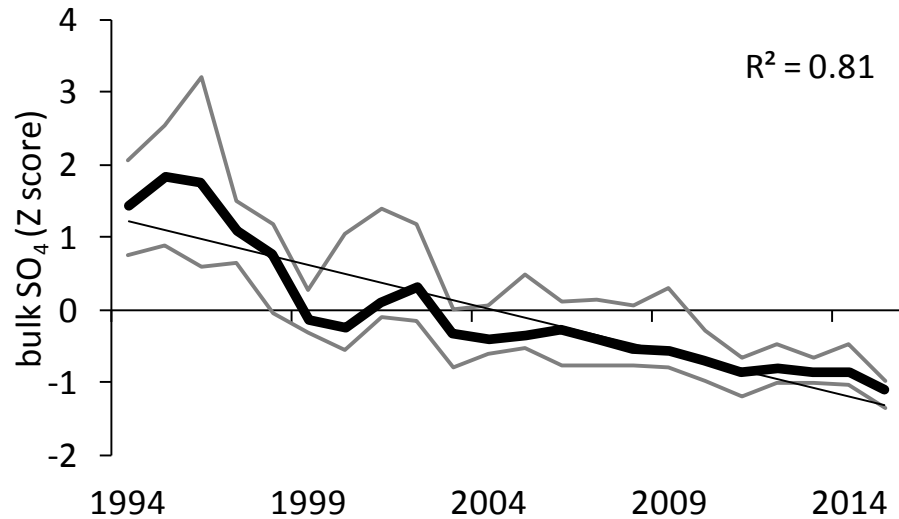
Acidified surface waters in the Czech Republic



Location of GEOMON watersheds



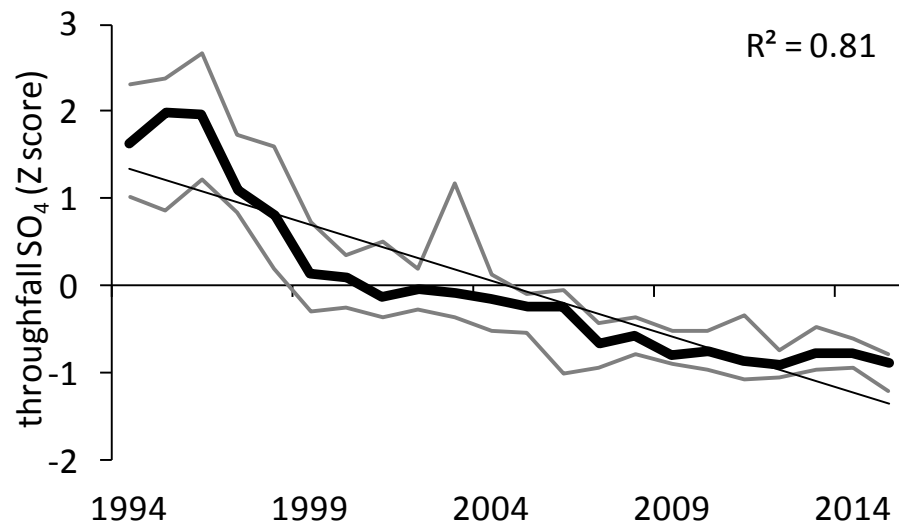
S deposition (1994 - 2015)



14 kg S ha⁻¹ yr⁻¹



3.7 kg S ha⁻¹ yr⁻¹



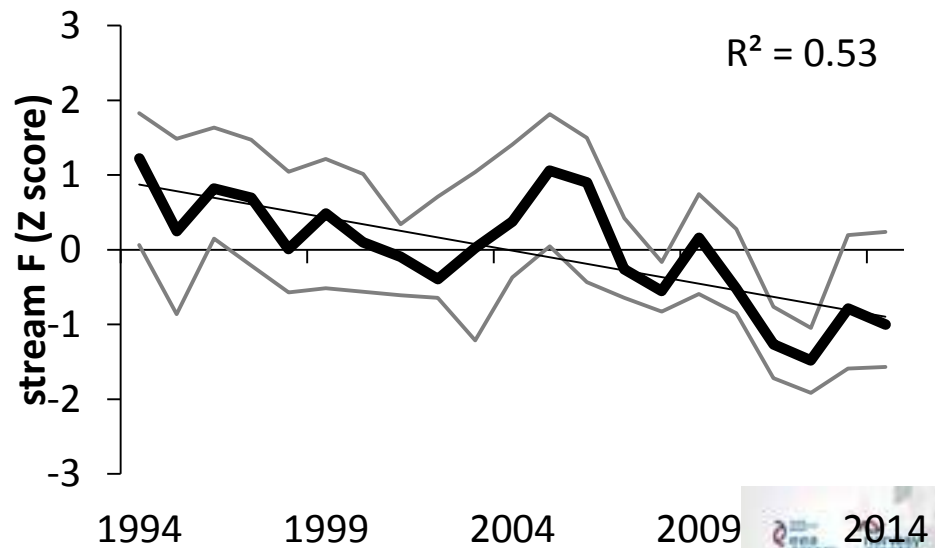
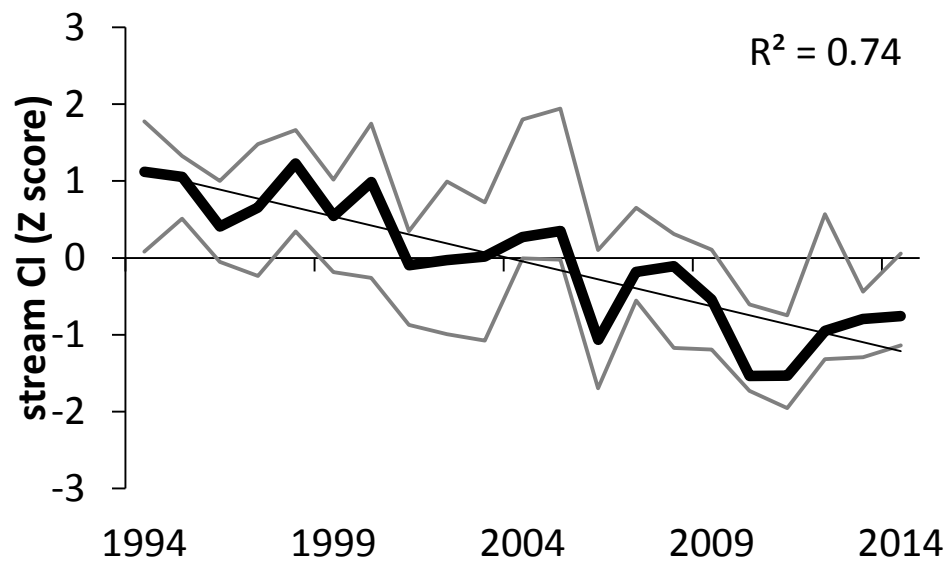
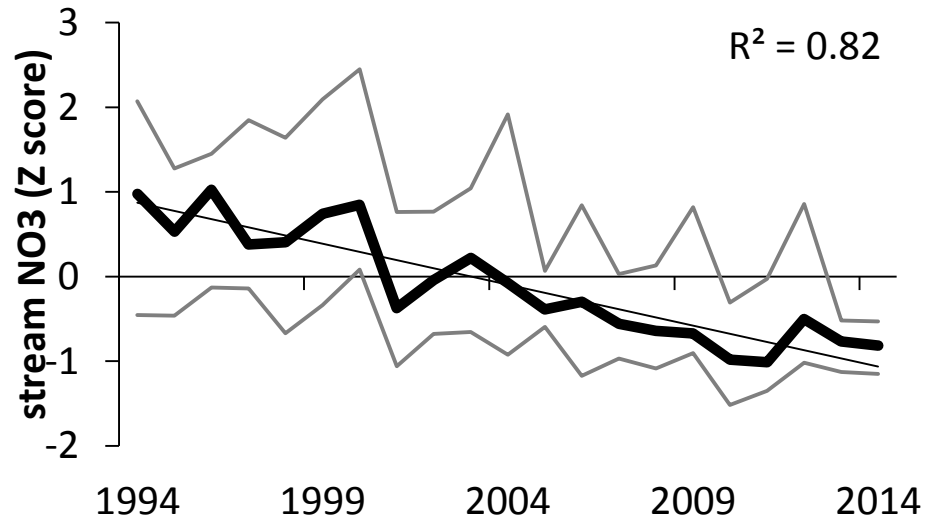
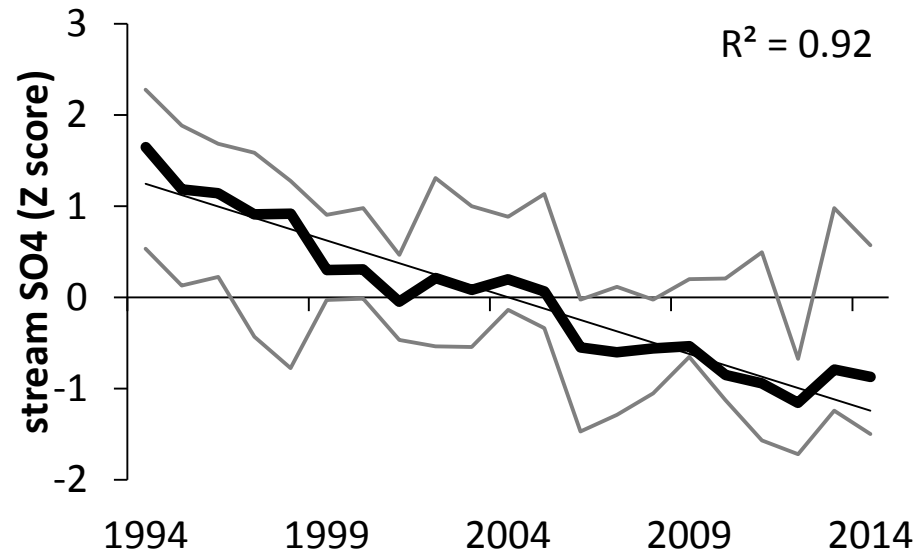
36 kg S ha⁻¹ yr⁻¹



9.7 kg S ha⁻¹ yr⁻¹

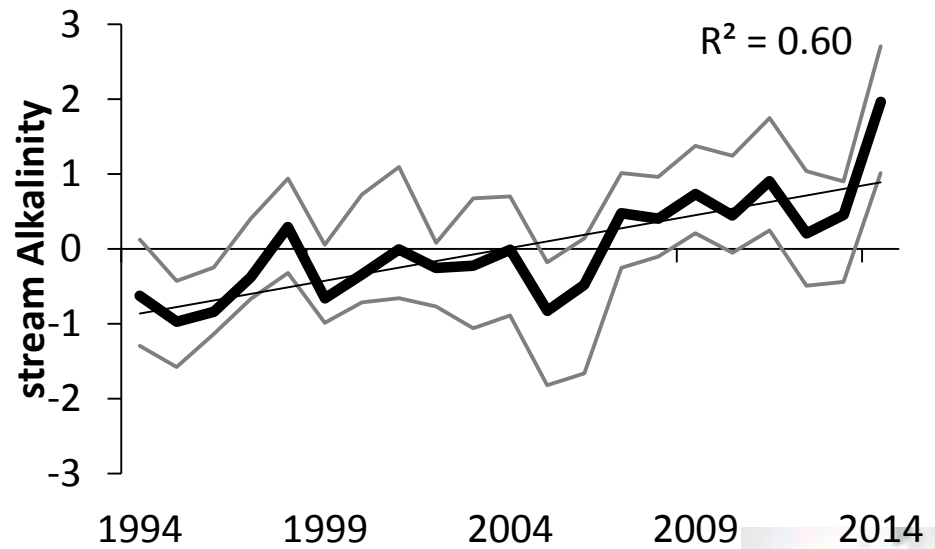
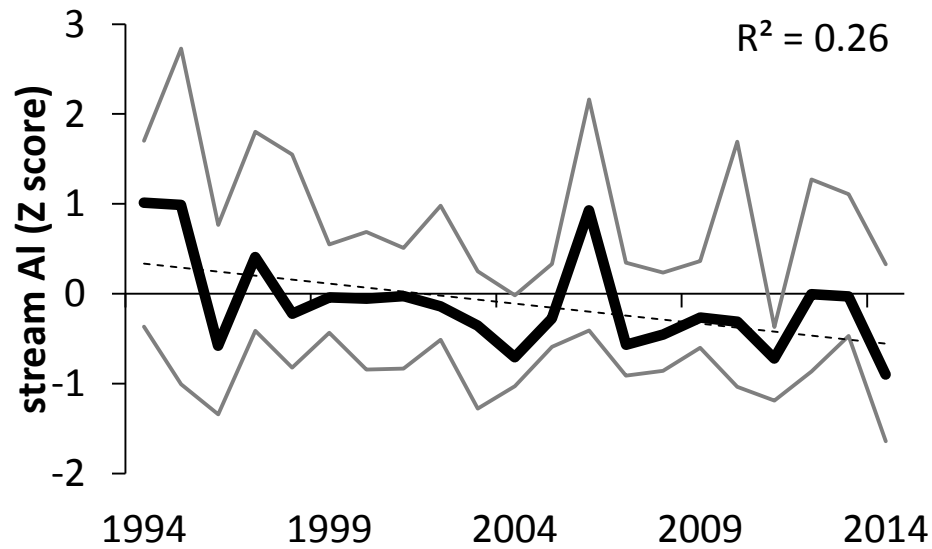
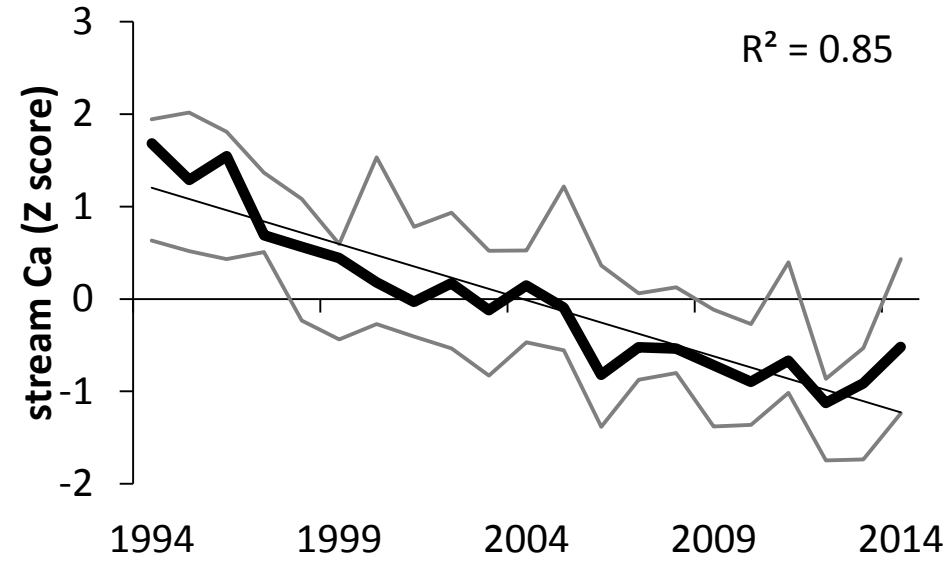
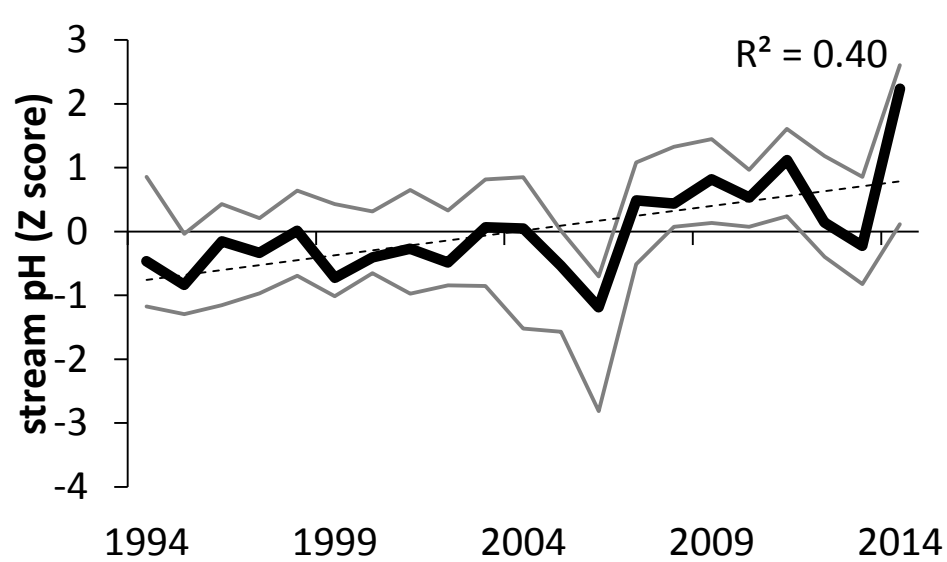
Dry deposition still important

GEOMON Runoff (anions)



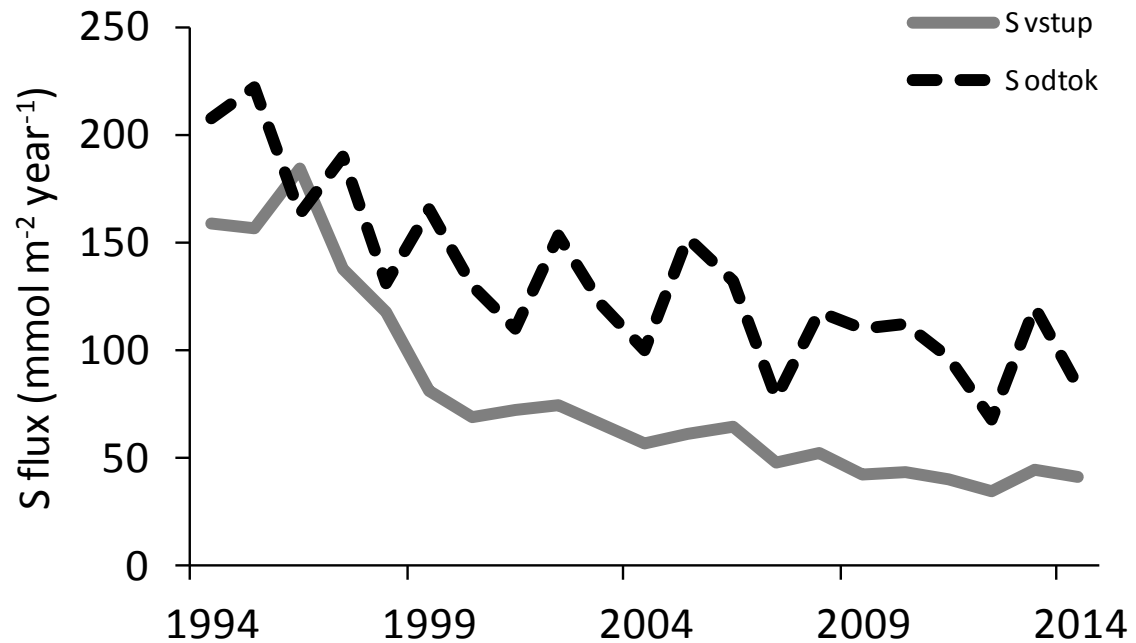
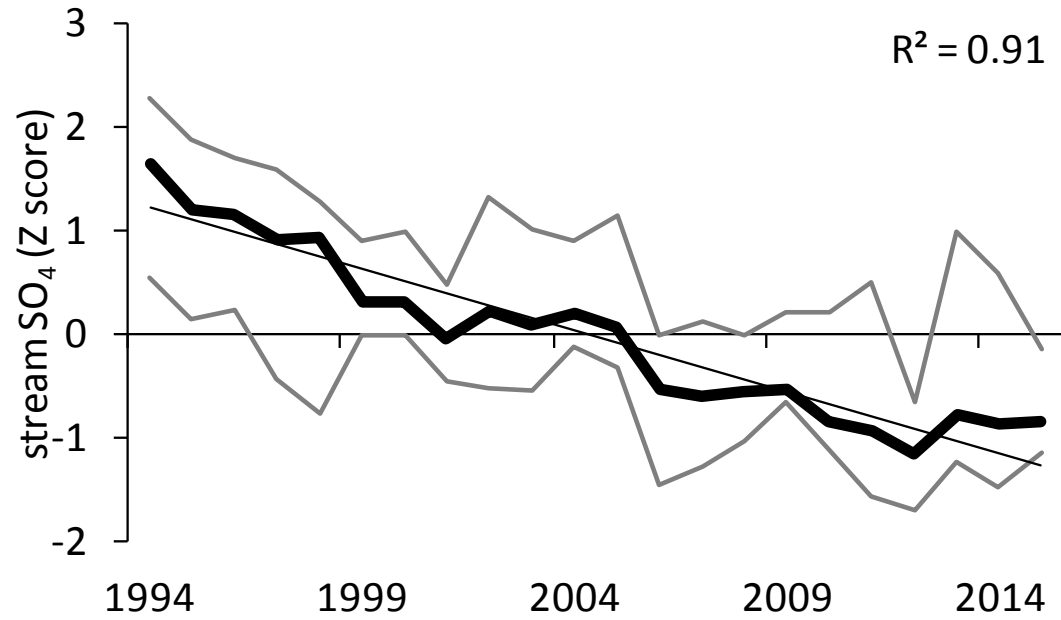
GEOMON

Runoff (cations, alk)



GEOMON

S budget



Susi Schneider



ANE CER JEZ LES LIT LIZ LKV LYS MOD NAZ PLB POM SAL UDL UHL



	ANE	CER	JEZ	LES	LIT	LIZ	LKV	LYS	MOD	NAZ	PLB	POM	SAL	UDL	UHL
cyanobacteria															
Calothrix elenkinii												xx			
Calothrix fusca												x			
Chamaesiphon confervicola			x						x						
Chamaesiphon incrustans												x			
Chamaesiphon polonicus				<1				<1							
Chroococcus spp.														x	
Heteroleibleinia spp.		xx					x					x	x		
Hydrococcus cesati											x				
Hydrococcus sp.													<1	<1	
Leptolyngbya spp.	<1		x										x		
Oscillatoria spp.		x							x						
Phormidium autumnale							<1		1						
Phormidium favosum		<1													
Phormidium inundatum															<1
Phormidium spp.												x	x	x	
Pleurocapsa minor										xxx	5				
Pseudanabaena spp.									xx						x
Pseudanabaena starmachii															
Scytonema spp.													<1		x
unidentified coccoid cyanobacteria	xxx														

green algae															
Actinotaenium cruciferum									x					x	
Closterium spp.		x		x			xx								x
Cosmarium spp.									xx						
Gongrosira spp.		2	<1		<1										
Hormidium flaccidum										5	10	x			
Hormidium rivulare					xx				<1						
Microspora palustris var minor			x		10			15							x
Microspora tumidula										x					
Microthamnion strictissimum					x				x						
Mougeotia a (6 -12u)					xx				xx					x	
Mougeotia d/e (27-36u)							x								
Oedogonium b (13-18u)		<1													
Oedogonium c (23-28u)										x			x		
Oedogonium d (29-32u)		xx													
Oedogonium e (35-43u)			x												
Spirogyra a (20-42u,1K,L)							<1								
Staurastrum spp.		x												x	x
Stigeoclonium spp.	x							15						x	
unidentified coccoid green algae					<1			<1							
Ulothrix tenerrima				xx											
Ulothrix zonata		x													
Zygogonium sp3 (16-20u)															x

chrysophytes															
Epipyxis spp.					x										

red algae															
Audouinella chalybaea			<1												
Audouinella pygmaea							x	x		x					x
Batrachospermum confusum															1
Batrachospermum spp.	xx														
unidentified red algae				x						x		x	x	x	

xanthophytes															
Vaucheria spp.			<1				<1								
AIP	n.d.	7.13	6.5	n.d.	5.8	7.2	7	5.7	6.75	7.1	7	7.3	7.1	n.d.	5.5
PIT	7.76	14.3	15.5	20	5.15	22.4	23	5.2	13.1	6.87	5.49	11.1	8.3	3.95	15.3
algal taxon richness	4	11	6	3	7	3	5	6	8	5	5	6	6	9	8
taxon richness cyanobacteria	2	4	2	0	0	2	0	1	4	1	4	4	4	4	2
taxon richness green algae	1	6	3	2	6	0	3	5	3	3	1	1	1	4	4



	PC1	PC2	PC3
Eigenvalue	7,64	5,88	4,86
Proportion Explained	0,31	0,24	0,19
Cumulative Proportion	0,31	0,54	0,74
2015			
pH	-0,30	0,37	-0,66
Na	-0,71	-0,14	0,30
K	-0,68	0,24	0,11
NH4	-0,07	0,15	-0,16
Ca	-0,77	-0,04	0,22
Cl	-0,69	-0,29	0,02
NO3	-0,17	0,38	-0,50
SO4	-0,71	-0,25	0,19
SiO2	-0,48	-0,57	-0,26
Al	0,05	-0,78	-0,01
Alk	-0,34	0,20	-0,70
cond	-0,64	-0,38	-0,18
DOC	0,19	-0,71	-0,23
Pb	0,15	-0,54	-0,48
P	0,18	-0,67	-0,13
1995-2014			
pH_20	-0,29	0,30	-0,66
pH.slope	0,40	0,11	0,08
NH4_20	-0,11	0,41	-0,36
NH4.slope	-0,36	-0,48	-0,35
Ca_20	-0,77	0,08	0,28
Ca.slope	-0,67	0,02	0,37
NO3_20	-0,50	0,10	-0,31
NO3.slope	0,28	-0,29	-0,27
Al_20	0,15	-0,74	0,10
Al.slope	0,06	-0,18	-0,70

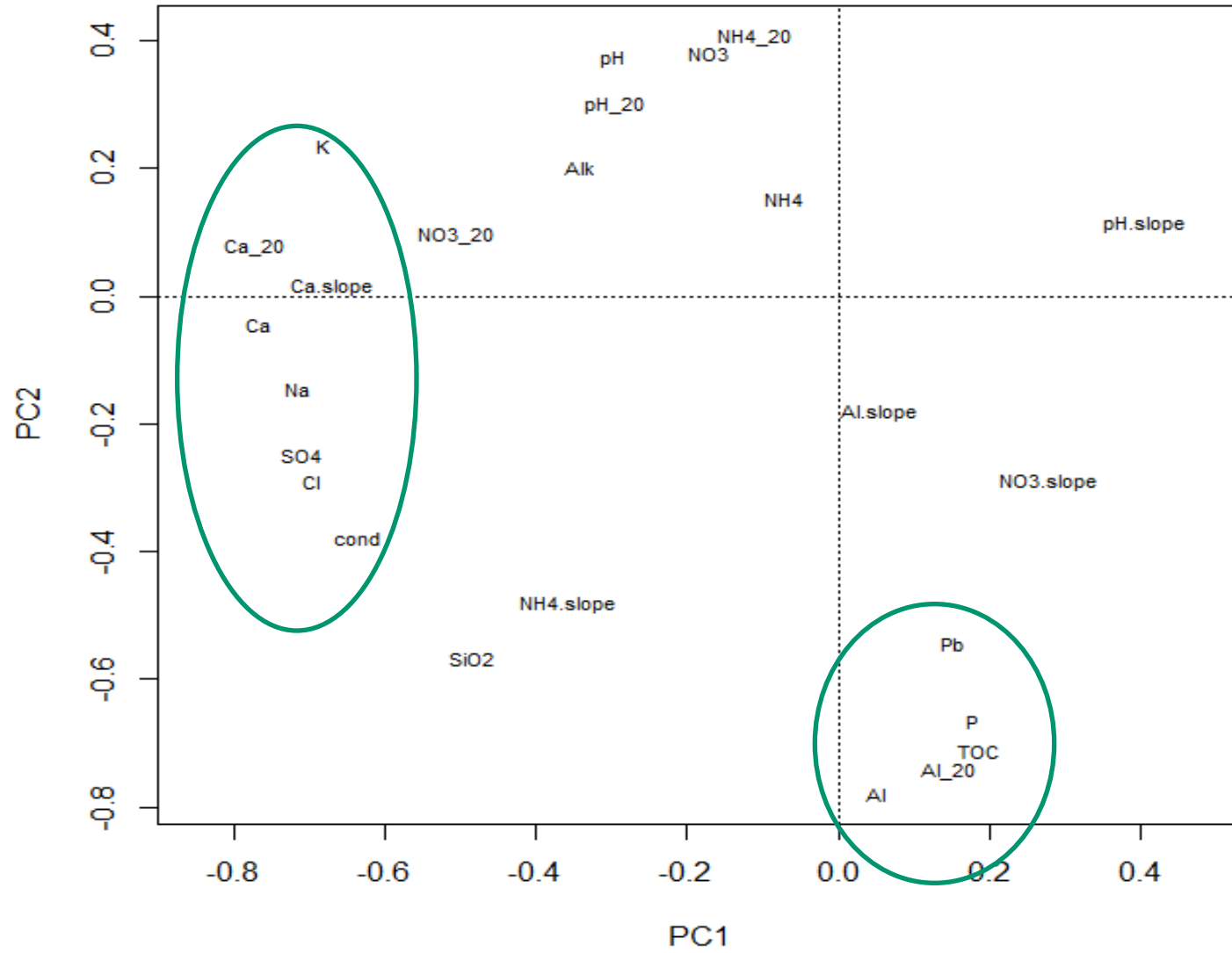
Main component analysis – streamwater chemistry

PC1 – ionic strength

PC2 – DOC

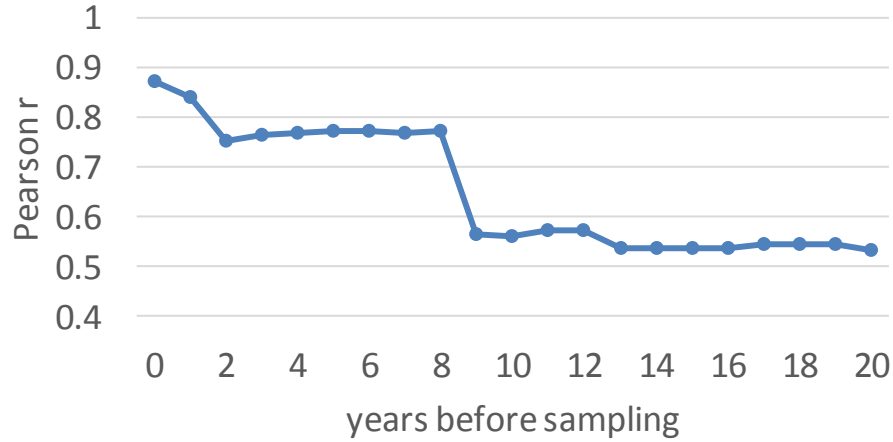
PC3 – acidity

All together -74% of variability

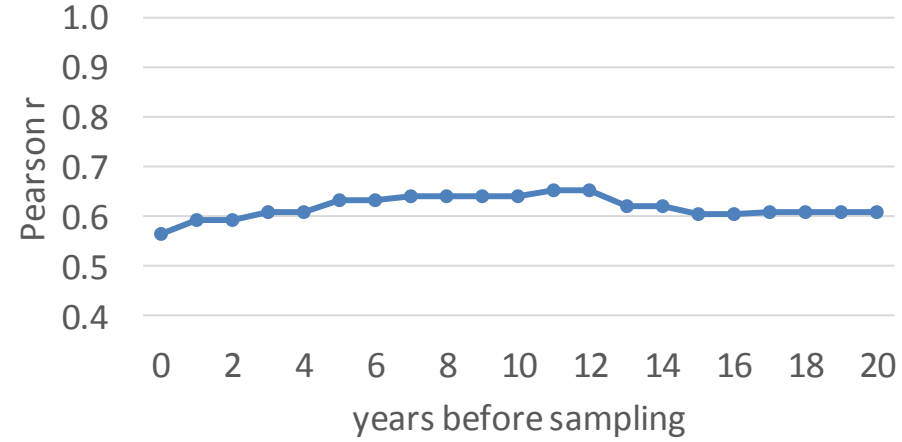


Correlation coefficients (Pearson r) between AIP (acidification index periphyton) of algal assemblages and pH, calculated for the time intervals before the algal sampling

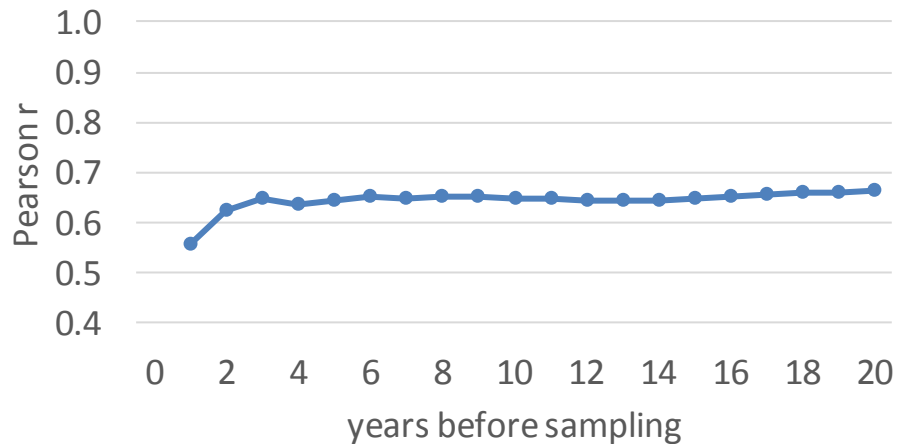
AIP - min pH



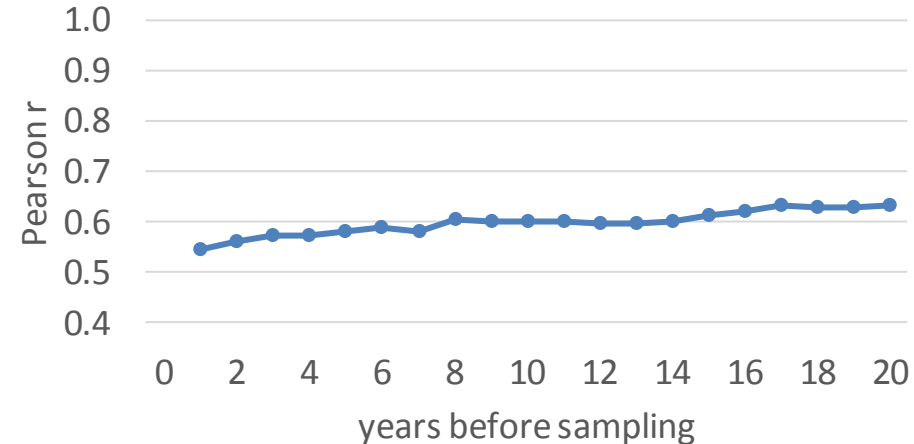
AIP - max pH



AIP - mean pH



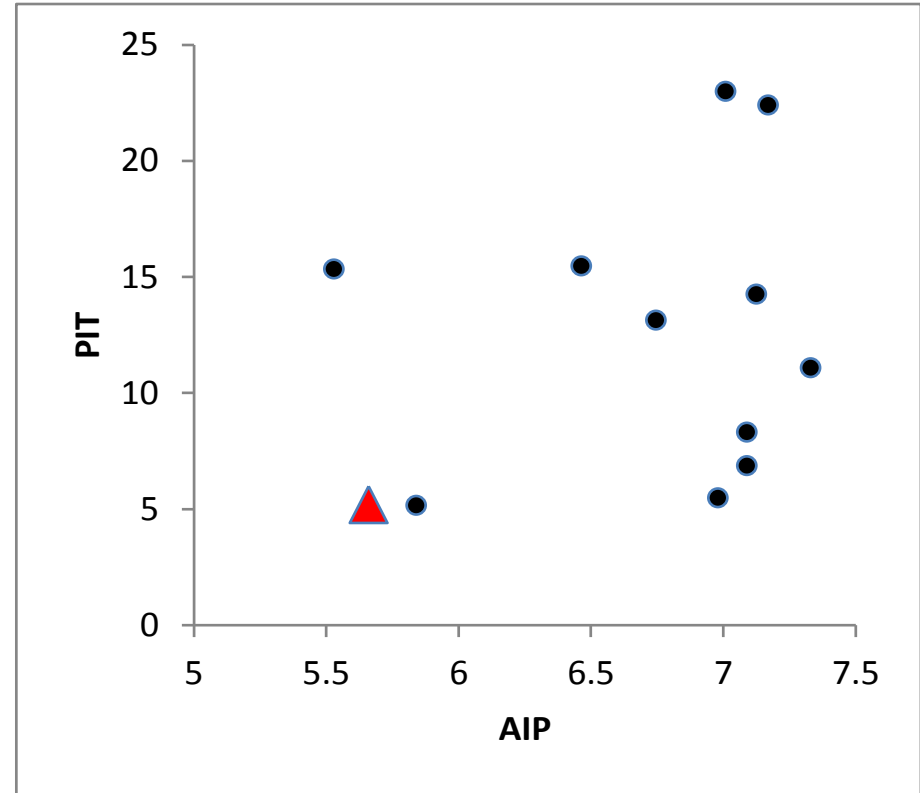
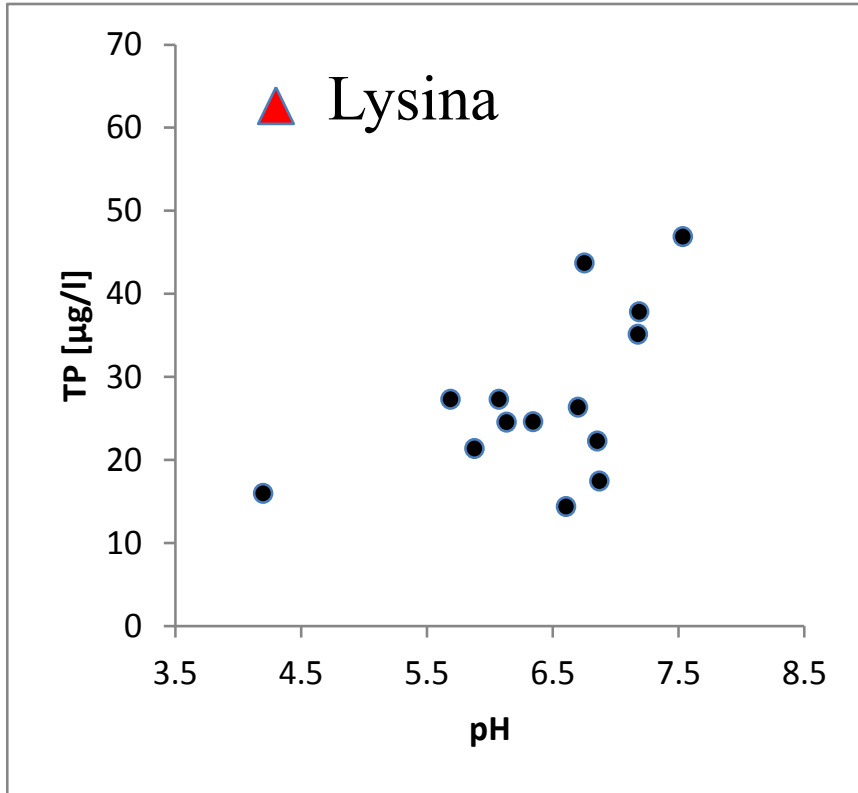
AIP - median pH



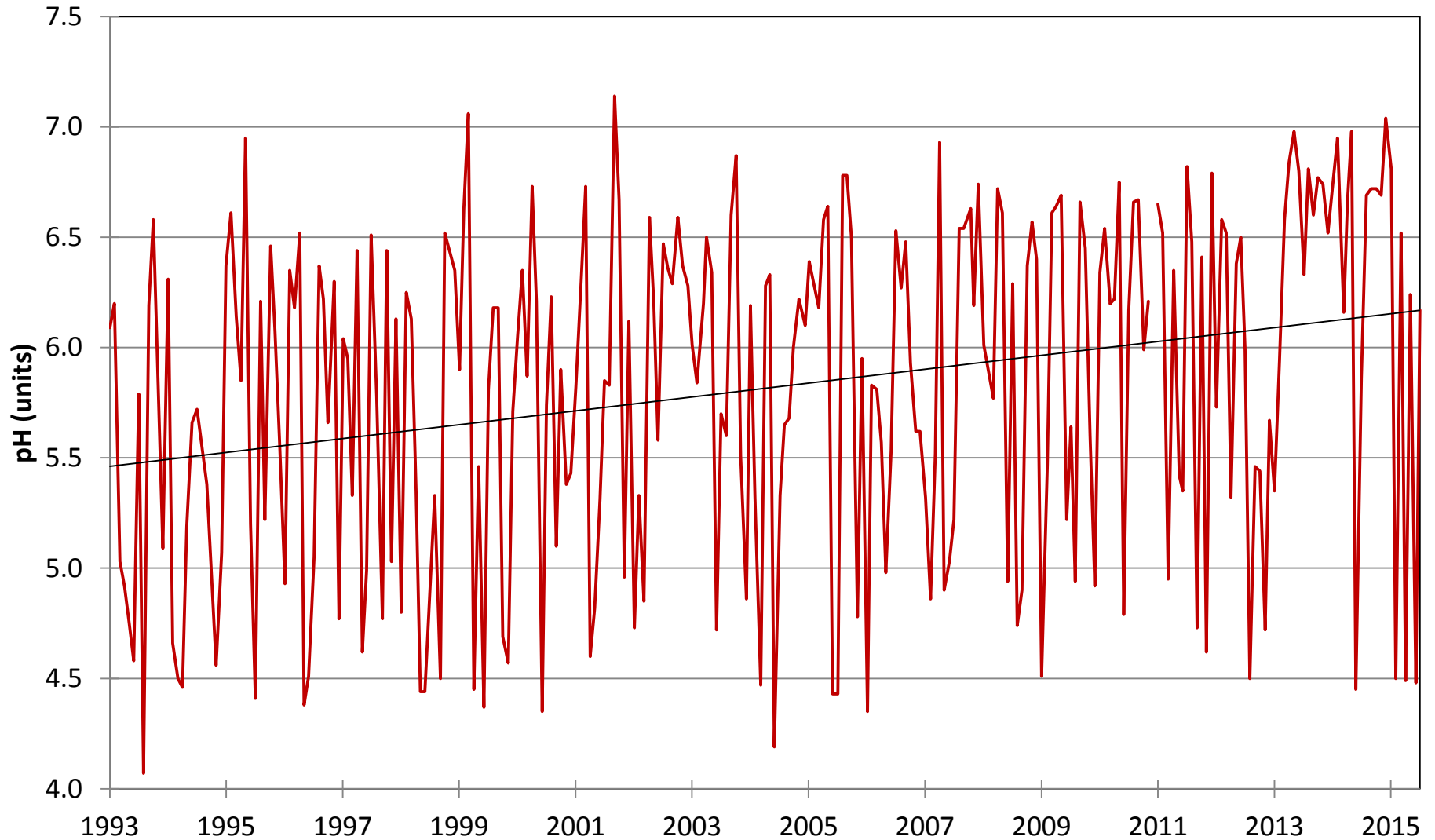
		AIP	PIT	taxon richness cyano- bacteria	taxon richness green algae	total algal taxon richness
catchment characteristics	catchment area	-0,03	0,39	0,05	0,25	0,37
	mean elevation	-0,18	0,04	0,37	0,27	0,46
	min_elevation	-0,31	-0,17	0,35	0,25	0,42
	max_elevation	-0,08	0,17	0,35	0,25	0,46
	Avg Temp	0,28	0,05	-0,41	-0,30	-0,50
	forested area	0,03	0,20	0,10	0,31	0,40
	longitude (X_WGS)	0,28	0,20	0,45	0,24	0,64
	latitude (Y_WGS)	-0,50	-0,16	0,11	0,25	0,32
2015	pH	0,75	0,13	0,66	-0,66	-0,08
	Na	0,36	0,39	-0,29	-0,52	-0,57
	K	0,48	0,11	0,07	-0,35	-0,18
	NH4	-0,50	0,13	-0,05	0,15	0,19
	Ca	0,50	0,40	-0,25	-0,44	-0,45
	Cl	0,40	-0,05	-0,32	-0,48	-0,66
	NO3	0,35	-0,08	0,64	-0,44	0,06
	SO4	0,22	0,29	-0,52	-0,25	-0,51
	SiO2	0,33	0,02	-0,15	-0,72	-0,77
	Al	-0,65	-0,45	-0,59	0,22	-0,32
	Alk	0,66	0,16	0,56	-0,67	-0,17
	cond	0,43	0,09	-0,26	-0,51	-0,62
	DOC	-0,33	-0,55	-0,24	0,04	-0,28
	Pb	-0,47	-0,43	0,04	-0,09	-0,16
P	-0,14	-0,43	-0,08	-0,14	-0,31	
1995-2014	pH_20	0,79	-0,06	0,61	-0,55	-0,06
	pH.slope	-0,34	0,11	-0,02	0,34	0,28
	NH4_20	0,28	-0,14	0,61	-0,18	0,32
	NH4.slope	0,24	-0,19	-0,19	-0,61	-0,66
	Ca_20	0,37	0,37	-0,41	-0,26	-0,35
	Ca.slope	0,05	0,02	-0,40	-0,10	-0,28
	NO3_20	0,29	-0,29	0,25	-0,42	-0,19
	NO3.slope	0,15	0,12	-0,02	-0,08	-0,16
	Al_20	-0,63	-0,35	-0,54	0,04	-0,40
Al.slope	0,38	-0,20	0,40	-0,41	-0,15	
PCA	PC1	-0,56	-0,15	0,25	0,51	0,49
	PC2	0,39	0,30	0,57	0,11	0,56
	PC3	-0,39	0,33	-0,62	0,51	0,08



Stream chemistry, PIT (nutrient) and AIP (acidification) indices



Uhlířská (ICP Waters) - lowest AIP



Conclusions

- acidification more important than eutrophication
- episodes more important than long-term averages
- expected results, but not much documented

Schneider et al. (in review, Hydrobiologia)

Soil modelling study VSD+

Maria Holmberg

Finnish Environment Centre SYKE, Helsinki

Joint TFM ICP IM & ICP Waters

Uppsala 9–11.5.2017

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Model development: Luc Bonten, Janet Mol, Max Posch, Gert Jan Reinds, Wieger Wamelink

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We acknowledge the support of Swedish EPA and our institutes and those of our collaborators as well as support of the eLTER project (EU/H2020 grant agreement No. 654359).



Outline

Background, rationale

Modelling setup

Site locations 40 sites

Calibrations 14 sites

Simulations 10 sites

Next tasks

On-going pressures

Climate change

- Warming: earlier start of growth, later end of growth
- Seasonal or annual drought
- Heavy rainfall events, drought spells

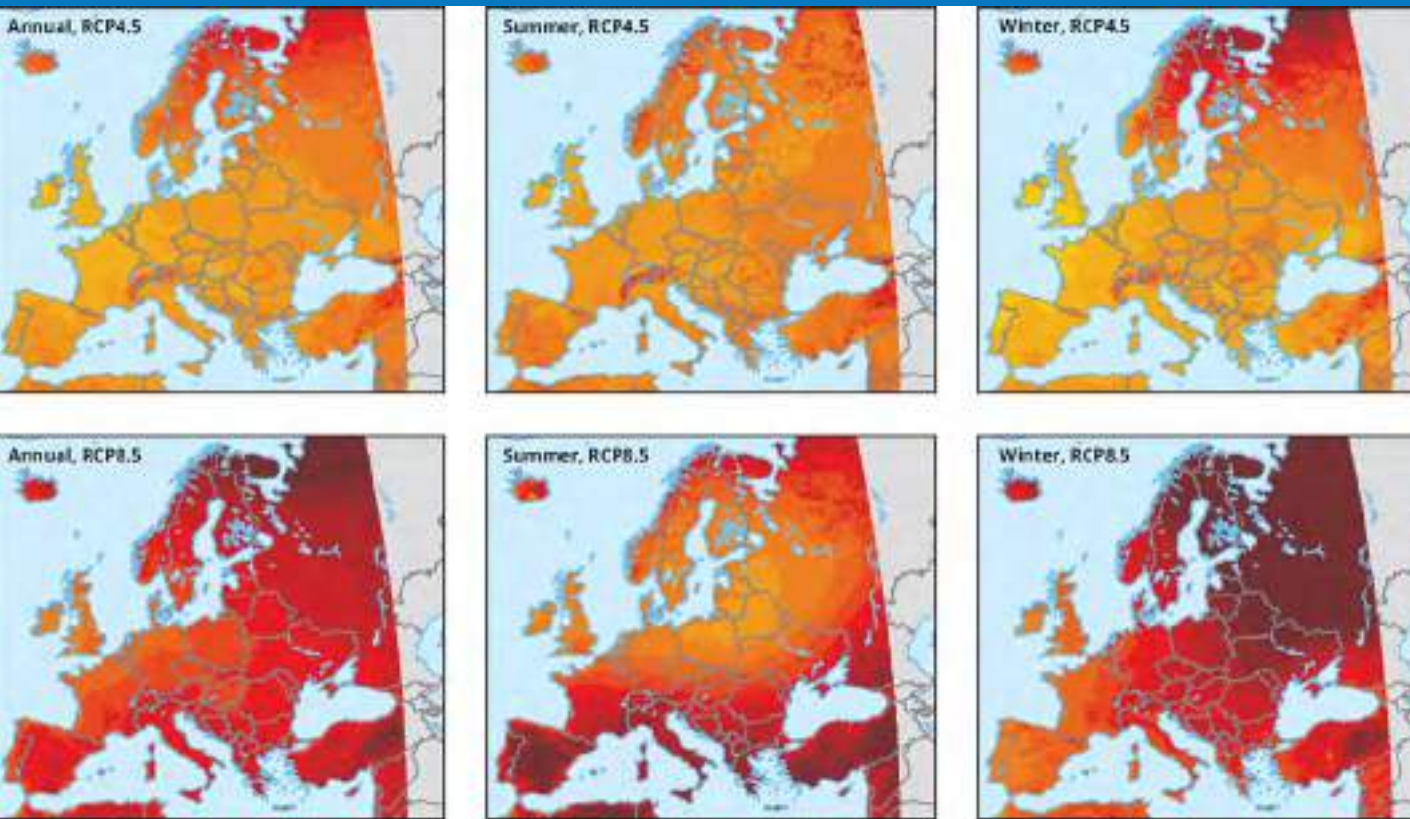
Air pollution

- Sulphur shows decreasing trends
- Nitrogen stable or some increase

Management

- Increased use of bioenergy, e.g. to mitigate climate change
- Decreased motivation for protection of biodiversity

Spatial variation in future climate: Temp.

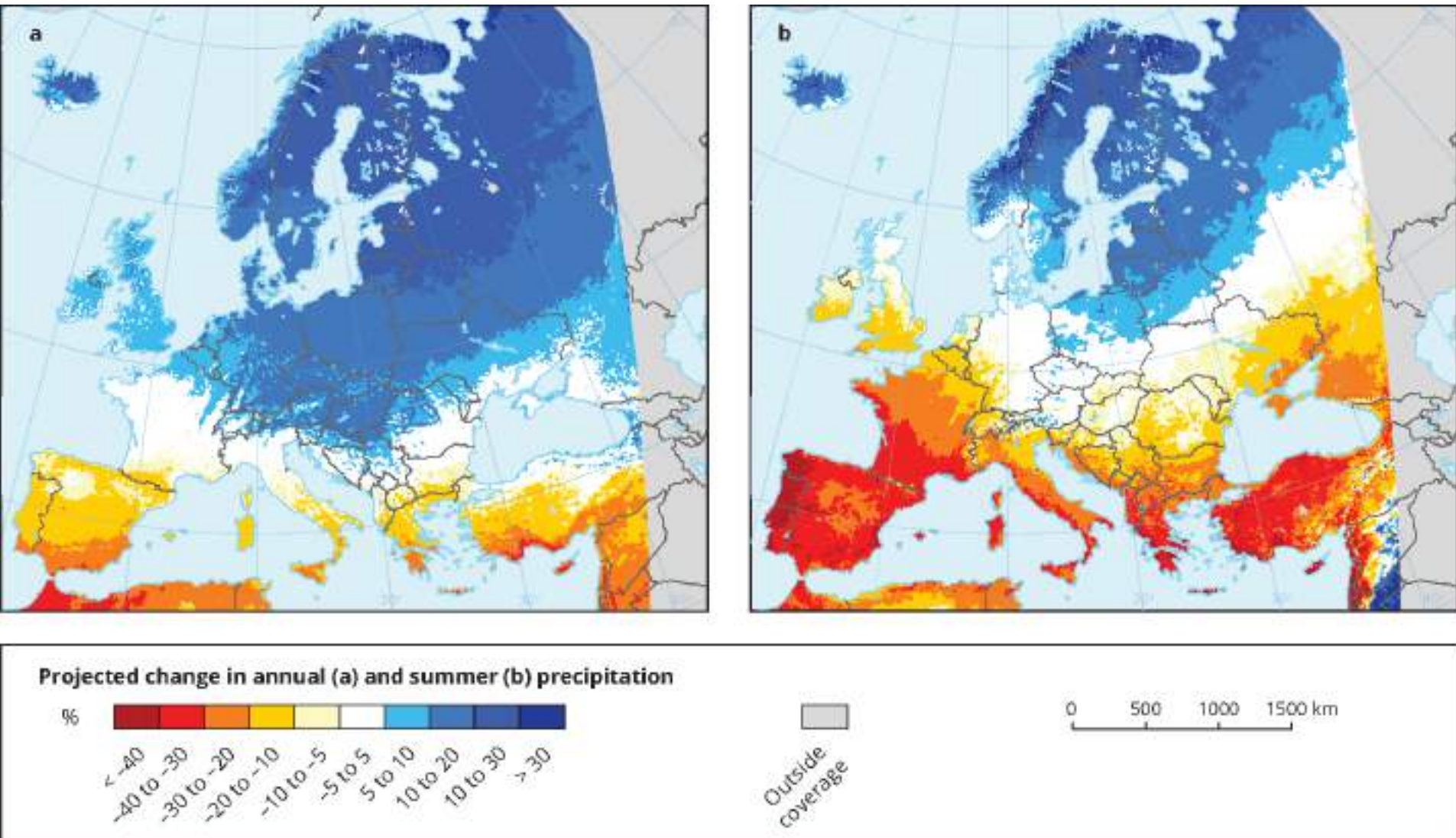


Projected change in annual, summer and winter temperature for the forcing scenarios RCP4.5 and RCP8.5



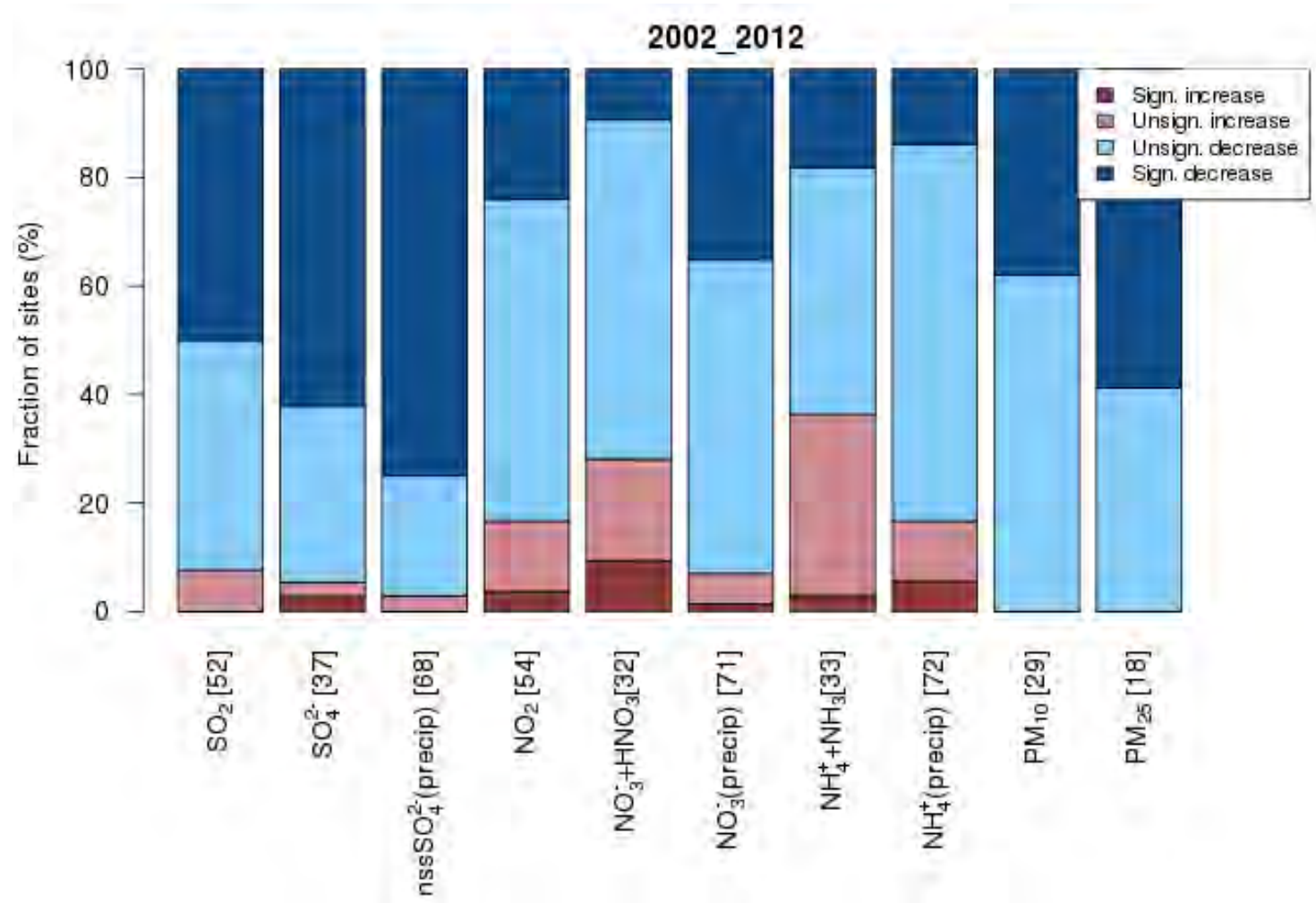
Projected changes in annual (left), summer (middle) and winter (right) near-surface air temperature (°C) in the period 2071-2100, compared to the baseline period 1971-2000 for the forcing scenarios RCP 4.5 (top) and RCP 8.5 (bottom). Model simulations are based on the multi-model ensemble average of RCM simulations from the EURO-CORDEX initiative. <http://www.euro-cordex.net/>

Spatial variation in future climate: Prec.



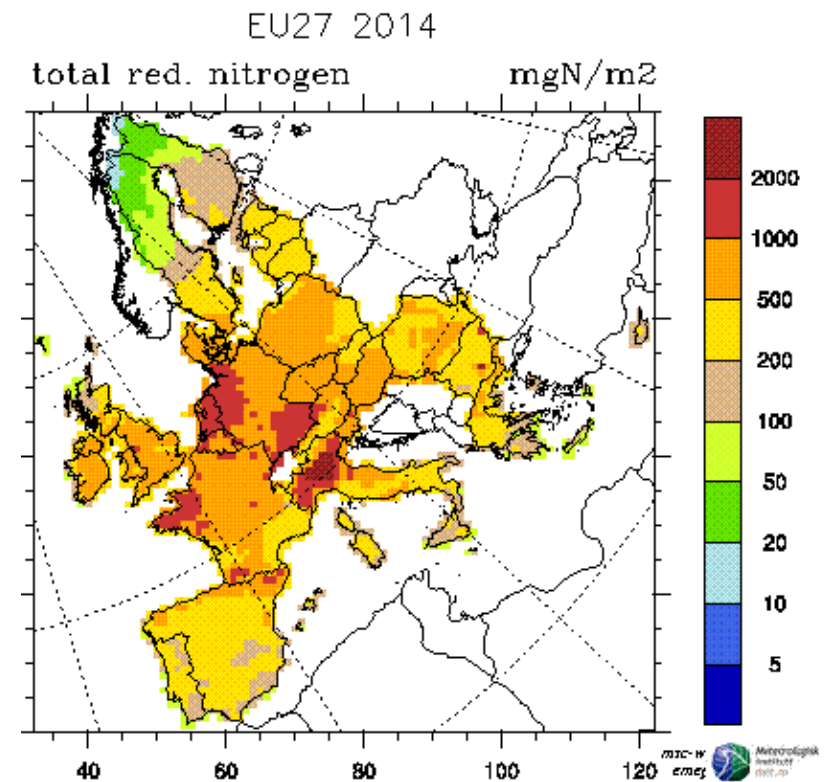
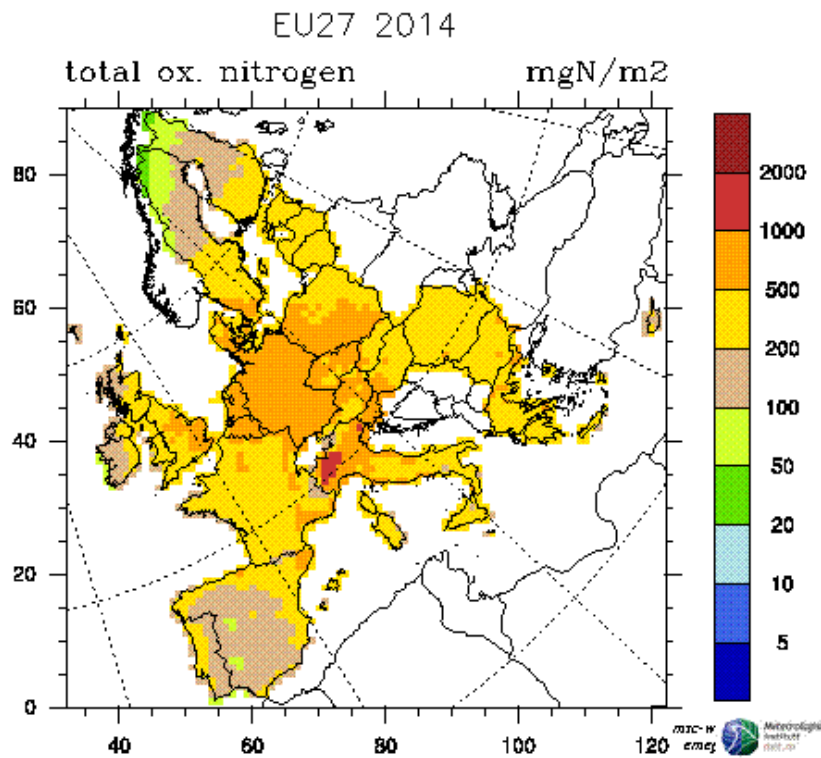
Projected changes in annual (left) and summer (right) precipitation (%) in the period 2071-2100 compared to the baseline period 1971-2000 for the forcing scenario RCP 8.5. Model simulations are based on the multi-model ensemble average of RCM simulations from the EURO-CORDEX initiative. <http://www.euro-cordex.net/>

Deposition trends: S decreased, N remains concern



EMEP Task Force on Measurement and Modelling Trend Report, Final Draft:
http://www.nilu.no/projects/ccc/tfmm/utrecht_2016/TFMM_TrendAR_20160513.pdf

Spatial variation in N deposition



EMEP SR 2015 <http://www.emep.int/mscw/>

Soil and habitat responses

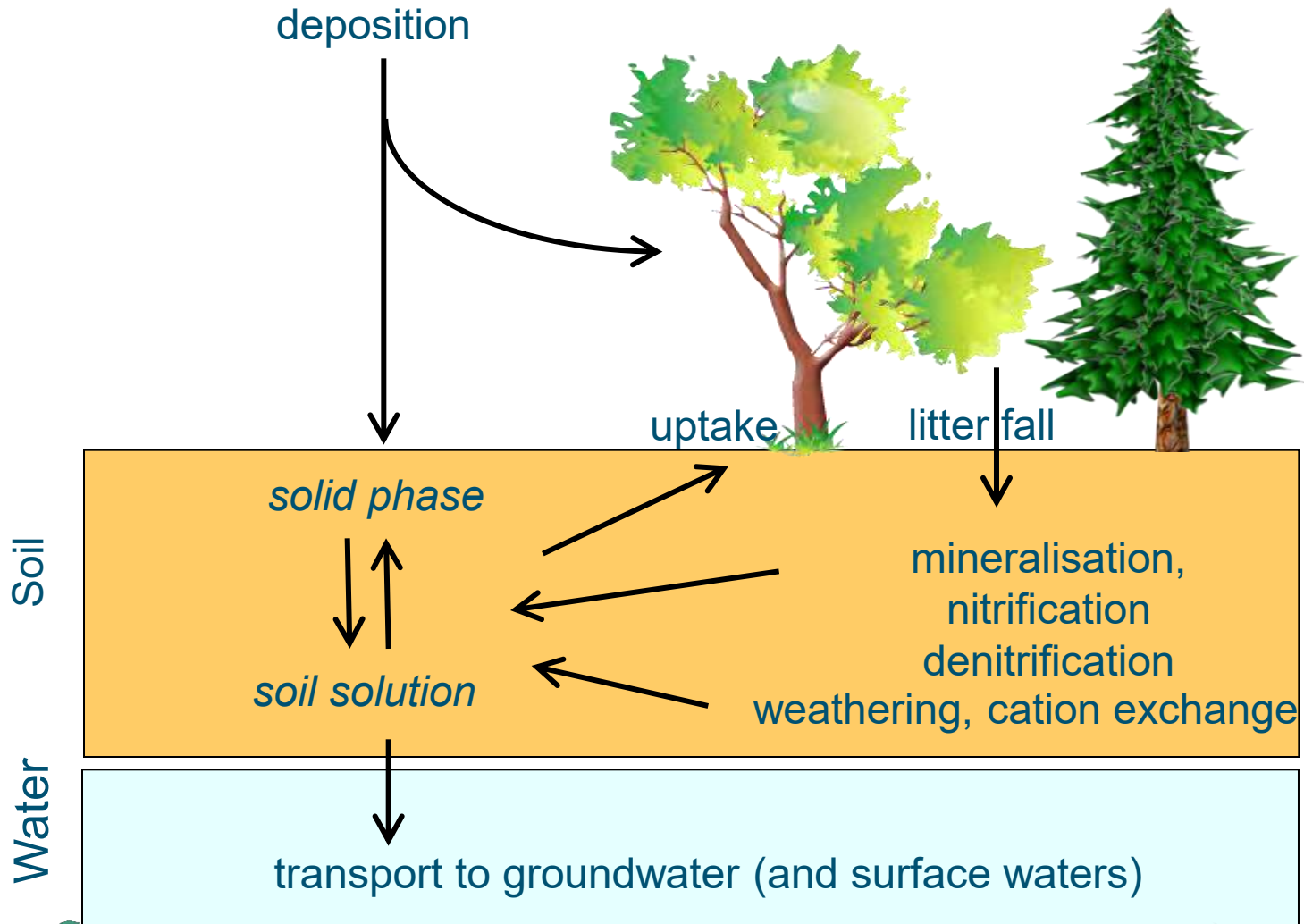
Air pollution, climate change, management

- Affect vegetation on the species and habitat levels
- Varies with region, latitude, policy
- Understanding soil dynamics will help evaluate vegetation response

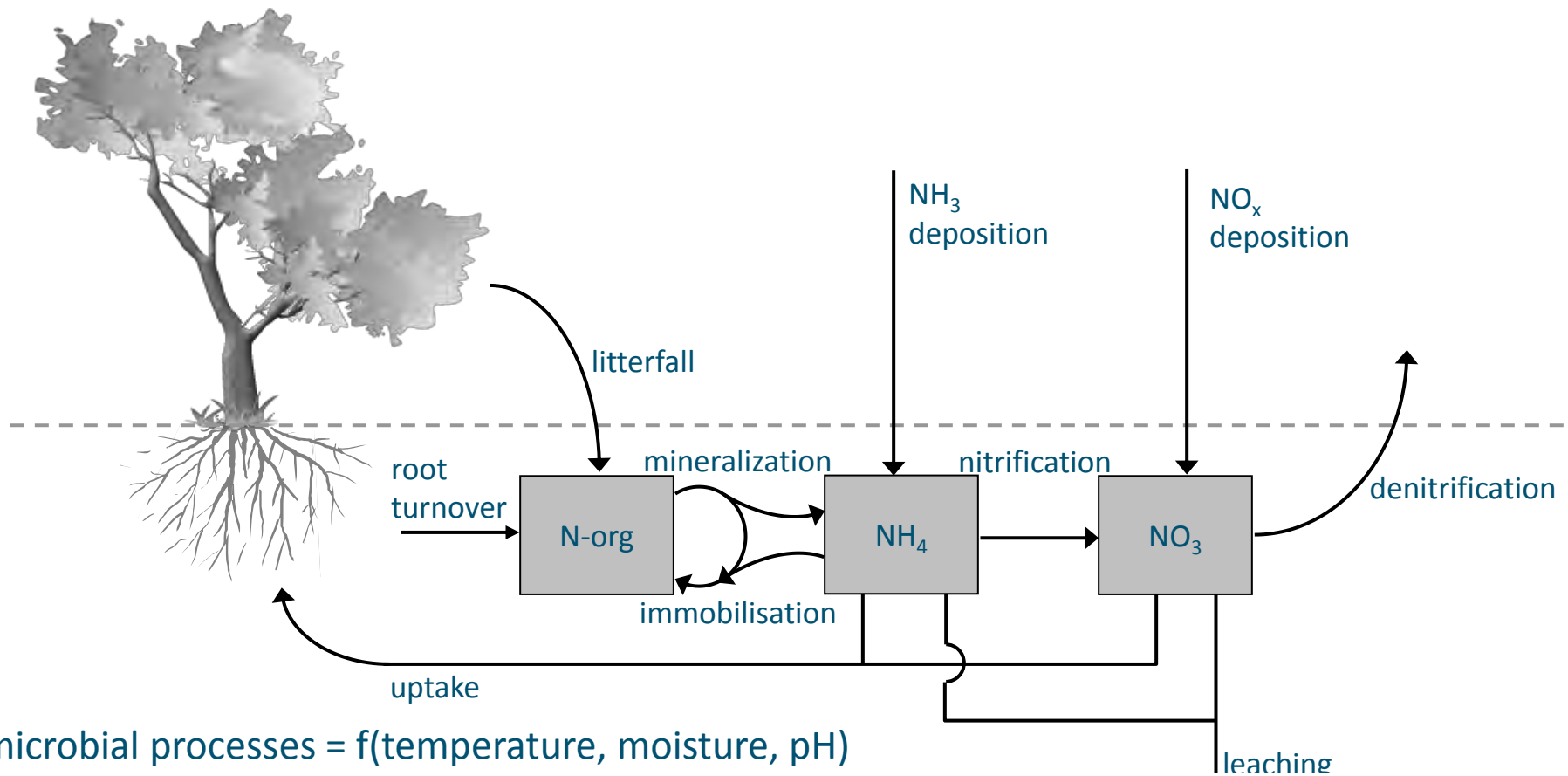
Overall aim to test VSD+ model chain to evaluate habitat responses

- Part I: Soil abiotic responses (Maria Holmberg et al.)
- Part II: Habitat responses (Thomas Dirnböck et al.)

VSD+



N processes in VSD+



microbial processes = $f(\text{temperature, moisture, pH})$

Bonten et al. 2016. Env Soft 79: 75-84.

<http://dx.doi.org/10.1016/j.envsoft.2016.01.009>

RothC 26.3 model in VSD+

Carbon pools

1. Decomposable plant material
2. Resistant plant material
3. Microbial biomass
4. Humified organic matter
- 5. *Inert organic matter***

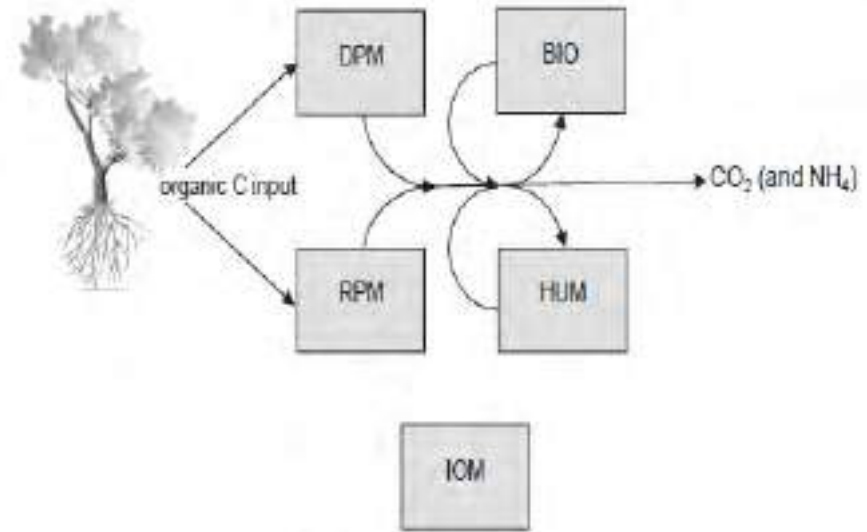
First-order processes for turnover of pools 1-4. Temperature, soil moisture and soil cover modify rates.

Bonten et al. 2016. *Env Soft* 79: 75-84.

<http://dx.doi.org/10.1016/j.envsoft.2016.01.009>

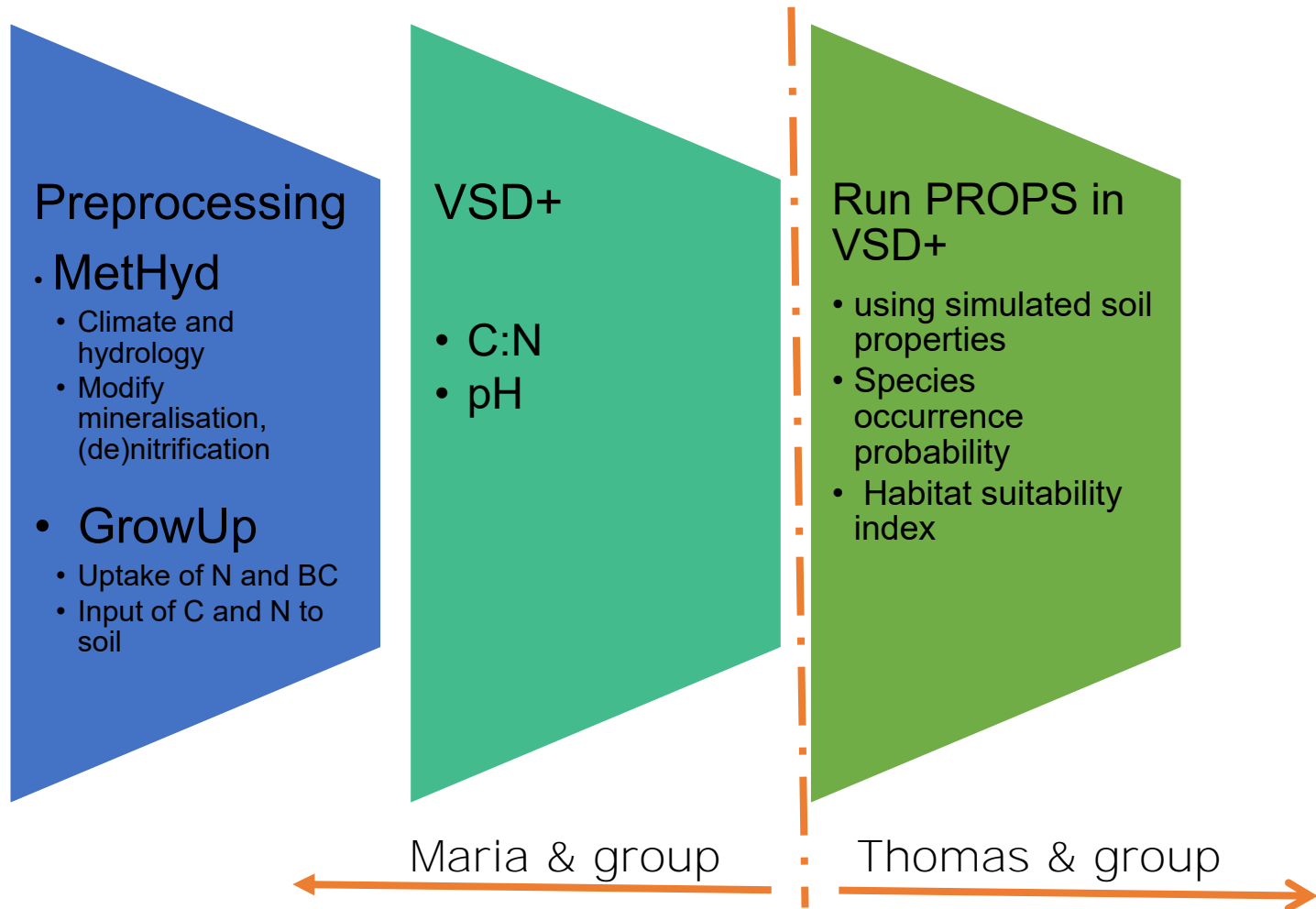
Coleman & Jenkinson 2104

http://rothamsted.ac.uk/sites/default/files/users/kcoleman/RothC_guide_DOS.pdf)

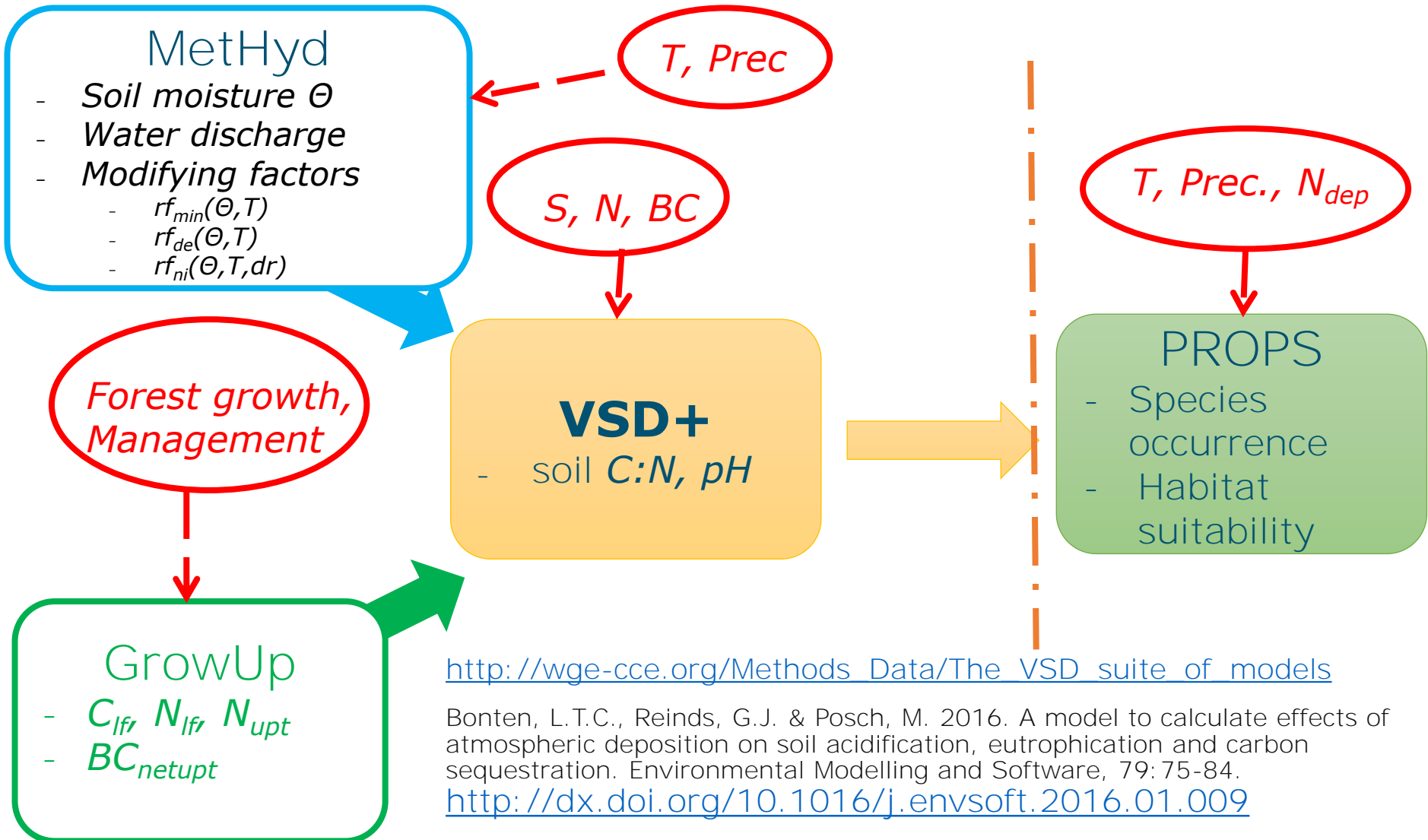


VSD+ model chain

<http://wge-cce.org/Methods Data/The VSD suite of models>



VSD+ model chain



http://wge-cce.org/Methods_Data/The_VSD_suite_of_models

Bonten, L.T.C., Reinds, G.J. & Posch, M. 2016. A model to calculate effects of atmospheric deposition on soil acidification, eutrophication and carbon sequestration. Environmental Modelling and Software, 79: 75-84.

<http://dx.doi.org/10.1016/j.envsoft.2016.01.009>

Objectives – soil modelling

Study impacts of N deposition and climate warming on soil conditions

Run VSD+ with deposition and climate scenarios

- Primary input
 - S, N dep
 - annual precip., annual temperature
 - Forest growth
- Primary output
 - Soil pH, (BS), C:N
- Key variables delivered to vegetation response analysis
 - N dep, temperature, precipitation
 - Soil pH, C:N

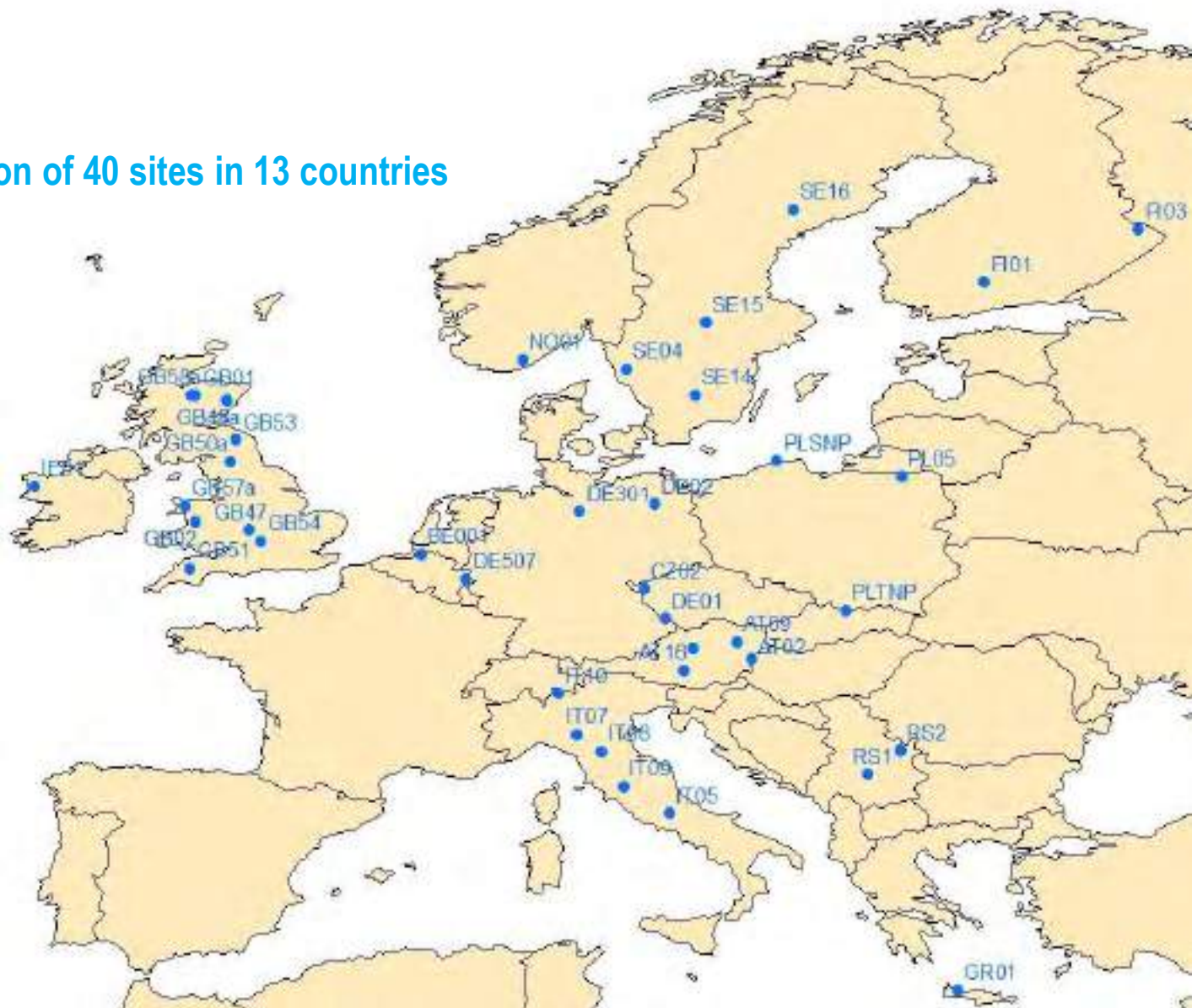
Illustrate regional variability in response

- Analyze cross-site differences

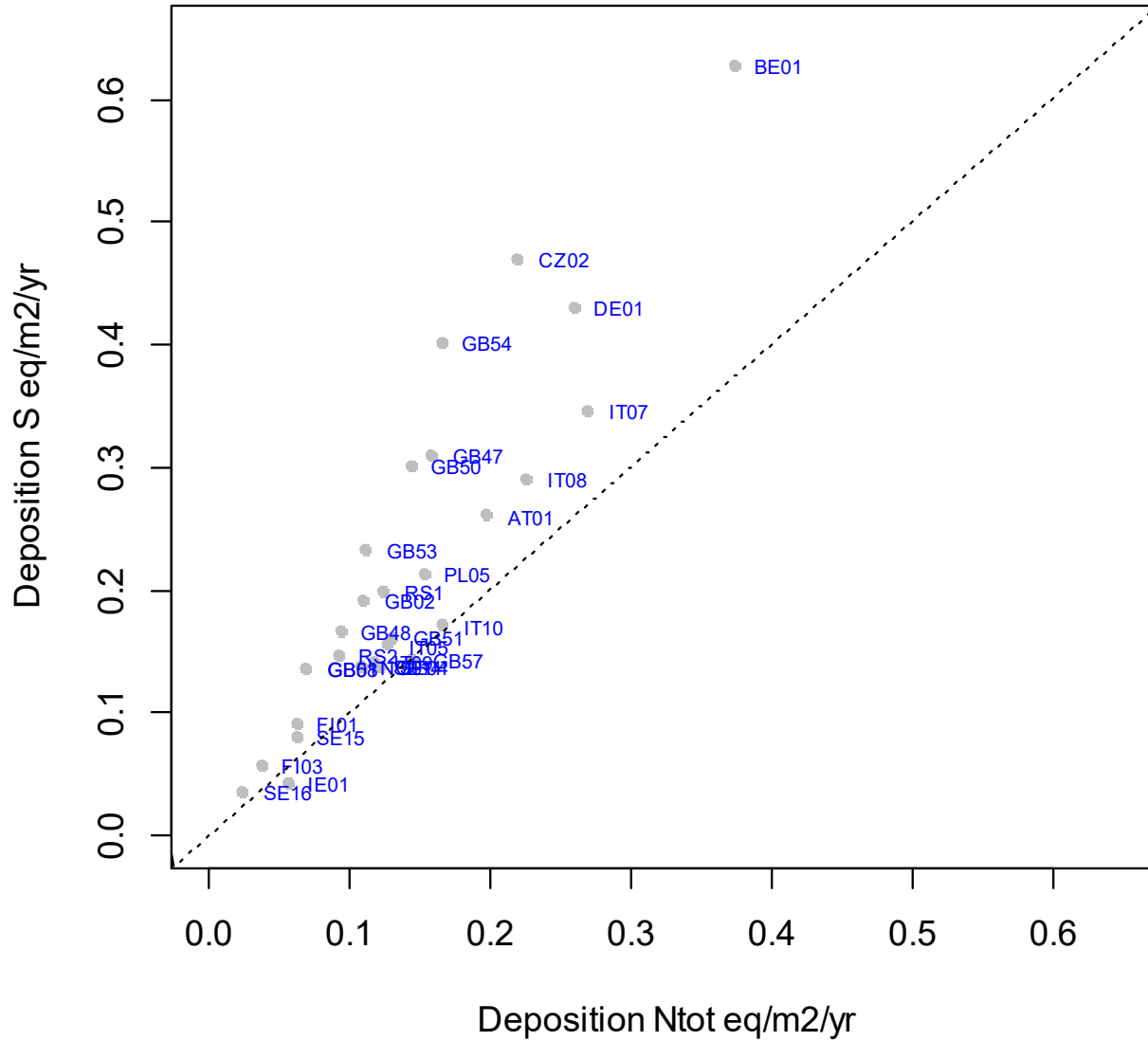
Work towards manuscript to special issue in STOTEN on theme “Detecting and explaining natural and anthropogenic changes by making use of large extent, long-term ecological research facilities of the international long-term ecosystem research (ILTER) network”. (dl July 2017)

Potential participation of 40 sites in 13 countries

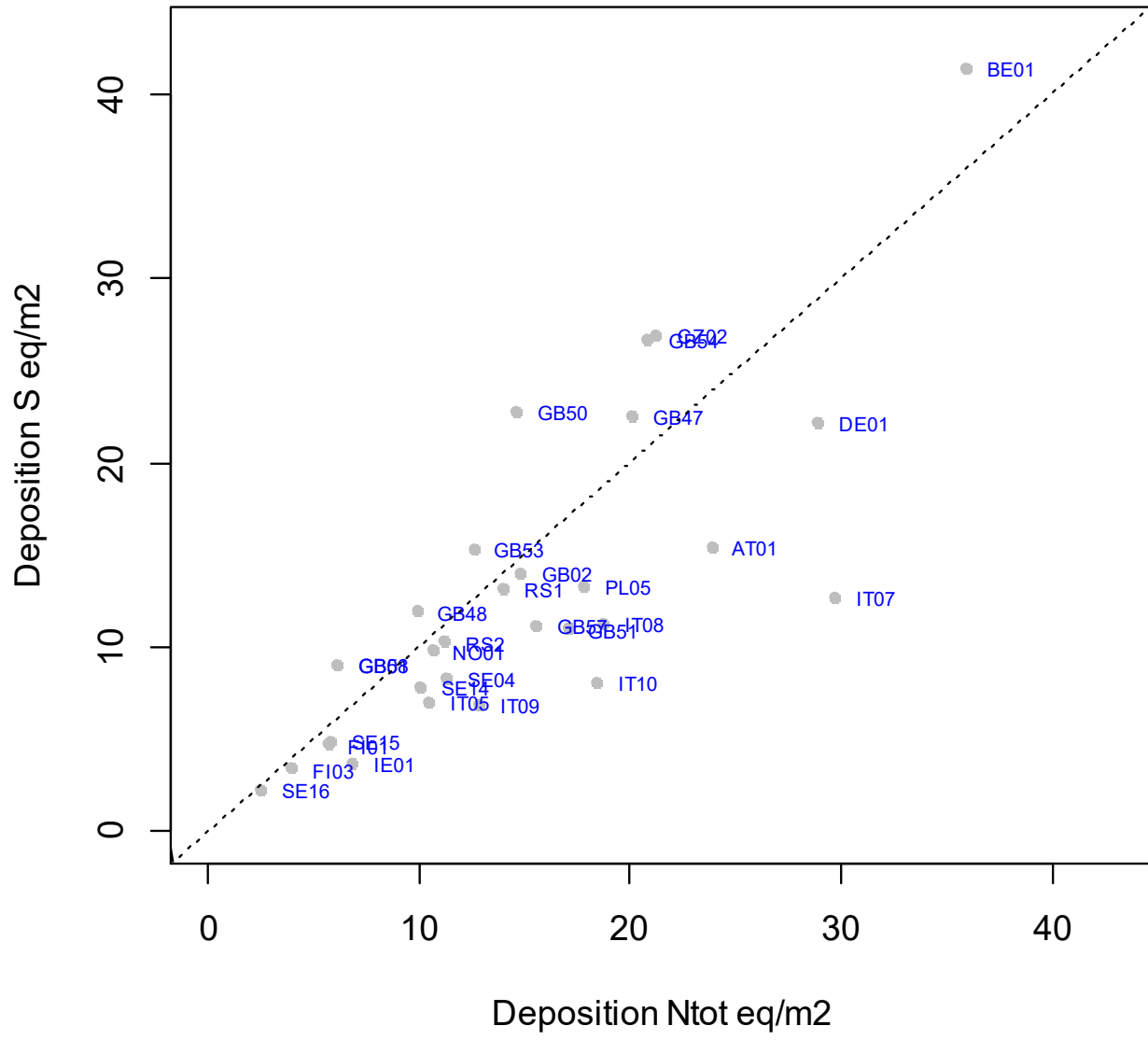
13 countries	40 sites
Austria	5
Belgium	1
Czech Republic	1
Germany	4
Greece	1
Finland	2
UK	10
Ireland	1
Italy	5
Norway	1
Poland	3
Serbia	2
Sweden	4



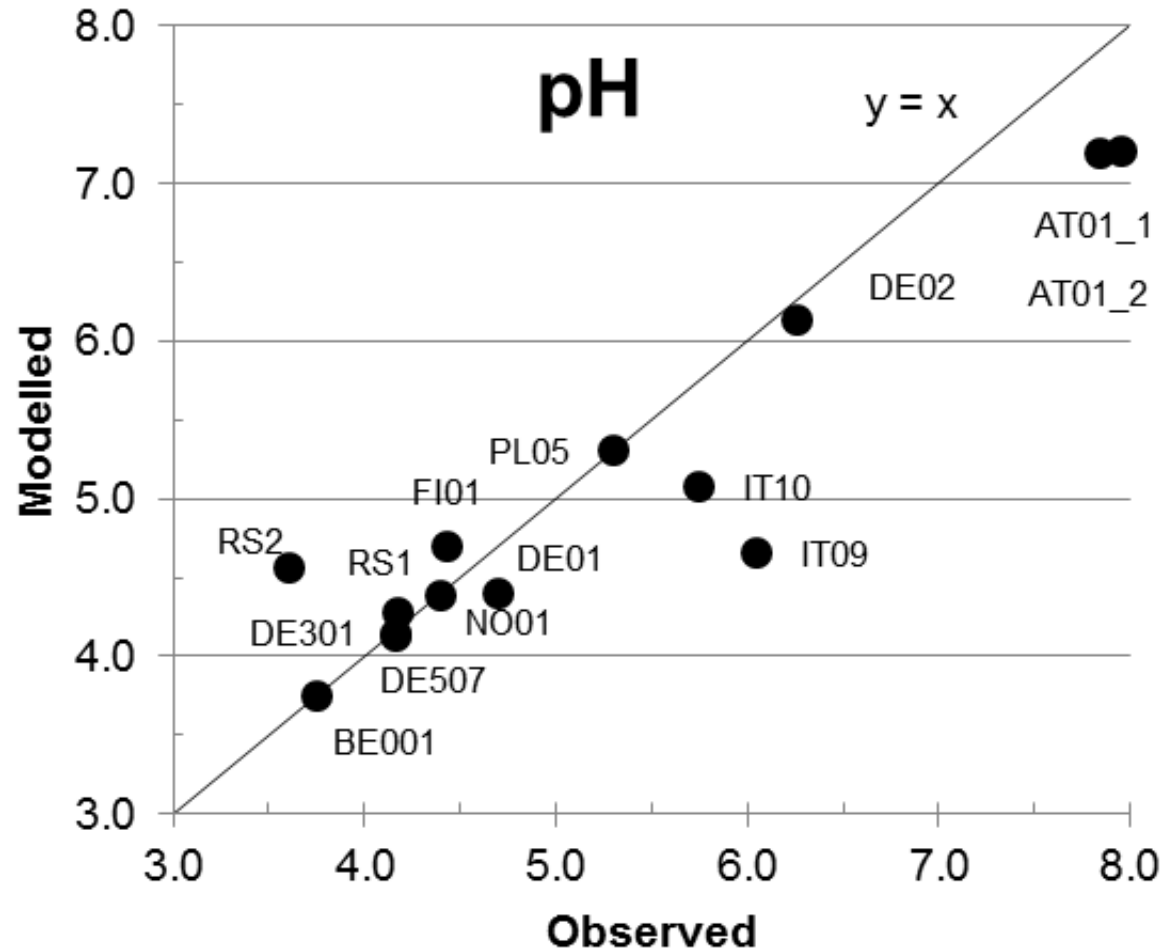
Peak deposition to forest



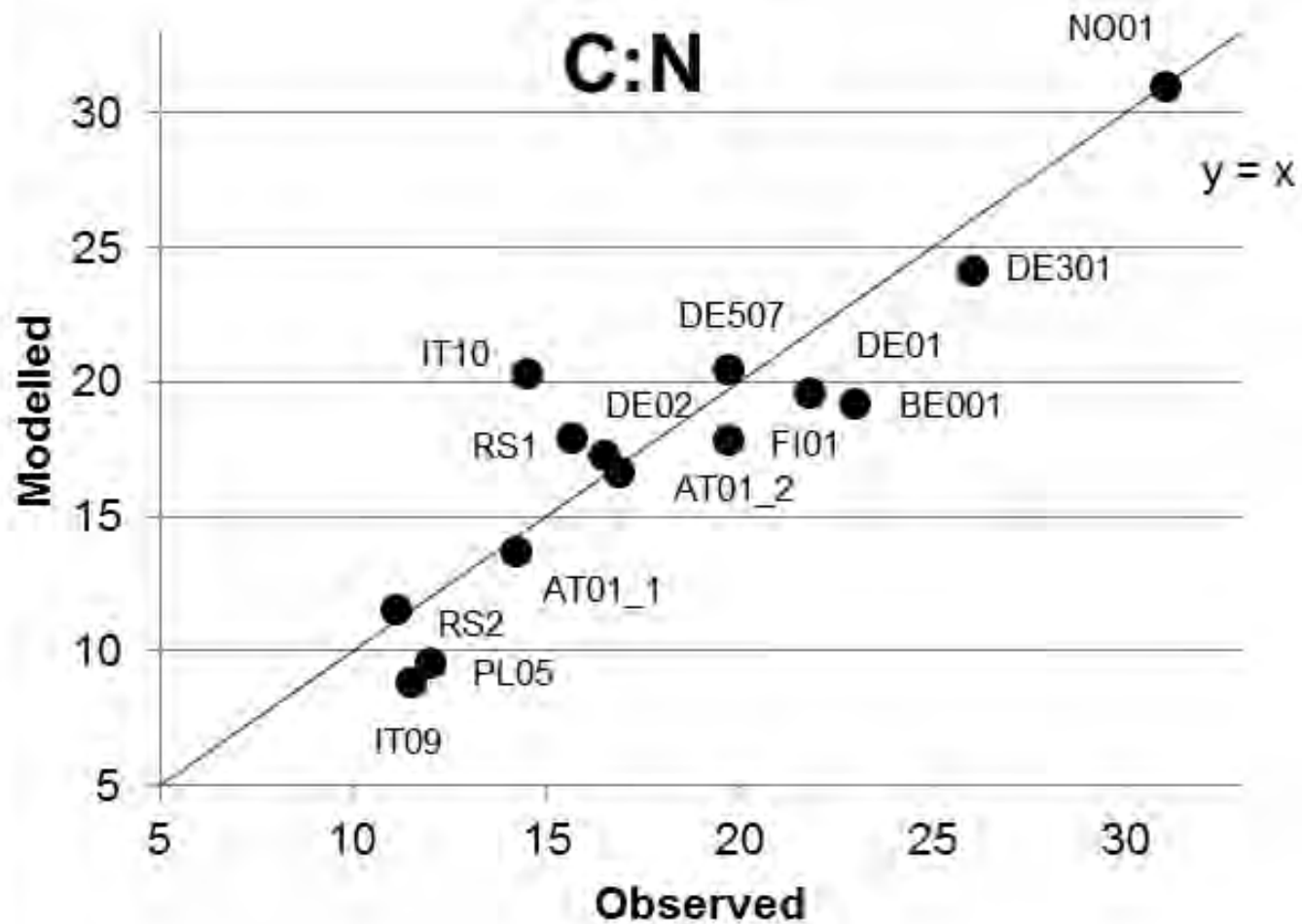
Cumulative forest deposition 1880 - 2100



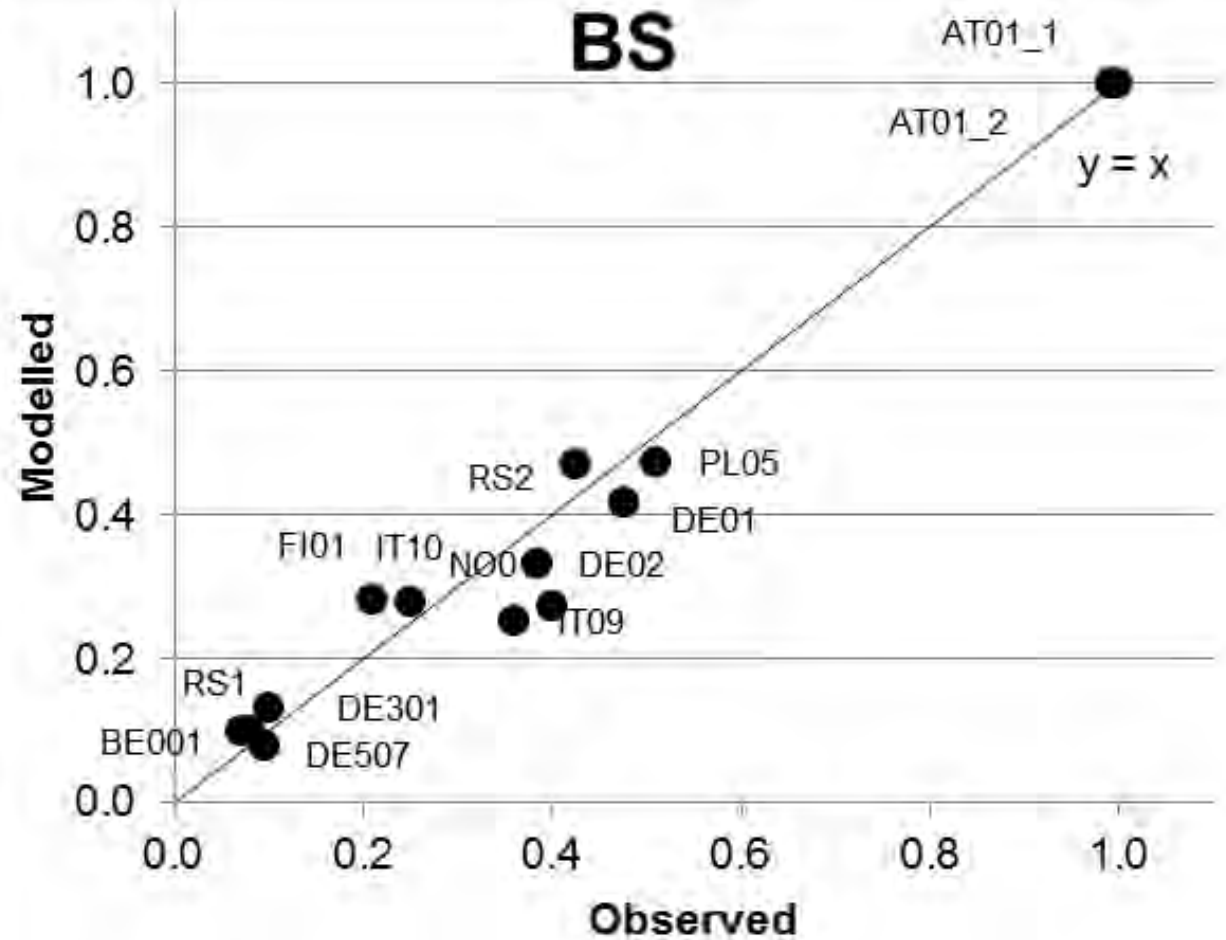
Comparison of modelled to observed values of soil solution pH at 14 sites.



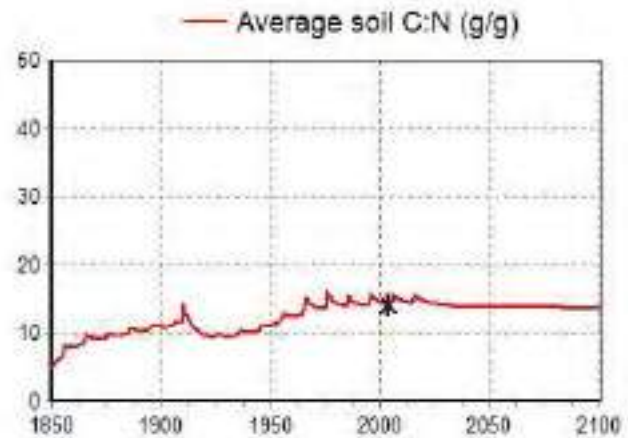
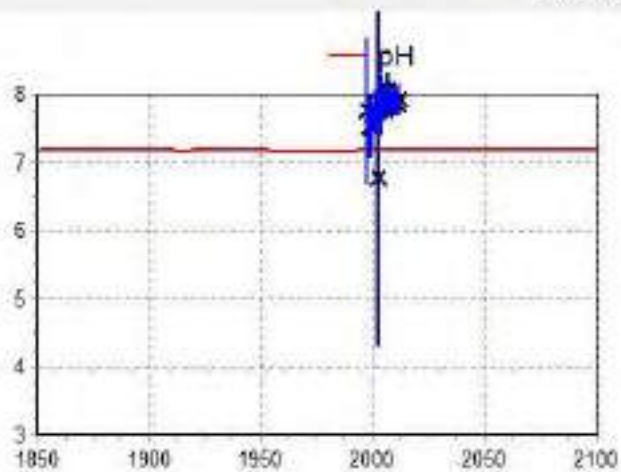
Comparison of modelled to observed values of soil C:N at 14 sites.



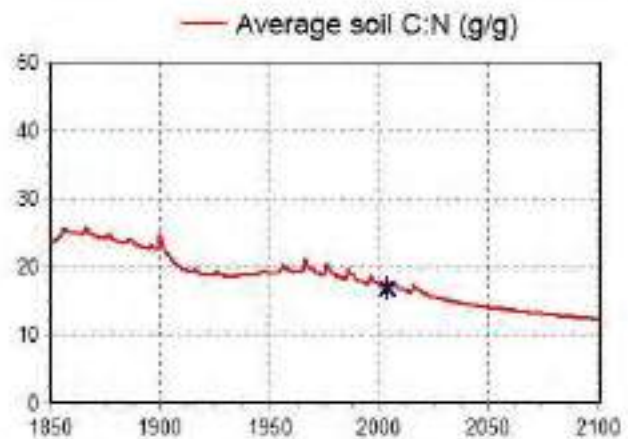
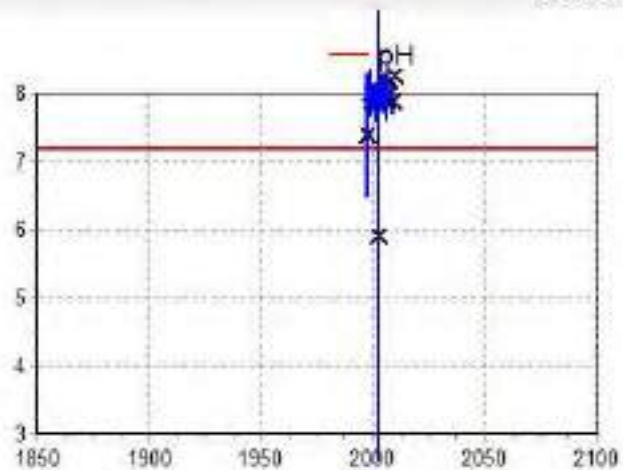
Comparison of modelled to observed values of soil BS at 14 sites.



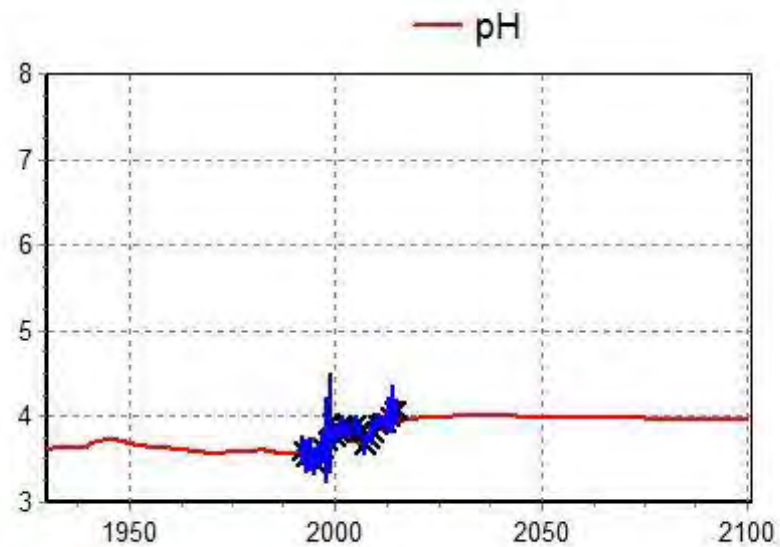
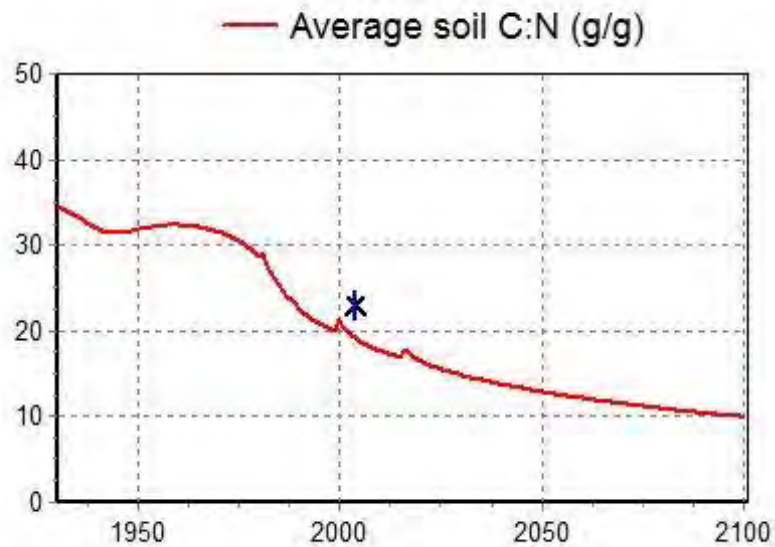
AT01a Zöbelboden IP1



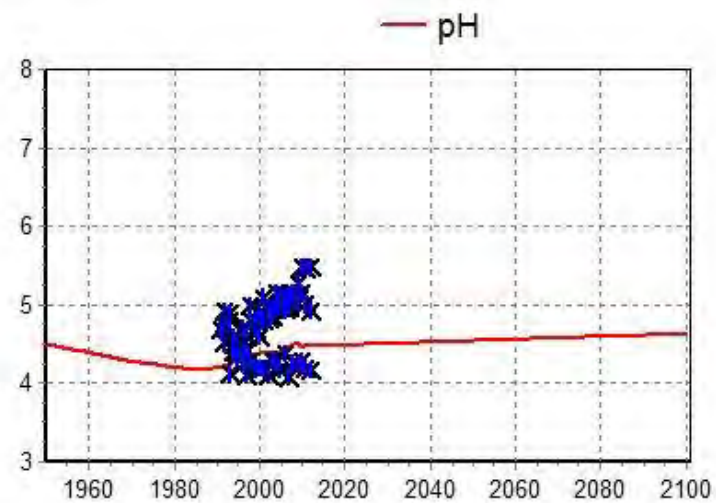
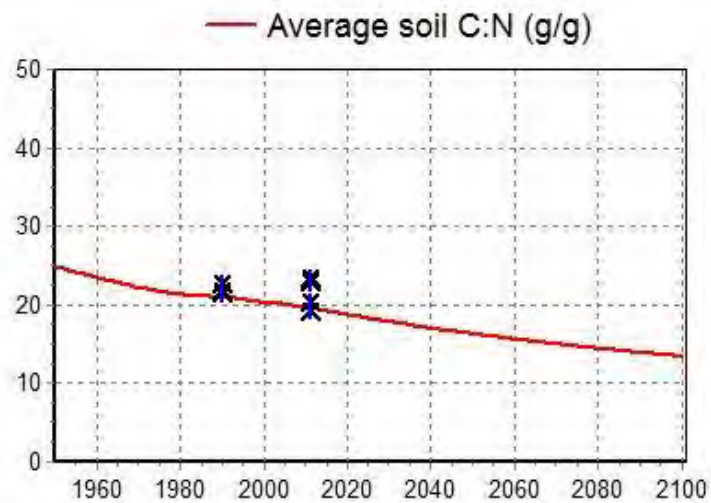
AT01b Zöbelboden IP2



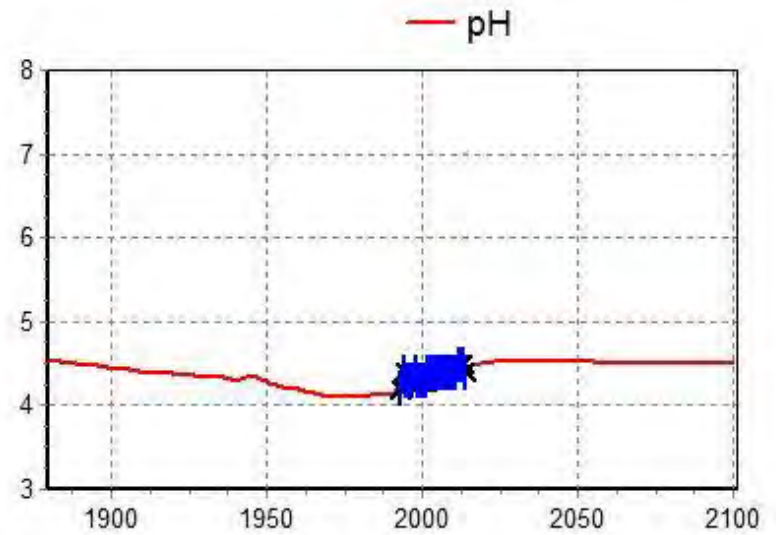
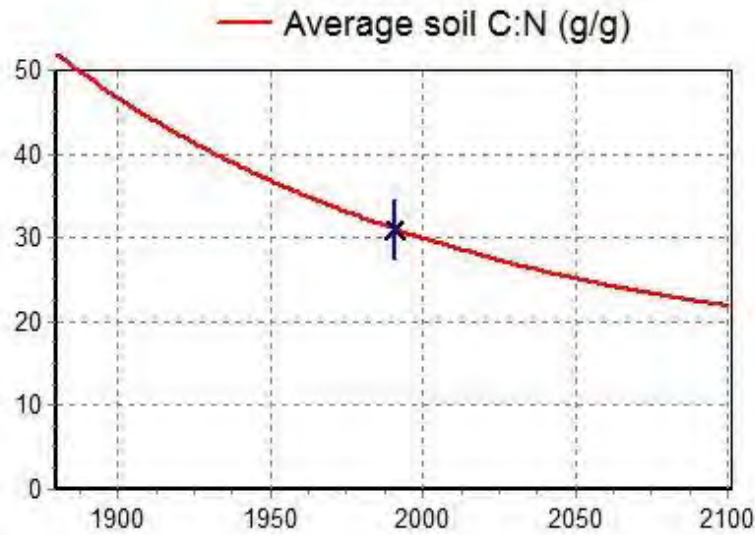
BE01 Brasschaat



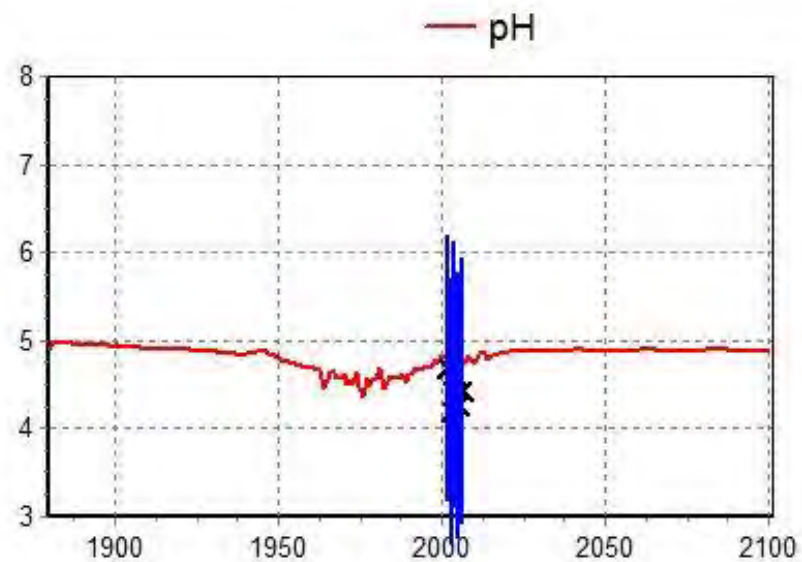
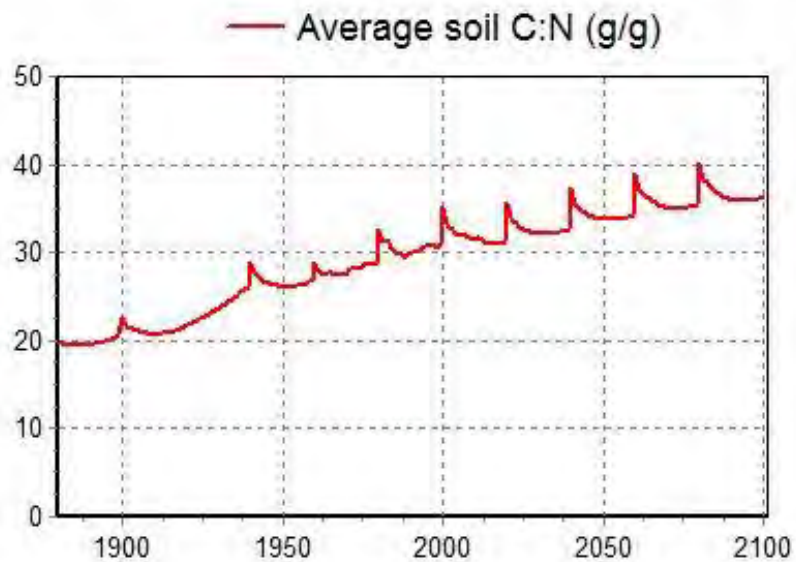
DE01 Forellenbach



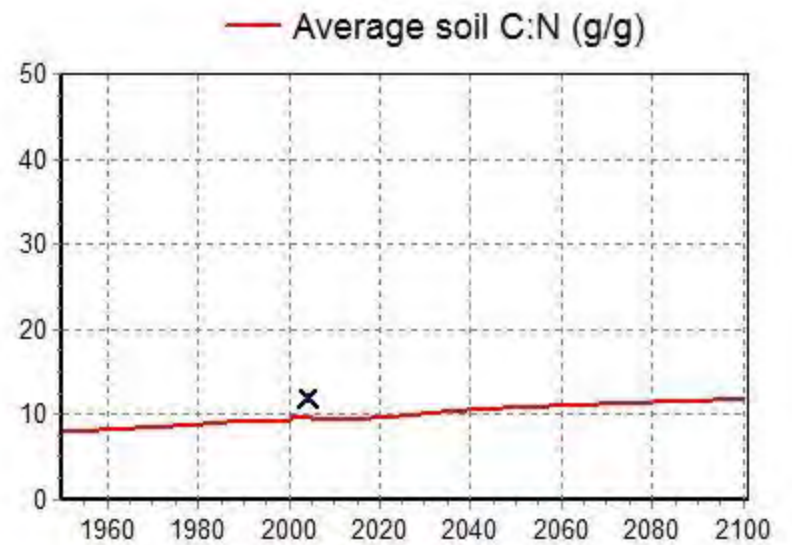
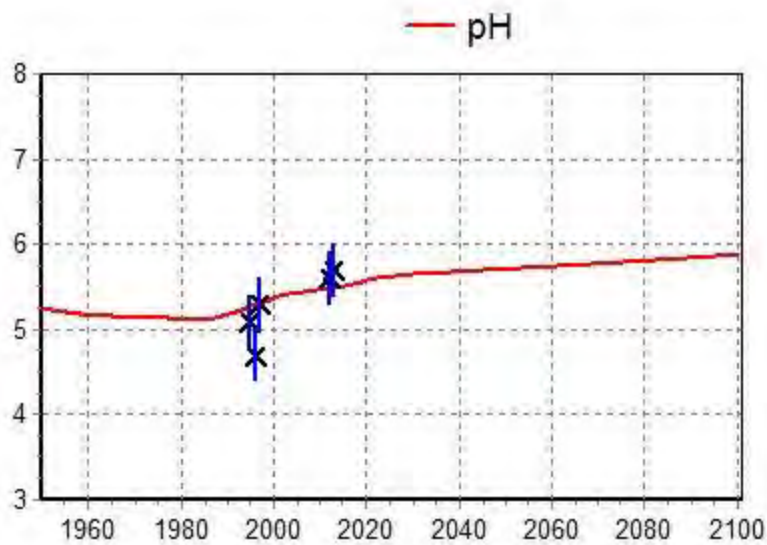
Birkenes NO01



FI01 Evo Valkea-Kotinen

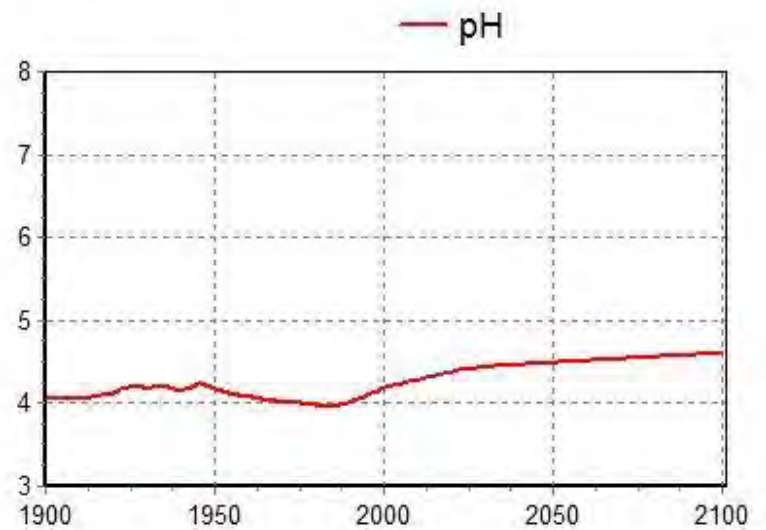
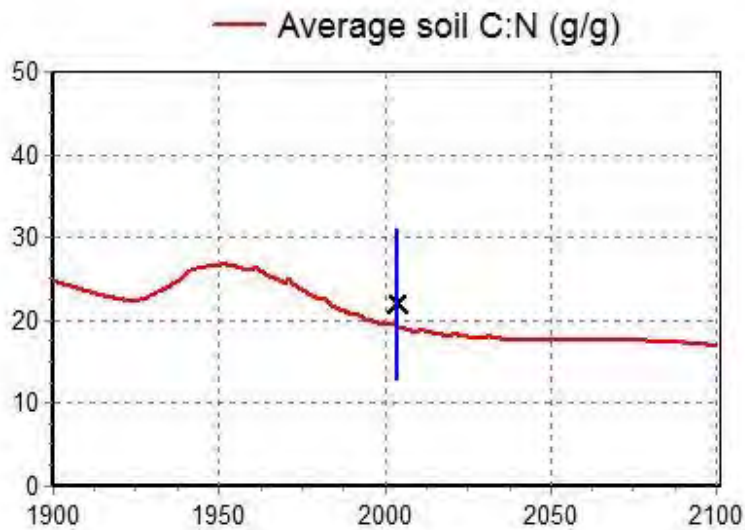


PL05 Borecka

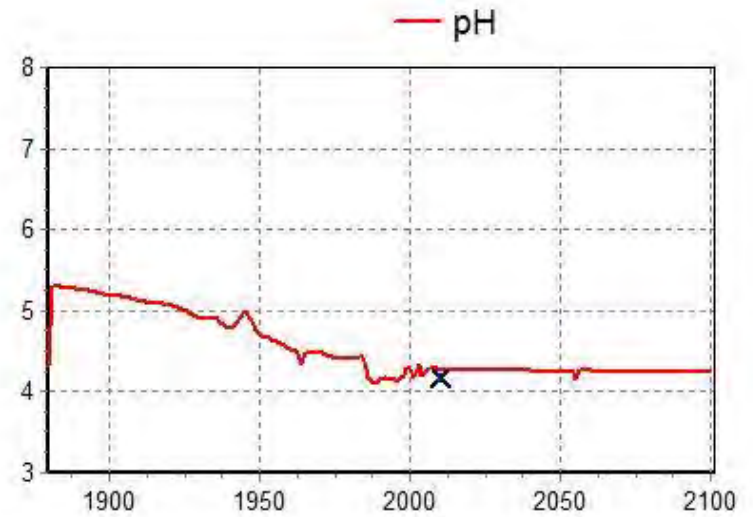
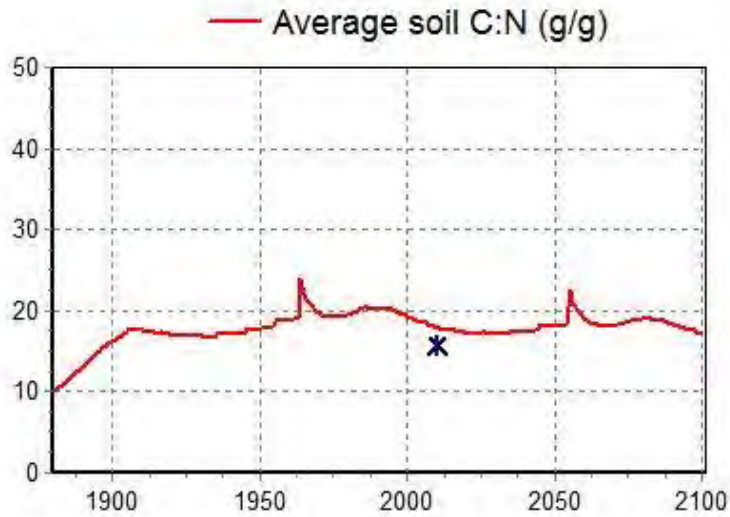


Time development | X-Y graphs |

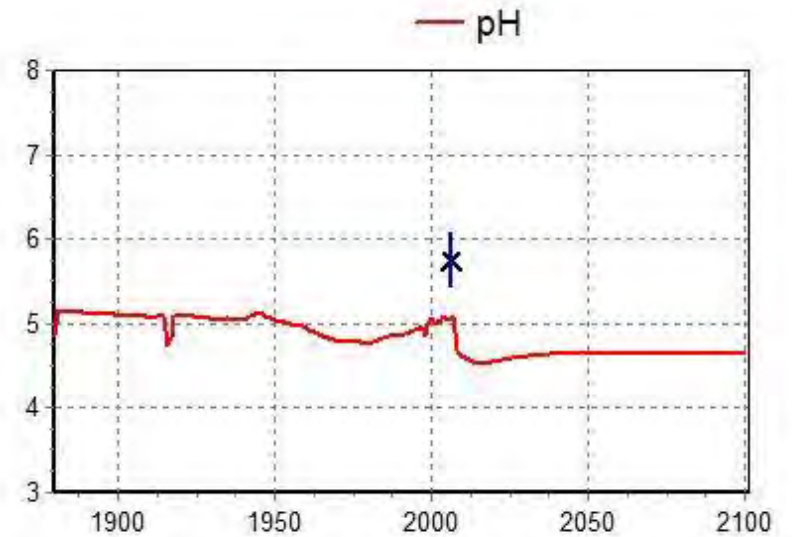
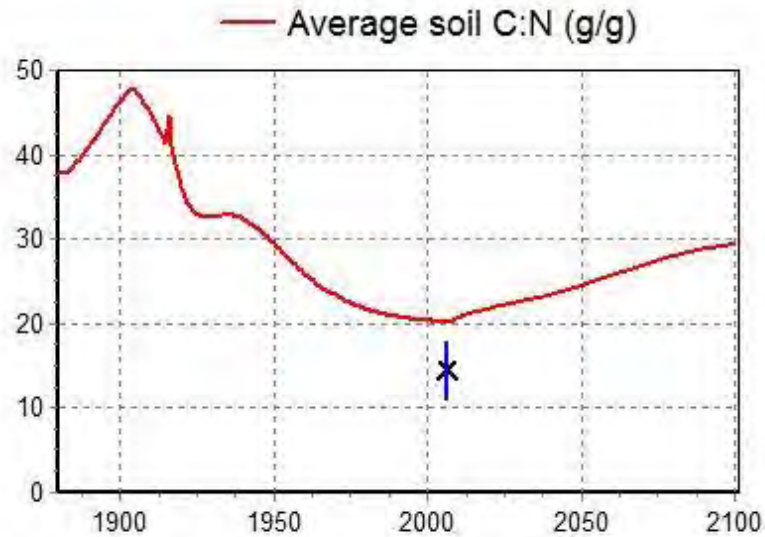
PL SNP Slowinski National Park



RS1 Kopaonik



IT10 LOM1



Next and future tasks

Ndep

- Simulate with deposition scenarios

T, P

- Simulate with climate scenarios

BS, C:N

- Evaluate soil abiotic results

The slide features several decorative icons of fish-like shapes with grid patterns, scattered around the central text. One is in the top left, one in the top right, one in the middle right, and one in the bottom right.

Thank you for your interest

maria.holmberg@ymparisto.fi



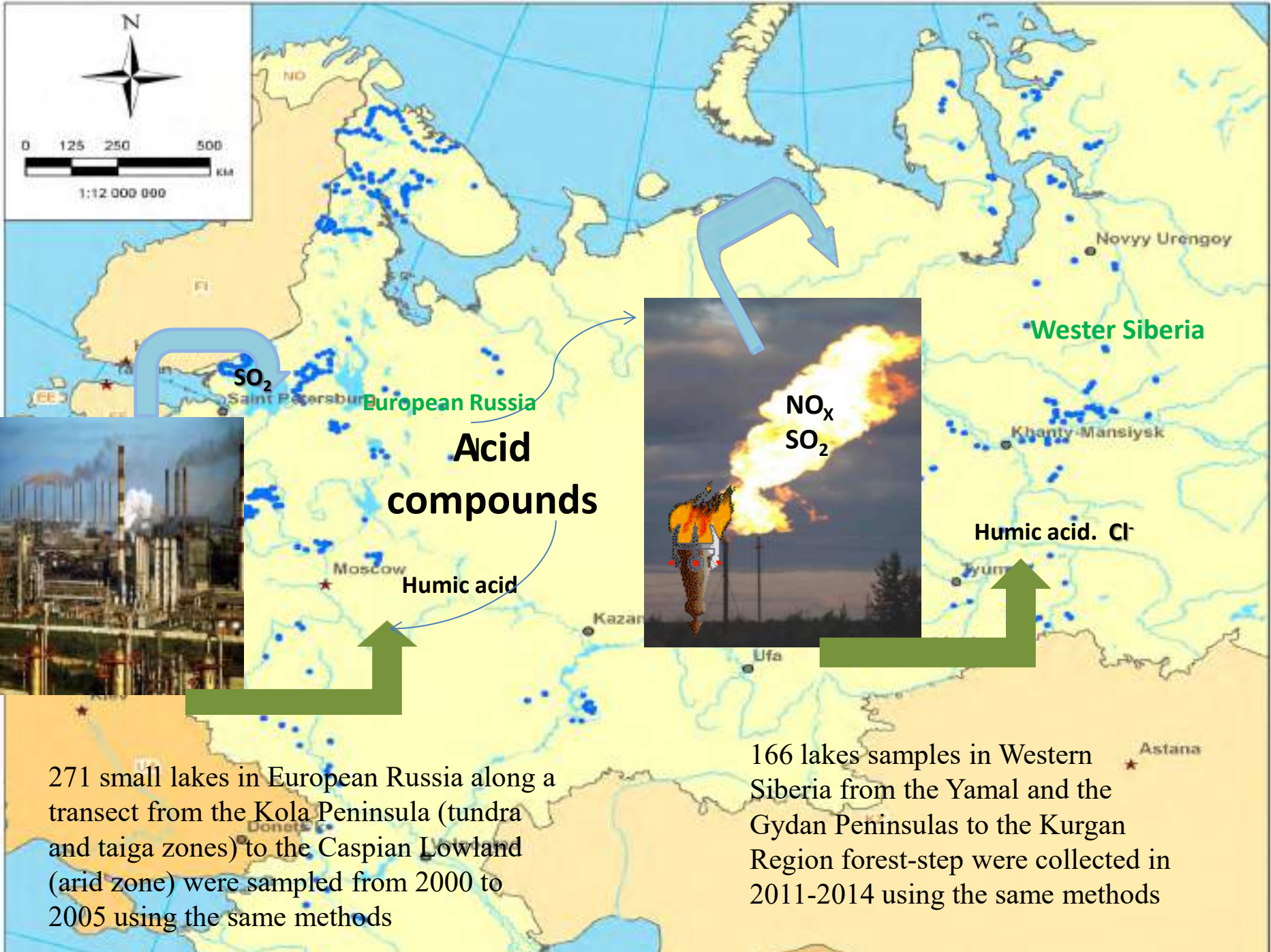
**V.I.Vernadsky Institute of
Geochemistry and Analytical
Chemistry RAS**



State University of Tyumen

**Water acidification and critical loads:
case study European Russia and Western Siberia**

**Moiseenko T.I.. Dinu M.I..
Kremleva T.A.. Gashkina N.A.
Khoroshavin V.Yu.**



271 small lakes in European Russia along a transect from the Kola Peninsula (tundra and taiga zones) to the Caspian Lowland (arid zone) were sampled from 2000 to 2005 using the same methods

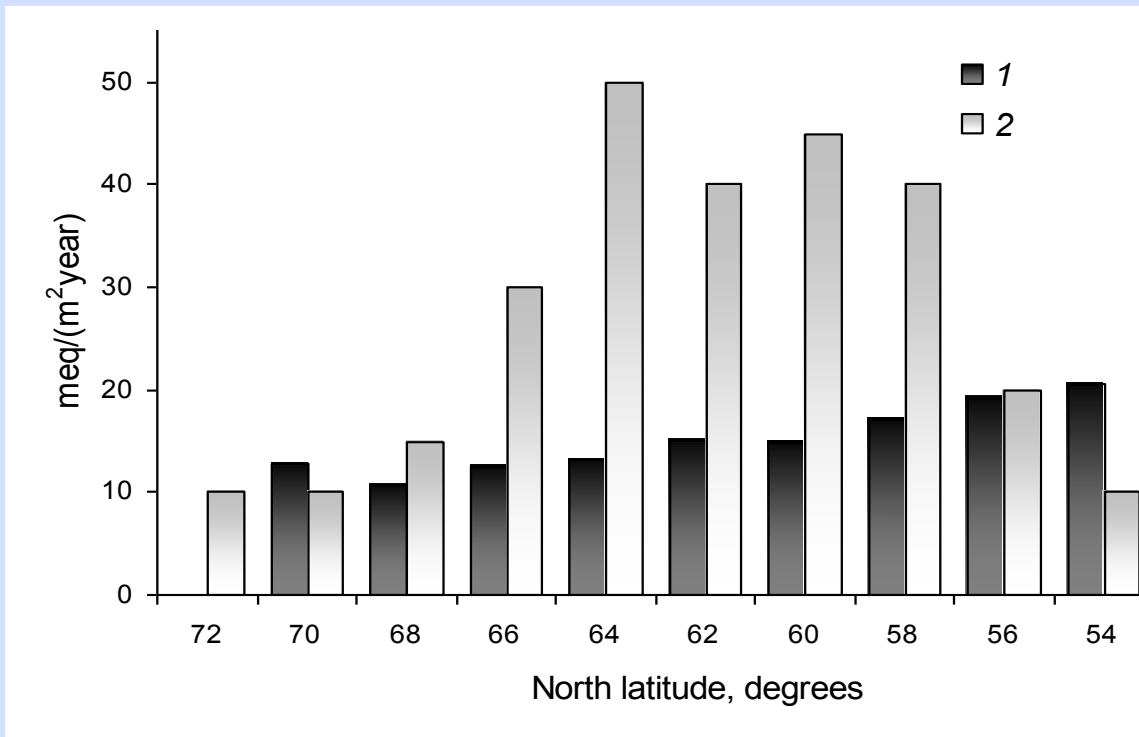
166 lakes samples in Western Siberia from the Yamal and the Gydan Peninsulas to the Kurgan Region forest-step were collected in 2011-2014 using the same methods

Gas flaring during of oil production in Western Siberia leads to air pollution by oxides of nitrogen, sulfur, chlorine.



Western Siberia is a leader among Russian regions for the oil and gas extraction: more than 6% of the world's oil production is concentrated in this region.

Deposition



Deposition of strong acids (meq/m²·year) by the latitudinal gradient **in European territory of Russia - 1** (EMEP, 2000) and **in Western Siberia - 2** (Semenov, 2002)

pH of precipitation

Region	pH of precipitation	
	Min	Max
North and North-West of ER	3.1	6.2
Center of ER	3.2	7.0
South of ER	3.1	7.1
Urals and the Urals region	4.0	7.2
Center of Western Siberia northern coast and north-eastern seas	3.6	7.0

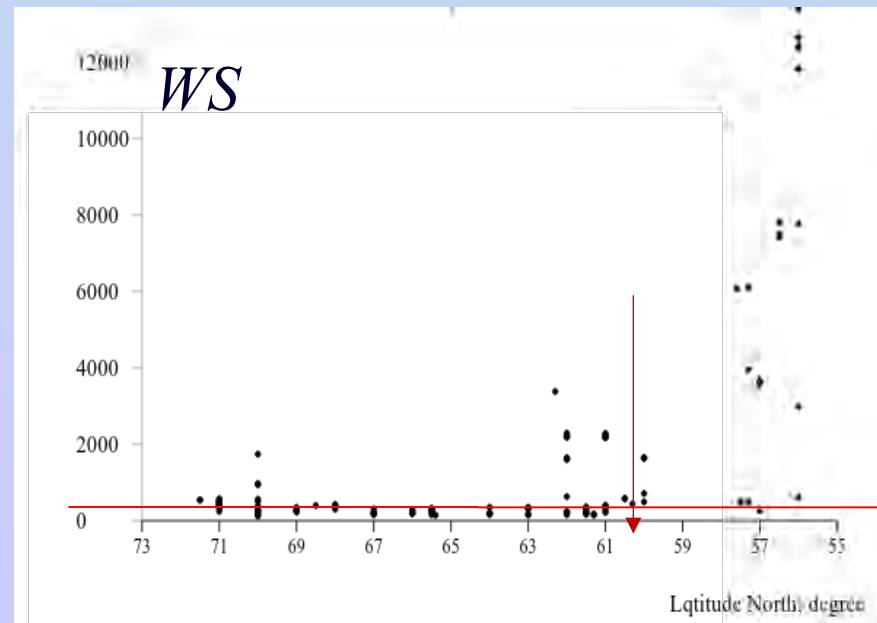
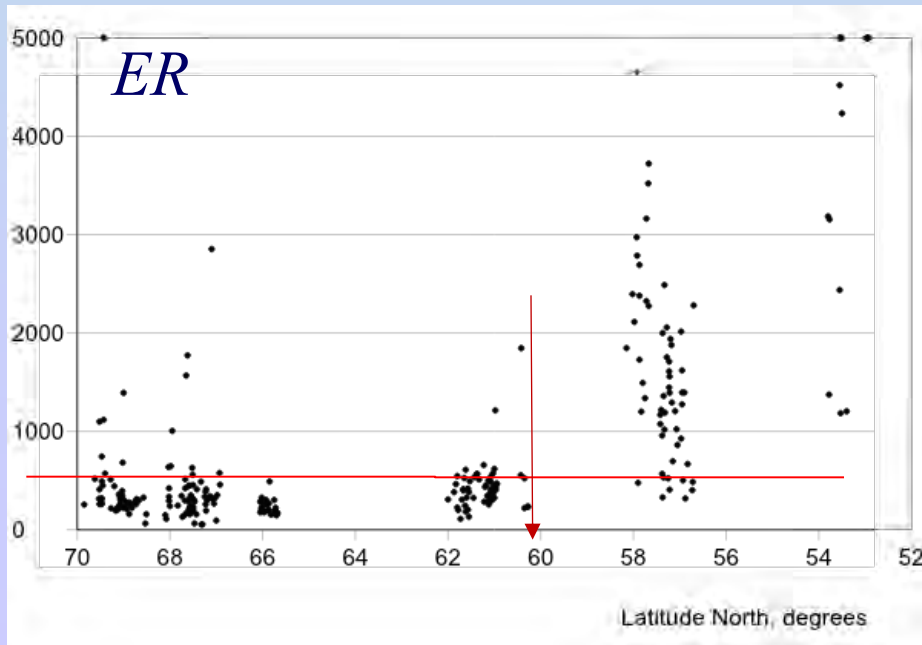
Protasov et al., 2000

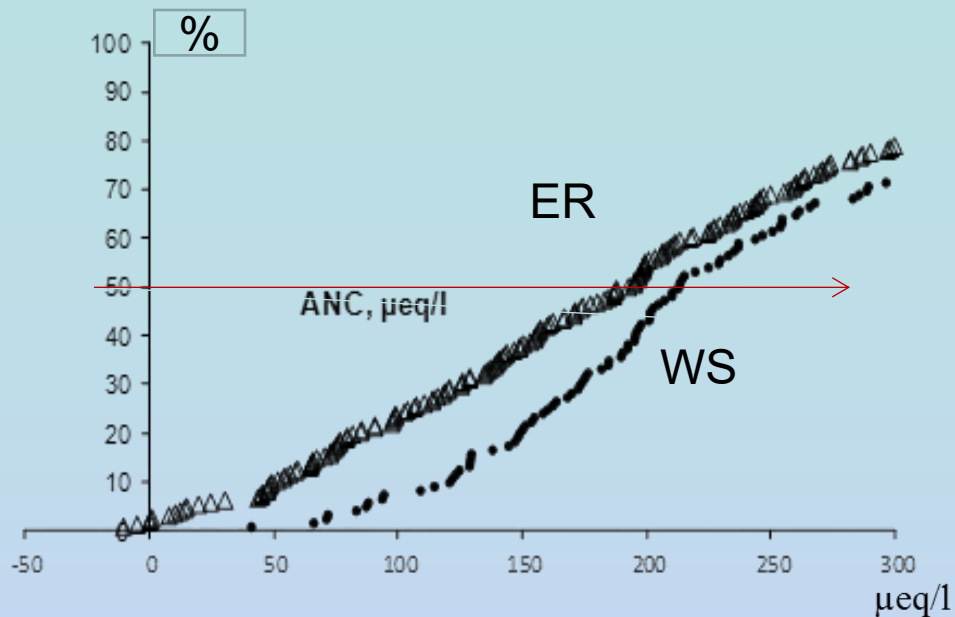
Buffer capacity

The main feature of both regions is **an increase of cations and alkalinity concentrations** in water towards the south: for ER - 55-60 °, for WS - 55-60 ° North latitude. The lakes of forest-steppe zones in ER and WS are highly resistant to acidification. Therefore, further discussion of these lakes was excluded.

Buffer capacity of northern and middle taiga region of ER and WS connect with the features of geology.

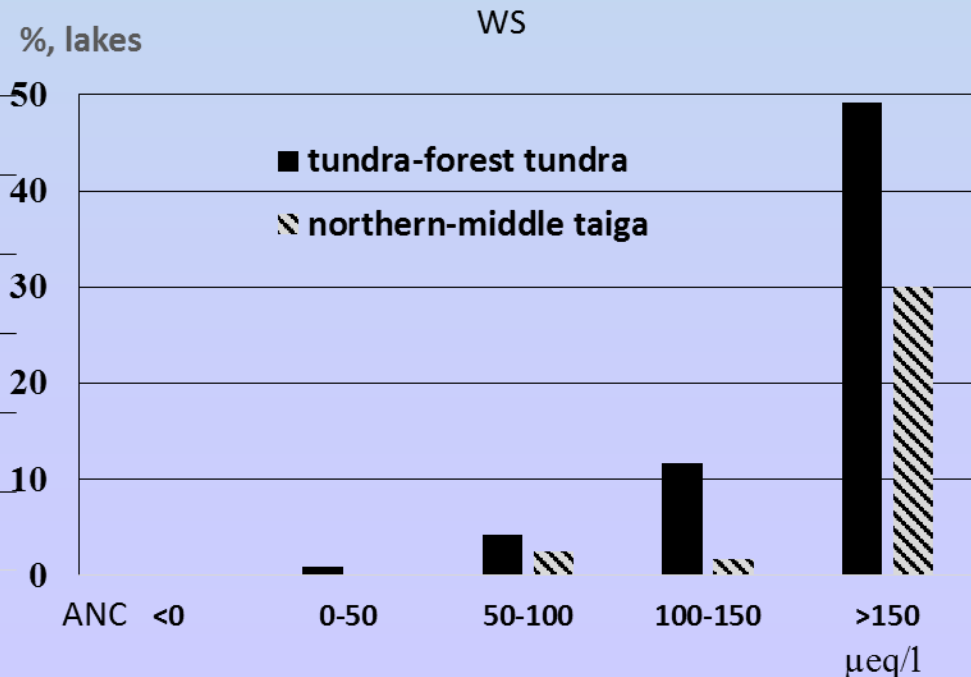
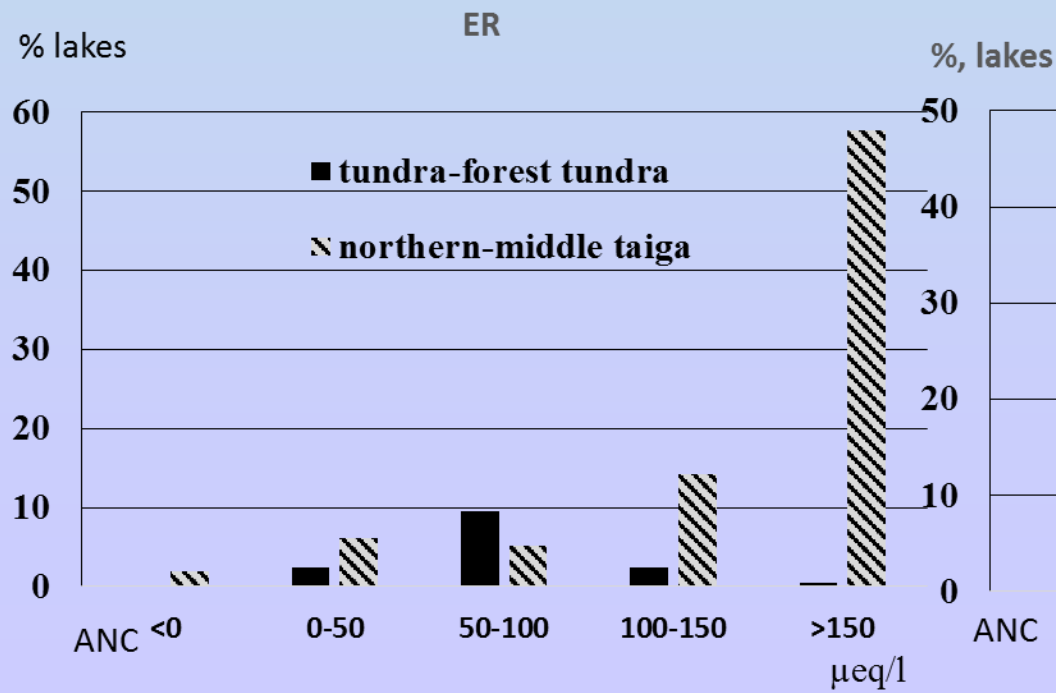
Concentration of cations (Ca+Mg+K+Na), $\mu\text{eq/l}$, in smalls lakes

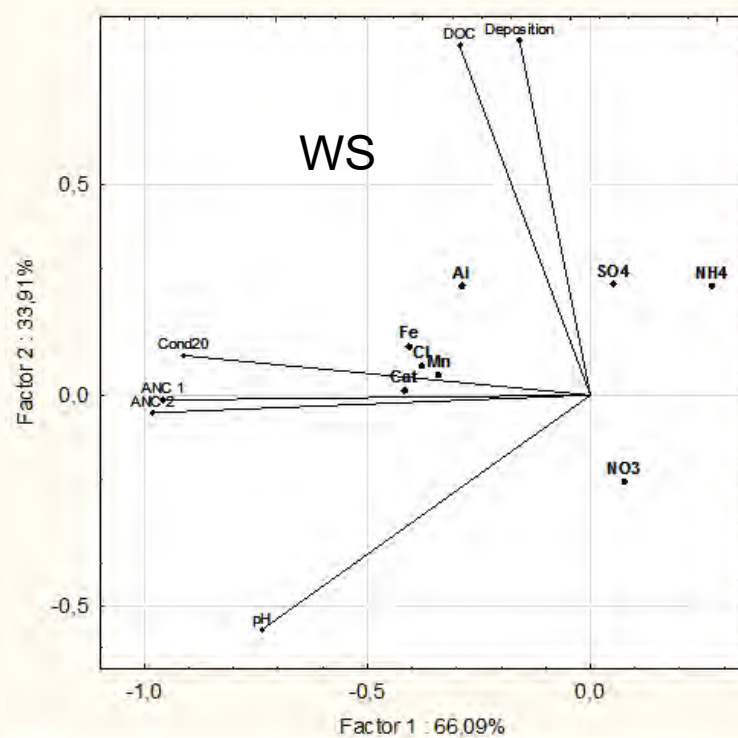
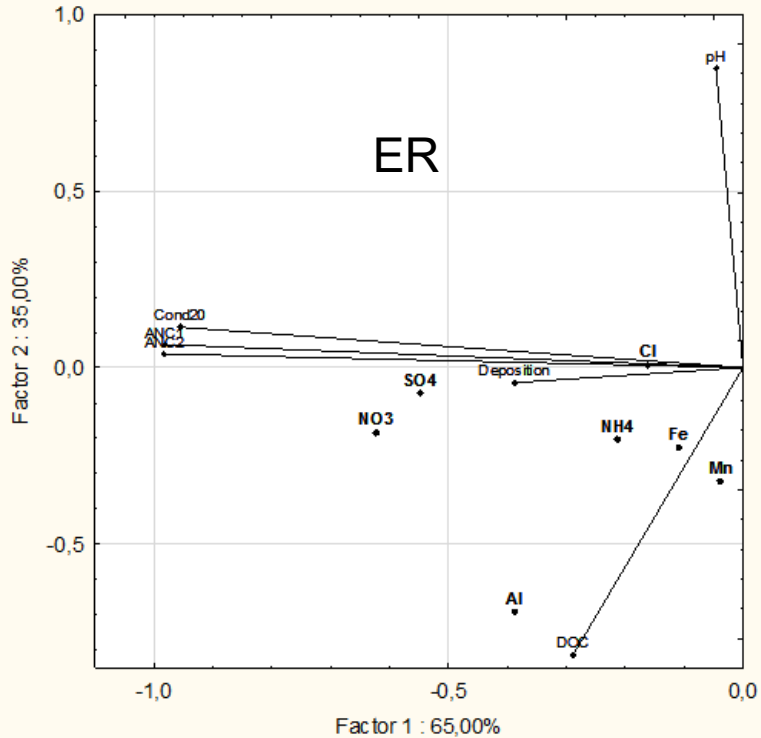




The acid neutralizing capacity of water (ANC, $\mu\text{eq/l}$) is used to estimate the anthropogenic acidification as a difference between cations and anions of strong acids (Henriksen et. al., 1992)

1. $\text{ANC1} = \text{Ca}^{2+} + \text{Mg}^{2+} + \text{Na}^{+} + \text{K}^{+} - \text{SO}_4^{2-} - \text{NO}_3^{-}$
2. $\text{ANC2} = \text{HCO}_3^{-} + \text{A}^{\text{n}} - \text{H}^{+} - \text{Al}^{3+}$



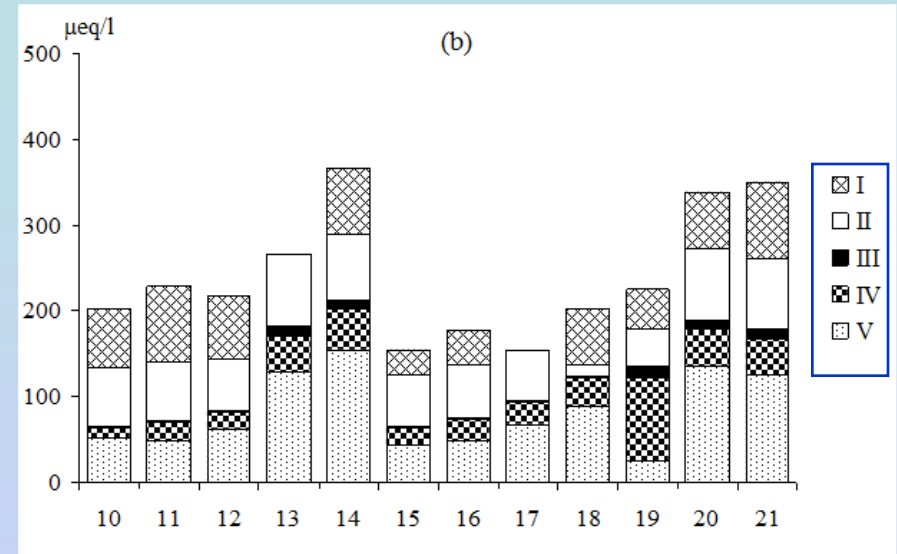
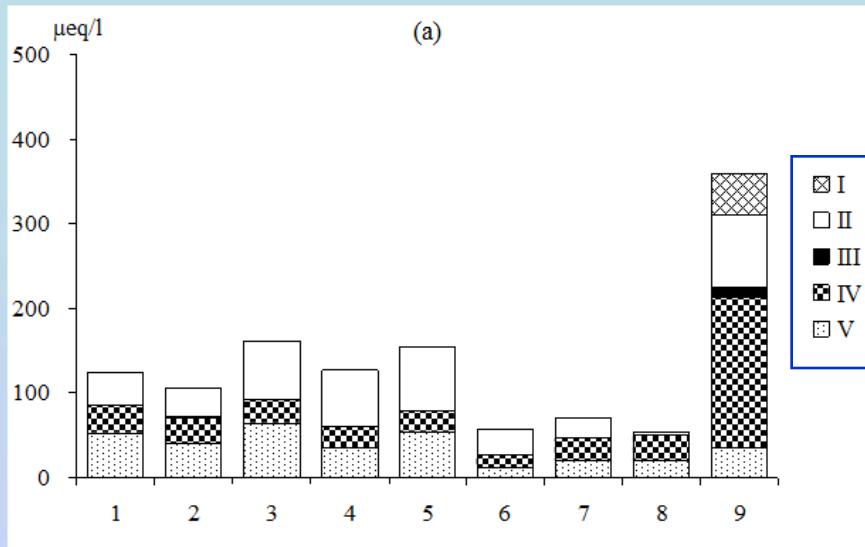


According to the results : **ANC 1 and ANC 2 are very close** that means the reliable of analyzes.

In **ER** lakes SO₄, Cl and NO₃ mostly have an affinity for Cond. This is due to increased population density and industrial at the same time with increase cations and buffer capacity of water, The direct connection of the deposition and pH (lake ER) confirms the greatest acid deposition in an areas with steady geological formations to acidification.

In the water **WS** Cl have an affinity for Cond because increased salt content of watersheds toward the South. WS is located on the site of paleo area and Quaternary rocks: contain a amount of chloride. SO₄ have an affinity for Deposition despite the higher sulfate concentrations in waters ER. The nitrate is inert to the selected axes because has the technological and natural origin.

The anionic composition (I – Alk. II - A^n . III – NO_3^- . IV – SO_4^{2-} . V – Cl^-) of the water lakes with **pH < 5** on the ER (a) (1 - in the tundra, 2-8 - in the northern taiga, 9 - in the middle taiga) and WS (b) (10-12 in the tundra, 13-14 - in the northern taiga, 15-21 - in the middle taiga).



Water acidification due to anthropogenic sulfate is characterized of ER.

In the acidic lakes of WS the water contained: chlorides, nitrates and sulfates.

Chlorides (lakes 13. 14. 20. 21) are dominated in majority lakes, but in some lake sulfates are dominated (lake 19).

Concentration of nitrates in water WS are higher in compared to the waters of ER.

Water of Western Siberia

NO₃

i) Delivery with the marsh waters, wetland and marsh is widely developed in the WS

(Nopz = 49.7·DOC – 114. (r=0.87. n=120);

i) the gas flaring forms the nitrogen oxides;

Cl⁻

i) WS is located on the site of paleosee area and Quaternary rocks contain a amount of chloride (Arkhipov et al.. 1987).

ii) The chlorides are present in the waters of WS as part of pollution of the oil and gas fields development (Kiriushin et al.. 2013).

Organic compound (A⁻)

Natural humus acids enters with the marsh waters.

Natural humus are characterized by a dual nature: weak and strong acids

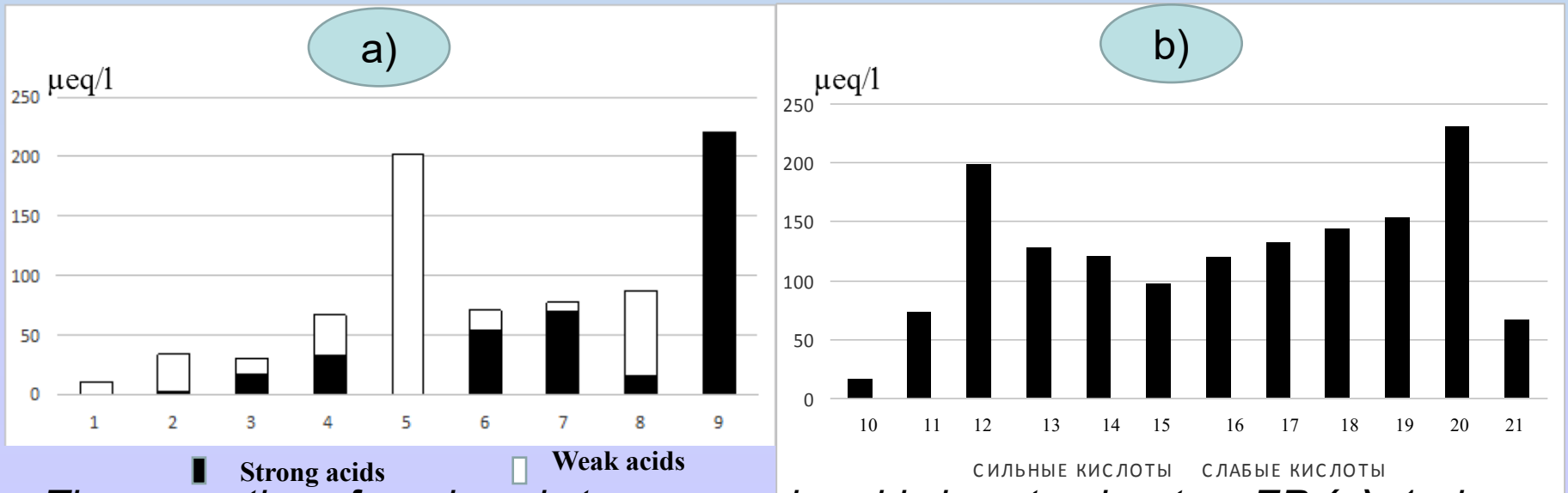
(Evans, Monteith et al.. 2008)

We calculated part of strong and week acids in water for acidify lakes in ER and WS based on method (Evans, Monteith, et.al. 2008). We used date about material balance of water chemical composition in each lake. The calculation algorithm takes into account the whole structure of the cationic and anionic groups including carbonates and bicarbonates.

$$\text{OrgA}_{\text{strong}} = \text{Ca}^{2+} + \text{Mg}^{2+} + \text{Na}^{+} + \text{NH}_4^{+} + \text{K}^{+} + \text{Al}_{\text{inorg}} - \text{SO}_4^{2-} - \text{Cl}^{-} - \text{NO}_3^{-} - \text{Alk}$$

$$\text{OrgA}_{\text{weak}} = \text{A}^{n-} - \text{OrgA}_{\text{strong}}$$

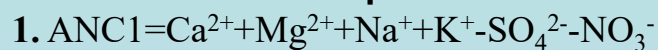
(Evans, Monteith et al.. 2008)



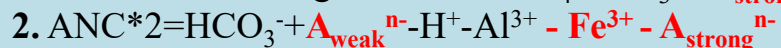
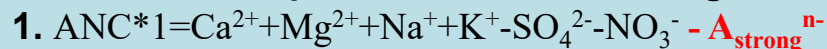
The proportion of weak and strong organic acids in natural waters ER (a): 1 - in the tundra. 2-8 - in the northern taiga. 9 - in the middle taiga and WS (b): 10-12 in the tundra. 13-14 - in the northern taiga. 15-21 - in the middle taiga

In table:

Black color without parentheses is the average ANC classics:



Red color with parentheses is the average ANC* with used of organic weak and strong acids



pH	Color, °Pt-Co scale					
	<10	10-30	30-60	60-100	>100	<i>n</i>
European Russia, 220 lakes, ANC*, µeq/l						
4-5	-5.0 (-20)	7 (25)	-	4 (-7)	130 (-9)	10
5-6	43 (31)	270 (187)	250 (149)	310 (279)	410 (222)	34
6-7	69 (54)	220 (176)	235 (168)	235 (131)	260 (108)	141
7-8	234 (233)	380 (335)	-	400 (305)	373 (269)	35
Western Siberia, 120 lakes, ANC*, µeq/l						
4-5	40 (-5)	62 (3.5)	140 (5.3)	175 (6.2)	250 (6.2)	18
5-6	150 (4.8)	170 (4.0)	230 (10)	300 (11)	350 (15)	54
6-7	300 (10)	375 (9.7)	350 (6.5)	-	400 (5.5)	35
7-8	472 (153)	1142 (8)	1100 (220)	1450 (75)	1560 (606)	13

Critical Loads and its exiding

$$CL = ([BC_o^*] - [ANC_{limit}]) Q - BC_d^*$$

$$BC_o^* = [BC^*]_t - F \Delta([SO_4^*] + [NO_3]) = [BC^*]_t - F (([SO_4^*]_t + [NO_3]_t) - ([SO_4^*]_o + [NO_3]_o))$$

$$F = \sin(\pi/2) [BC]_t / S$$

$$[SO_4]_{o^*} = 15 + 0.16 [BC_t]^* \quad (\text{Henriksen et al., 1992})$$

European Russia

$$BC_o = [BC]_t - F (([SO_4]_t - [SO_4]_o))$$

Zones of tundra, forest-tundra and northern taiga (Kola Peninsula):

$$[SO_4^*]_o = 15.3 + 0.02 [BC^*]_t. \quad r=0.71. \quad p<0.001;$$

Zone of the middle taiga (Karelia):

$$[SO_4^*]_o = 15.4 + 0.11 [BC^*]_t. \quad r=0.64. \quad p<0.001;$$

Zone of mixed forests:

$$[SO_4^*]_o = 15.2 + 0.05 [BC^*]_t. \quad r=0.68. \quad p<0.001.$$

S is 400 $\mu\text{eq} / \text{l}$ for the tundra and taiga zones.

S is 1100 $\mu\text{eq} / \text{l}$ for mixed ER forests.

$$CL_{ex} = CL - SO_4^*_{dep} - NO_{3dep} + BC^*_{dep}$$

Western Siberia

$$BC_o = [BC]_t - F (([SO_4]_t - [SO_4]_o) + ([NO_3]_t - [NO_3]_o) + ([Cl]_t - [Cl]_{Na}))$$

$$[NO_3]_o = 0.118 [A^n]_t.$$

(Cl_{Na} is compensated by Na)

Zones of tundra, forest-tundra and northern taiga:

$$[SO_4]_o = 2.67 + 0.021 [BC]_t. \quad r=0.72. \quad p<0.001;$$

Zone of the middle taiga:

$$[SO_4]_o = 16.9 + 0.015 [BC]_t. \quad r=0.76. \quad p<0.001;$$

Zone of southern taiga:

$$[SO_4]_o = 12.4 + 0.002 [BC]_t. \quad r=0.69. \quad p<0.005.$$

S is 500 $\mu\text{eq} / \text{l}$ for the tundra and northern taiga zones.

S is 1250 $\mu\text{eq} / \text{l}$ for the middle taiga

S is 3000 $\mu\text{eq} / \text{l}$ for the southern taiga.

$$CL_{ex} = CL - SO_{4dep} - NO_{3dep} - Cl_{dep} + BC_{dep}.$$

Thus, the necessary data for estimating the flow of cations into water systems ensuring neutralization of anthropogenic acids have been determined. Taking into account the complete and correctly obtained hydro-chemical information.

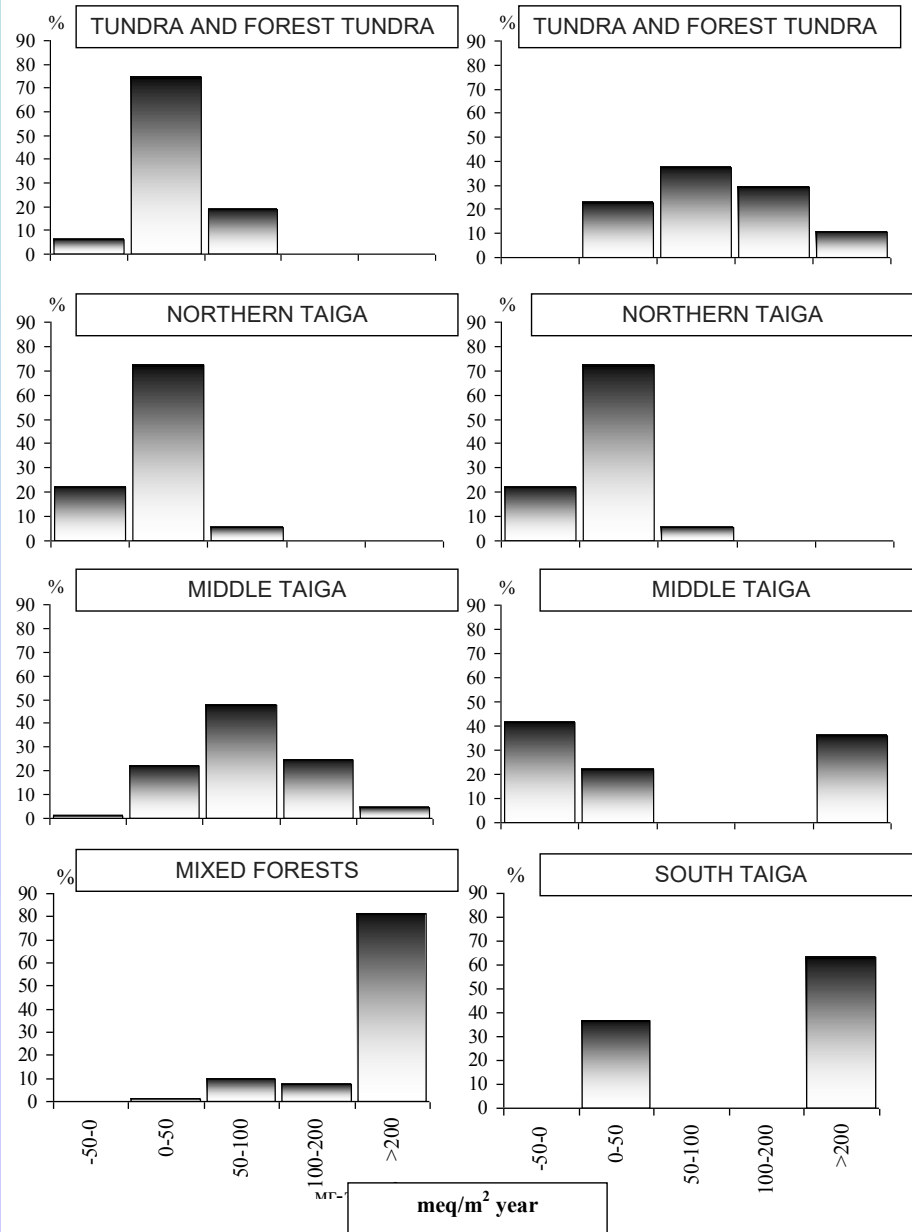
ANClimit was adopted as 50, $\mu\text{eq} / \text{l}$

CL

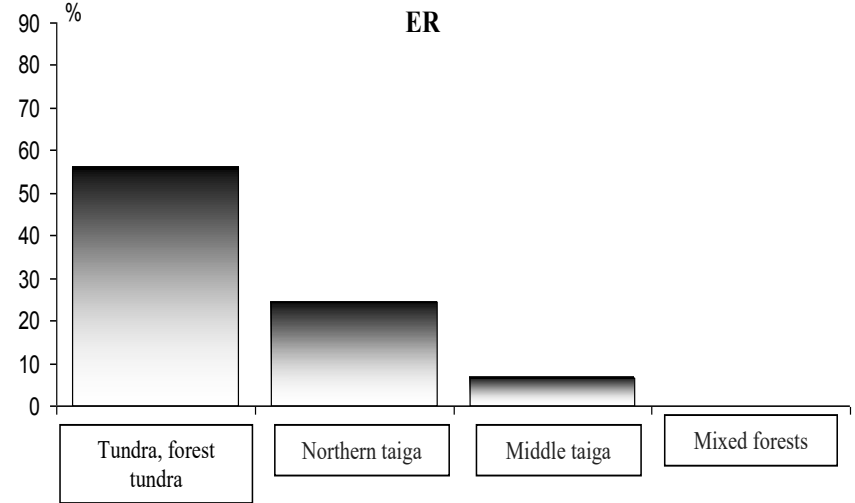
CLex

ER

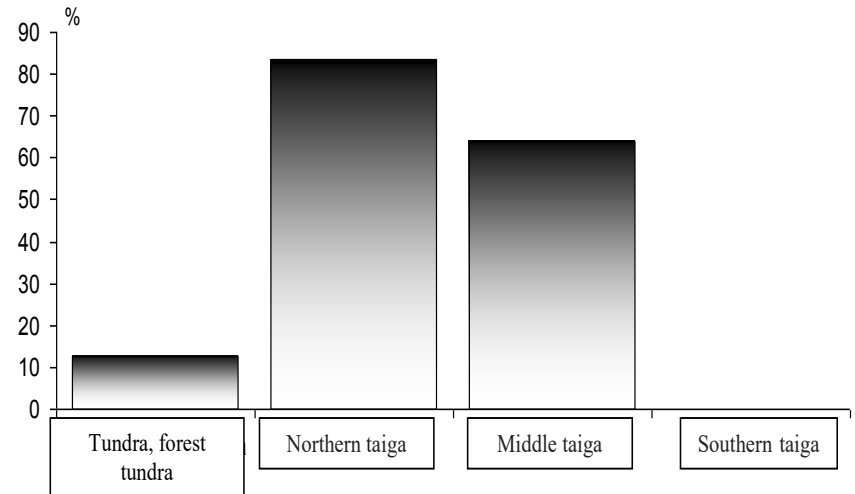
WS



ER

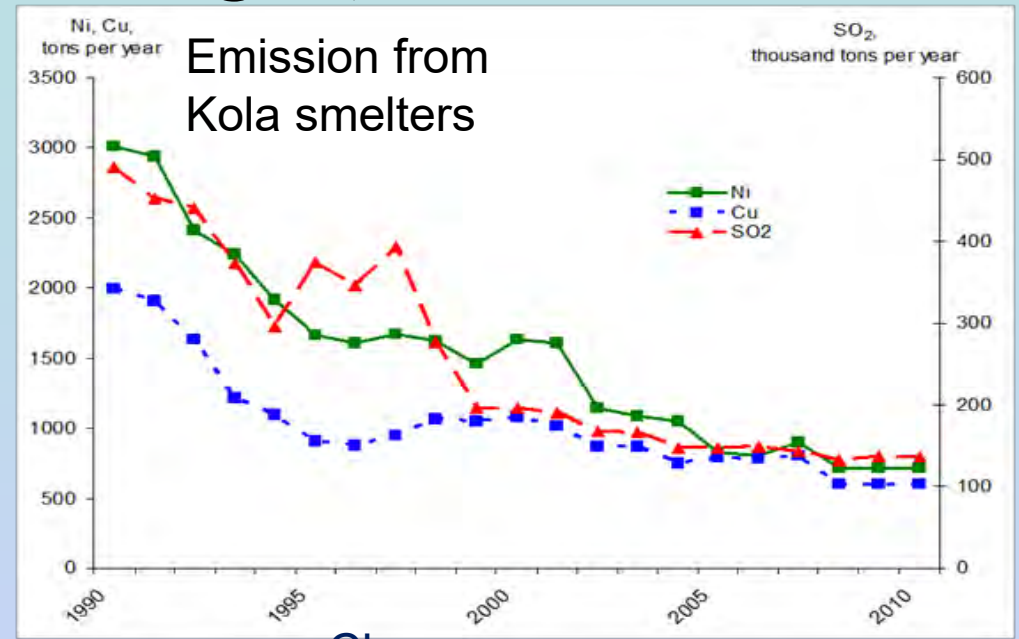
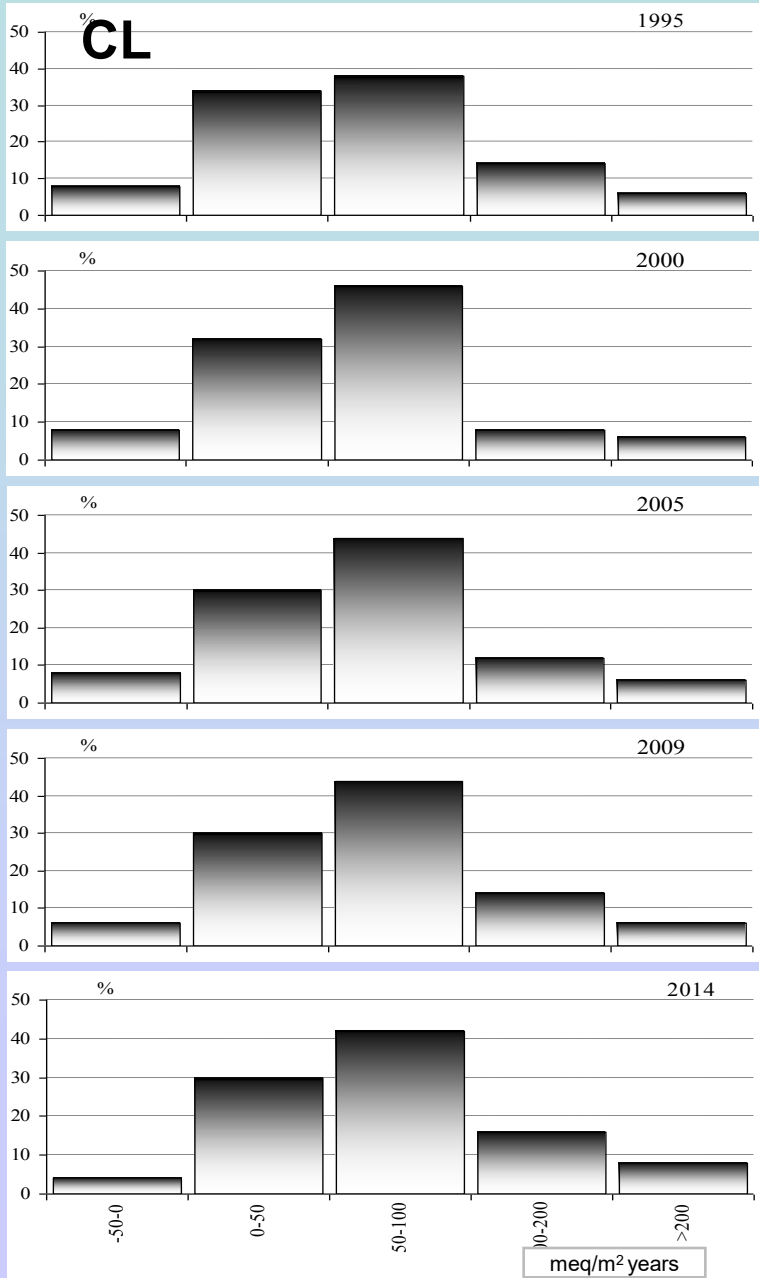


WS

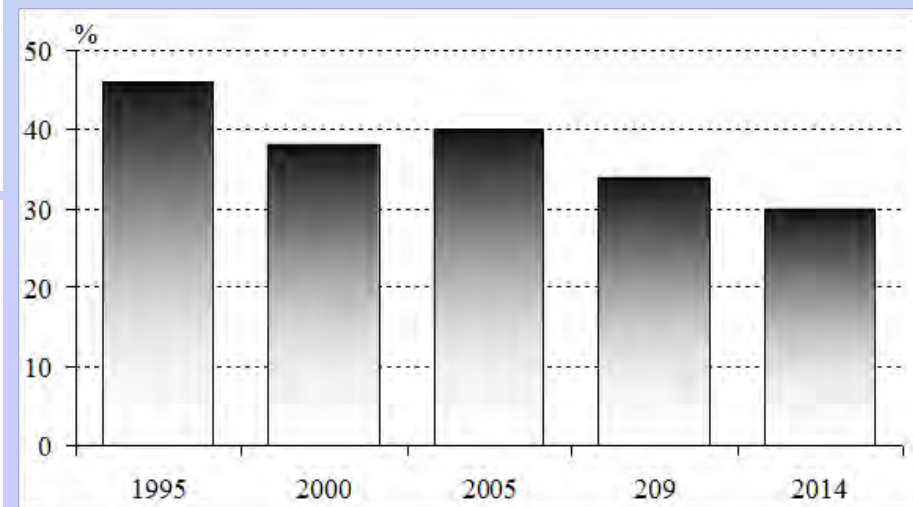


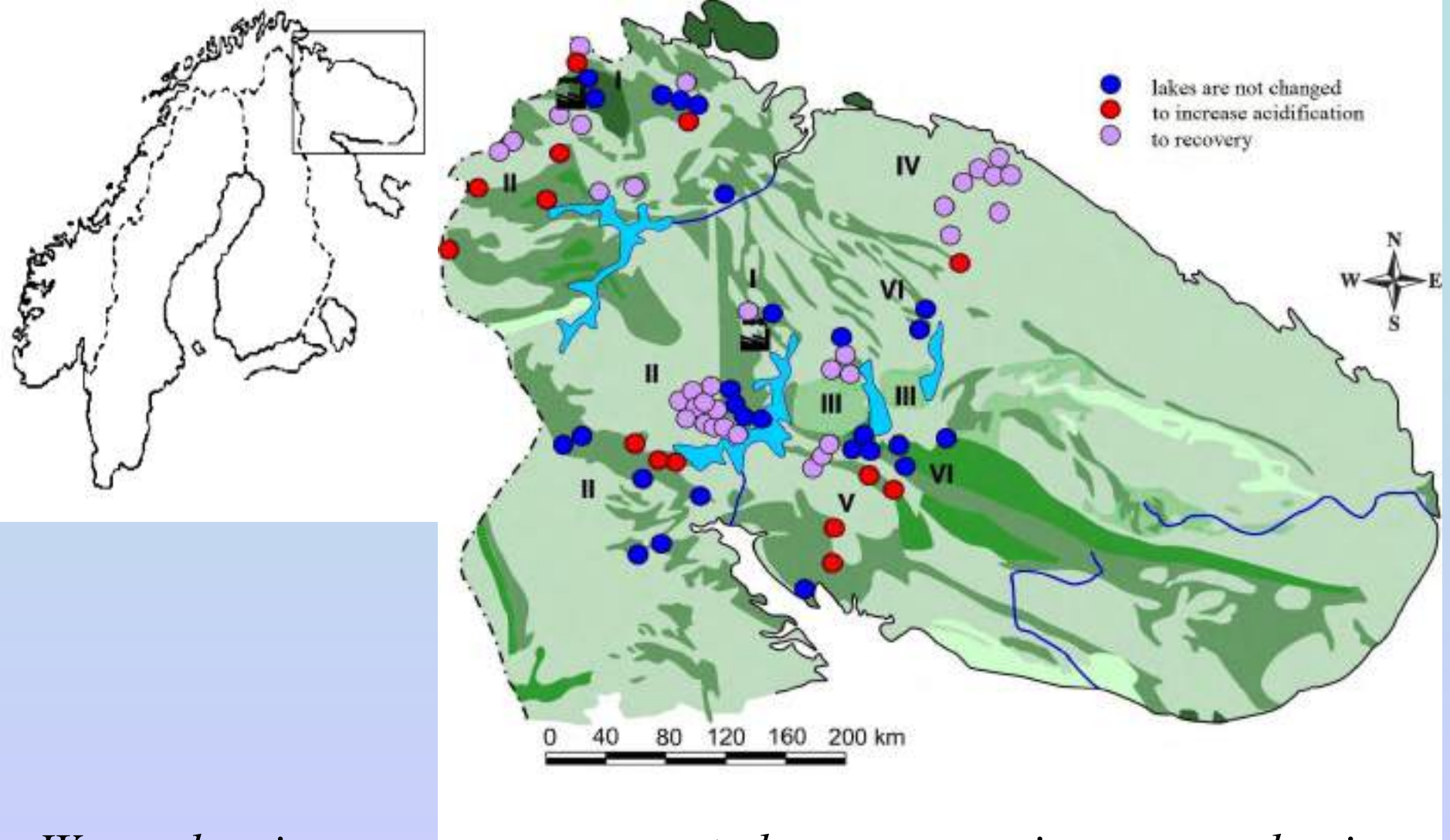
Long-term trends - recovery

(Kola region)



Clex





*Water chemistry responses are **not** always proportionate to reductions in sulfate flow. Exists three scenarios of changes water chemistry: i) further water acidification is progressing. ii) pH and alkalinity levels remain the same; iii) lake water chemistry recovers.*

Character of Spatiotemporal Variations in the Chemical Composition of Lake Water under the Influence of Emission from Copper–Nickel Plants: Prediction of Acidification

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119991 Russia *e-mail: ngashkina@gmail.com

Received July 28, 2015; accepted December 15, 2015

Abstract In this paper, we analyze the influence of variations in the emission of sulfur dioxide and solid substances by the Pechenganikel and Severonikel copper–nickel plants in Murmansk oblast on the chemical composition of lake water and development of acidification. The dynamics of ~100 lakes examined in 1990, 1995, 2000, 2005, and 2009 and response of the chemical composition of the lake waters on the impact of acidifying substances was explored depending on the magnitude of load (distance from the plants), geologically controlled vulnerability of the lake catchments to acid precipitation, and the size of the lakes. Possible further changes in the sulfate concentration and pH values of lake waters were estimated for scenarios assuming an increase or a decrease in sulfur dioxide emission from the plants. It was shown that, in the zone of maximum and high load, a 20% change in sulfur dioxide emission will result in a mean change in sulfate concentration of ± 8 $\mu\text{eq/L}$ (which is comparable with the regional background) and a change in pH value of ± 0.1 in acid-sensitive lakes and will have almost no effect on these parameters in lakes insensitive to acid precipitation.

Keywords: lakes, anthropogenic sulfates, vulnerability, acidification, load variations, dynamics, prediction

DOI: 10.1134/S001670291612003X

Thanks for
attentions !



Relationship between critical load exceedances and empirical impact indicators at ICP Integrated Monitoring sites

Update 2017: preliminary results

Maria Holmberg, [Jussi Vuorenmaa](#), Maximilian Posch
Sirpa Kleemola + NFCs...

Joint ICP Waters & ICP IM Task Force meeting,
Uppsala, Sweden, 11.5.2015



Previous study at ICP IM sites

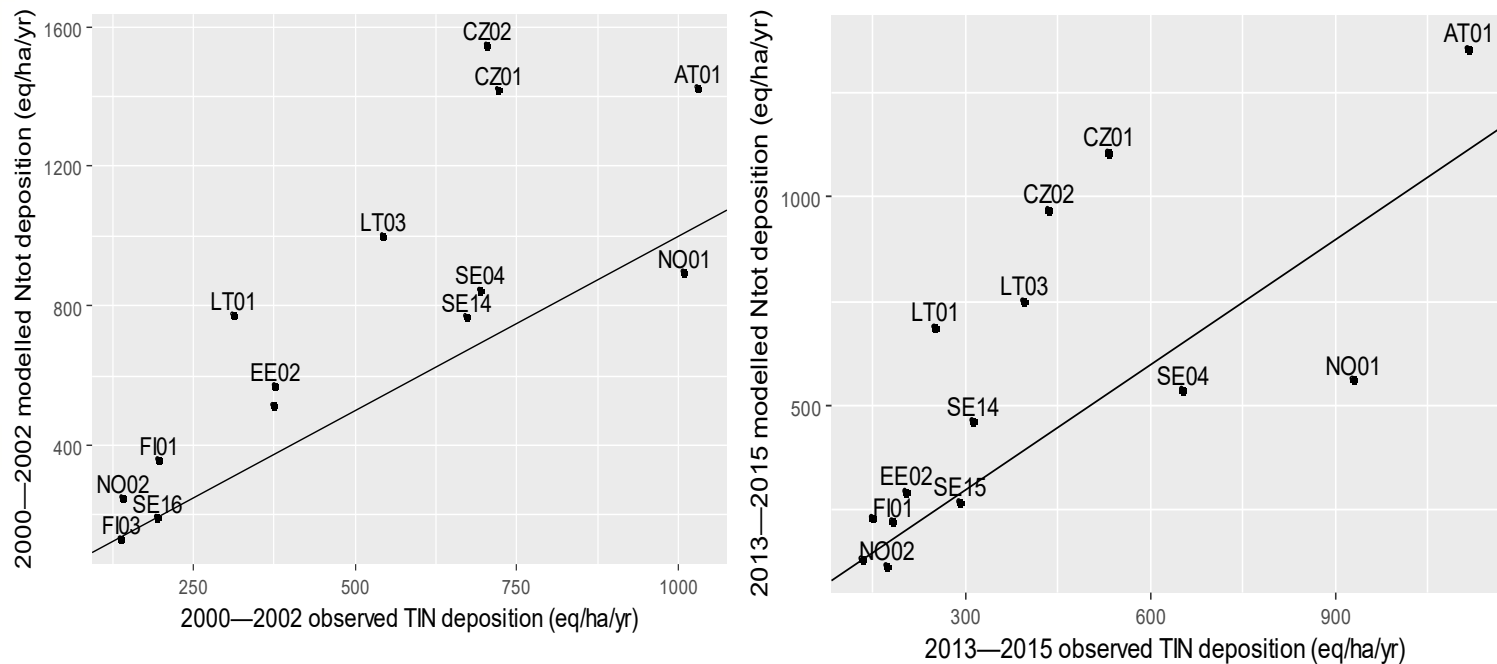
- *Is there a link between modelled critical thresholds and empirical results for acidification parameters and nutrient nitrogen at IM sites?*
- Vuorenmaa & Holmberg (2010, 19th AR) and Holmberg et al. (2013, *Ecol Ind*) concluded that:
 - 👍 Empirical impact indicators as derived from observations in ICP IM catchments were in good agreement with exceedances of critical loads of acidification and eutrophication
 - 👍 Data from the ICP IM thus provided evidence of a connection between modelled critical loads and empirical monitoring results for acidification parameters and nutrient nitrogen.
 - 👍 The collected empirical data of the ICP IM allow testing/validation of the key concepts in the CL calculations.
 - 👍 Increases confidence in the regional scale CLs mapping approach used in the integrated assessment modelling.

Material and methods

- Previous study with modelled N deposition data of 2000 and empirical TIN data of 2000-2002 **was revisited with new data on modelled N deposition (2010) and empirical TIN RW (2013-2015).**
- Critical loads for eutrophication, their exceedances and concentrations and fluxes of total inorganic nitrogen (TIN = $\text{NO}_3 + \text{NH}_4$) in runoff were determined for a selection of 14 sites: AT01, CZ01, CZ02, EE02, FI01, FI03, LT01, LT03, NO01, NO02, SE04, SE14, SE15, SE16.
- The exceedances (ExCLnutN) were calculated as differences between the level of total N deposition ($\text{N}_{\text{tot}} = \text{NO}_3 + \text{NH}_4$) and the mass balance critical loads of nitrogen (CLnutN).

N in deposition #1.

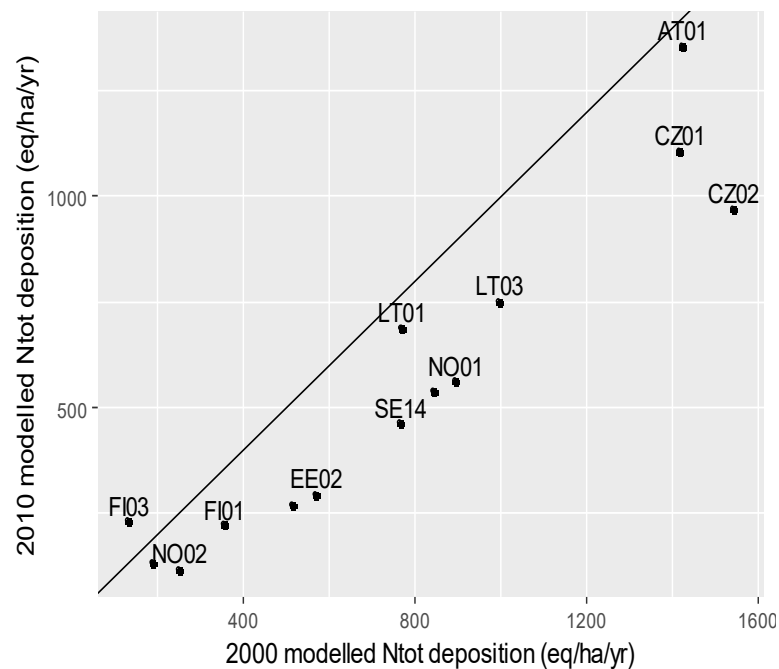
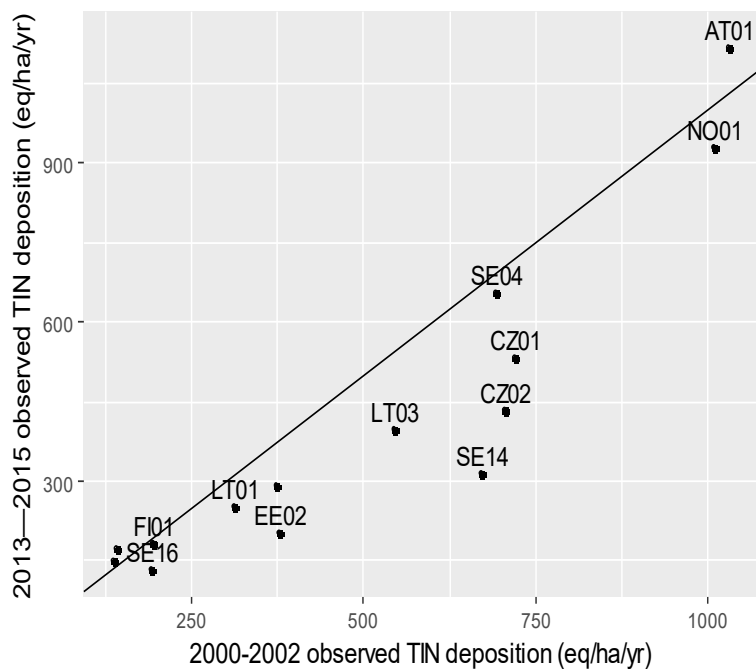
Comparison of modelled to observed N input to the sites in two study periods. The line is drawn at slope 1:1.



For most sites, the modelled N deposition > observed flux in bulk deposition

N in deposition #2.

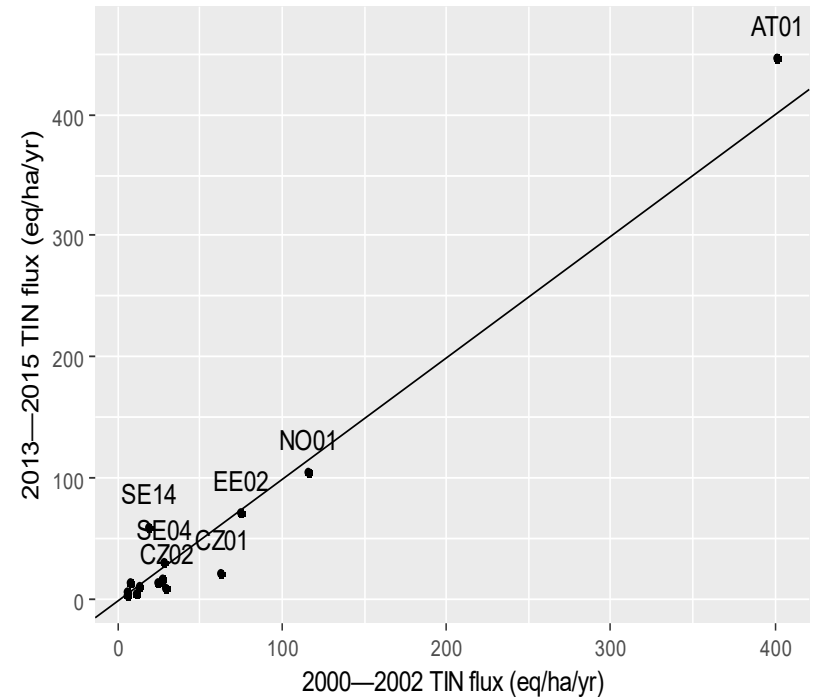
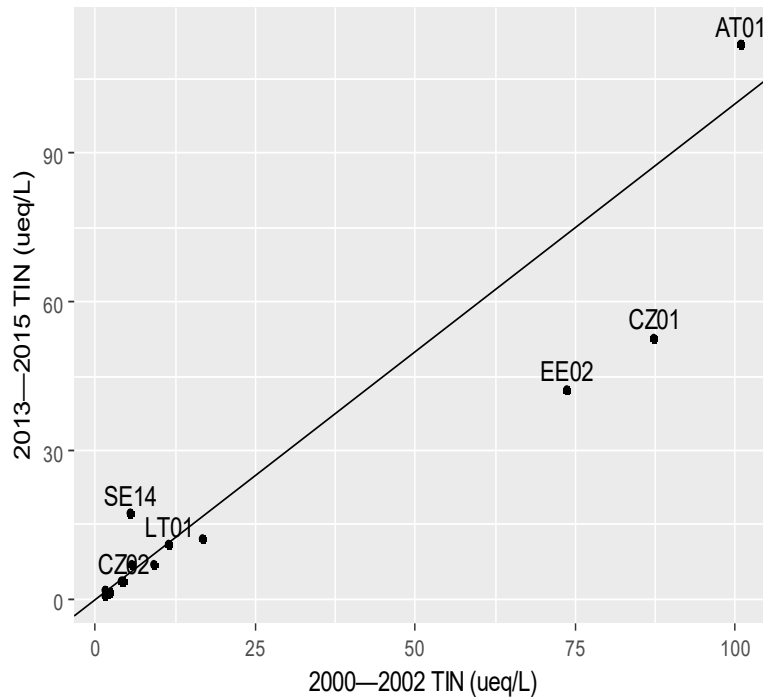
Comparison of N input (observed and modelled) to the sites for the period 2013–2015 (y-axis) versus period 2000–2002 (x-axis)



Both the observed and the modelled estimates for N input to the sites have decreased for almost all sites between the two observation periods.

N in runoff #1

Observed concentration (left) and flux (right) of TIN in runoff, averages for period 2013-2015 (y-axis) compared to those for 2000-2002 (x-axis).



TIN concentrations and fluxes in runoff decreased at most sites between the two observation periods.

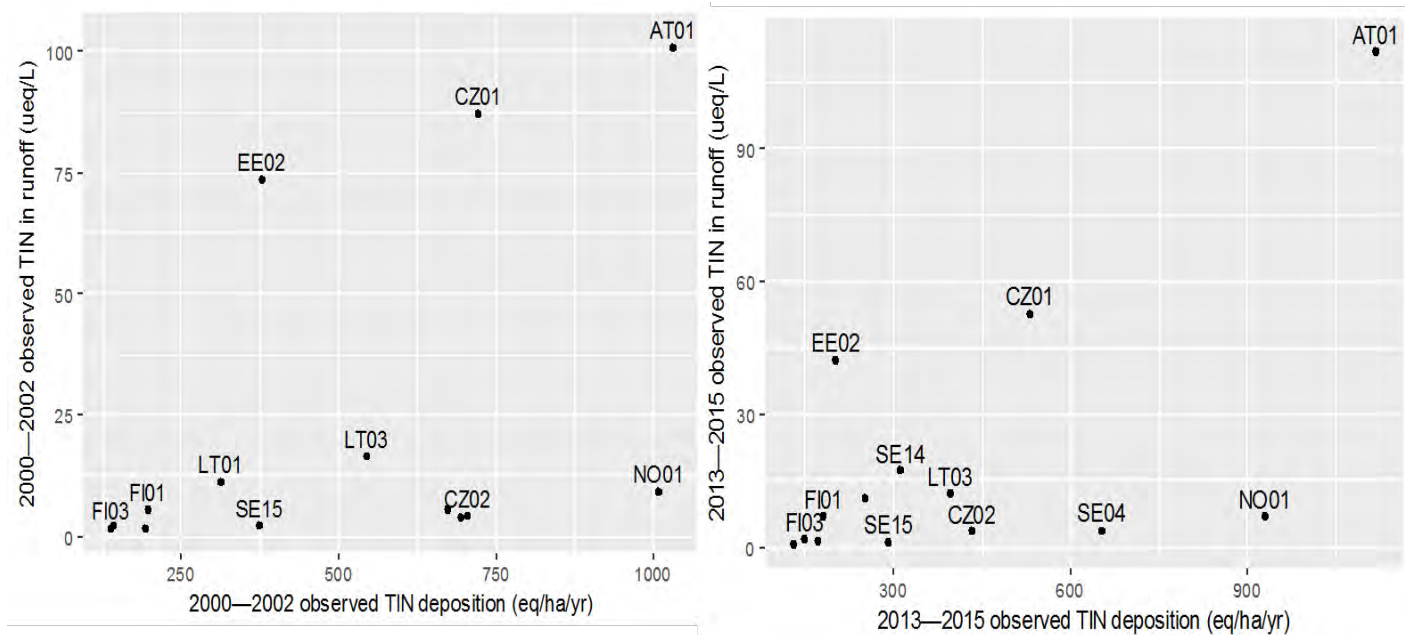
Summary table: 2000 & 2000-2002 vs. 2010 & 2013-2015

			2000-2002	2000-2002	2000	2000	2013-2015	2013-2015	2010	2010
Country	IM Site code	Site	TIN conc. ($\mu\text{eq L}^{-1}$)	TIN flux (eq $\text{ha}^{-1} \text{yr}^{-1}$)	Ntot dep. (eq $\text{ha}^{-1} \text{yr}^{-1}$) (modelled)	ExCLnutN (eq $\text{ha}^{-1} \text{yr}^{-1}$)	TIN conc. ($\mu\text{eq L}^{-1}$)	TIN flux (eq $\text{ha}^{-1} \text{yr}^{-1}$)	Ntot dep. (eq $\text{ha}^{-1} \text{yr}^{-1}$) (modelled)	ExCLnutN (eq $\text{ha}^{-1} \text{yr}^{-1}$)
Austria	AT01	Zöbelboden IP1	100.8	401.3	1424	1117	111.8	446	1355	1049
Czech Republic	CZ01	Anenske Povodi	87.2	62.5	1417	1114	52.8	22.3	1107	804
	CZ02	Lysina	4.2	29.4	1545	1172	3.52	10.2	968	595
Germany	DE01	Forellenbach	102.8	1373.0	1616	1140	46.1	354	1481	1011
Estonia	EE02	Vilsandi	45.4	70.6	570	189	42.3	71.1	292	-45
Finland	FI01	Valkea-Kotinen	5.7	13.4	357	56	7	11.7	220	-141
	FI03	Hietajärvi	1.6	5.7	130	-108	1.86	6.2	228	-982
Lithuania	LT01	Aukstaitija	11.3	7.4	770	465	11	13.7	685	378
	LT03	Zemaitija	16.6	27.4	997	699	12.2	16.8	750	428
Norway	NO01	Birkenes	9.2	115.9	896	442	6.91	105	560	108
	NO02	Kårvatn	2.2	36.5	249	-408	1.31	13.5	113	-530
Sweden	SE04	Gårdsjön	3.7	27.0	845	660	3.72	31.1	535	152
	SE14	Aneboda	5.5	19.1	767	534	17.3	60.2	460	226
	SE15	Kindla	2.0	11.0	514	210	1.13	5.4	268	-36
	SE16	Gammtratten	1.6	5.7	191	-99	0.69	3.3	128	-161

No change or increased, decreased

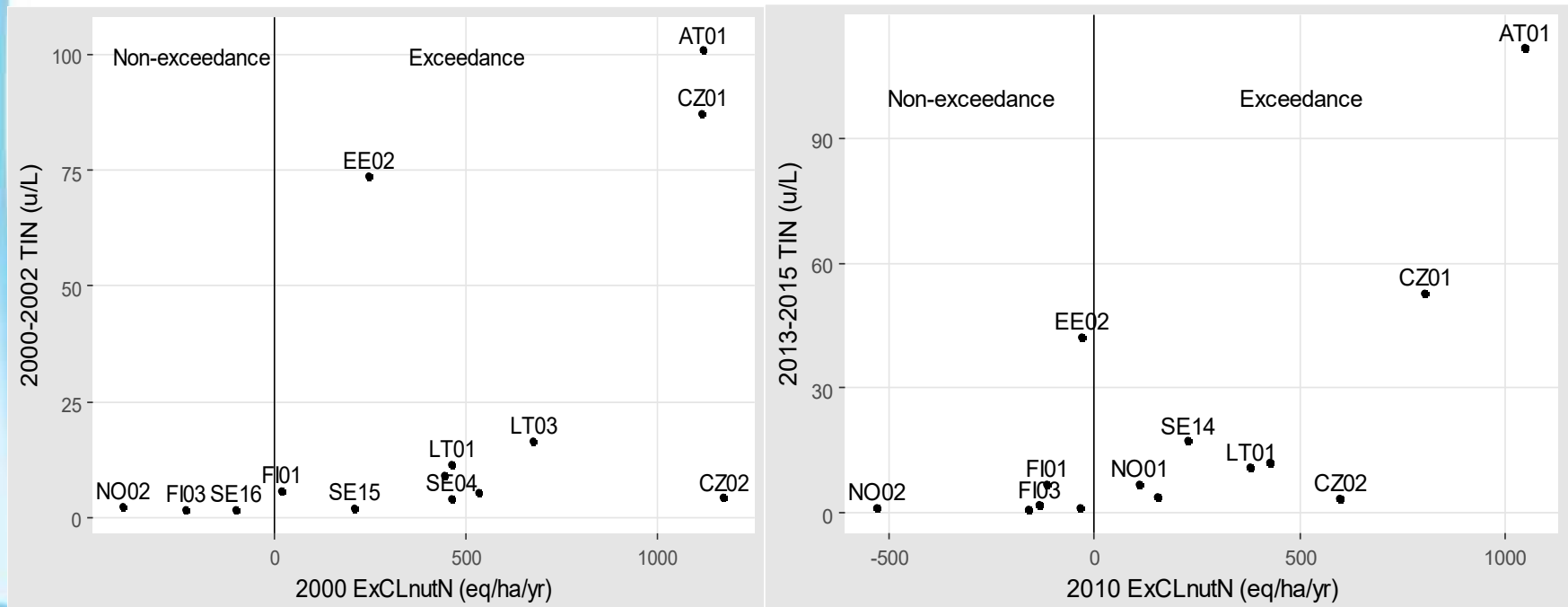
N in runoff #2

The observed concentration of TIN in runoff (y-axis) versus the observed deposition flux of TIN (x-axis). Period 2000 – 2002 in left panel, period 2013 – 2015 in right panel.



In general, concentrations of TIN in output increase with increasing TIN deposition

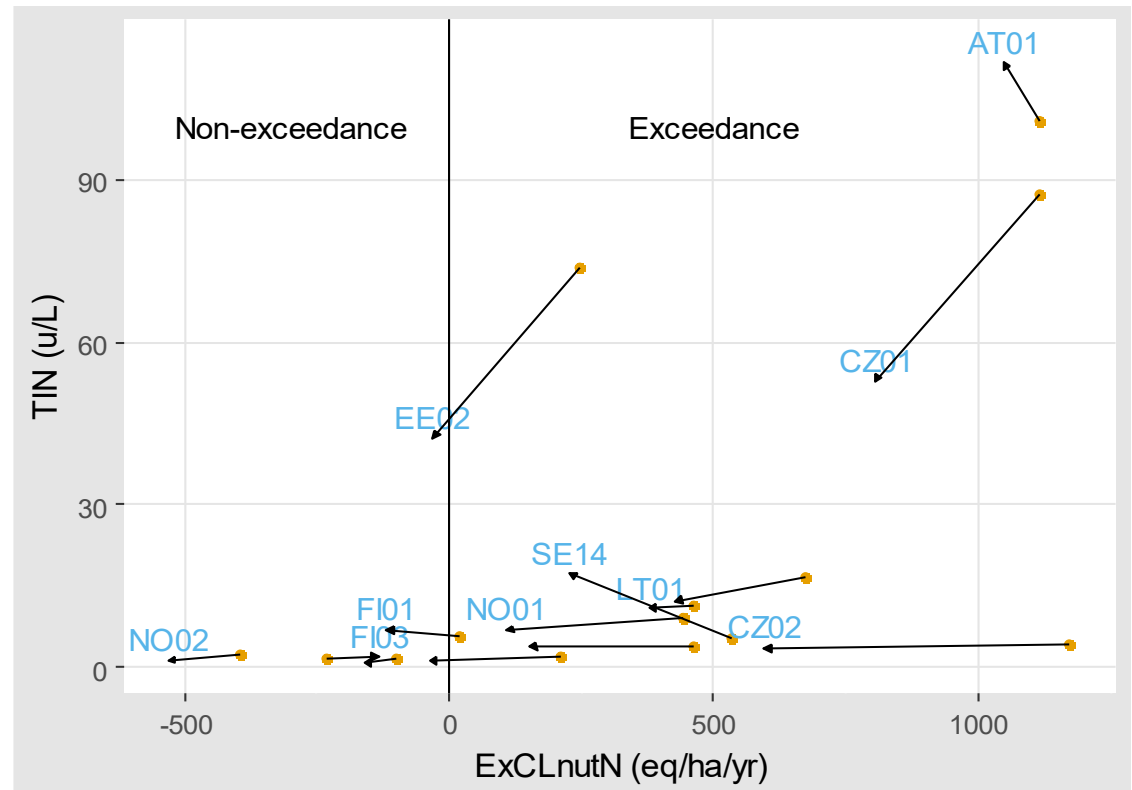
Observed concentration of TIN in runoff (y-axis), average for periods 2000–2002 (left) and 2013–2015 (right), versus the calculated exceedance of critical loads of nutrient N (x-axis), using modelled deposition values for 2000 (left) and 2010 (right).



CLnutN exceeded -> higher TIN concentrations in runoff

The observed concentration of TIN in runoff (y-axis) versus the calculated exceedance of critical loads of nutrient N (x-axis), using modelled deposition values.

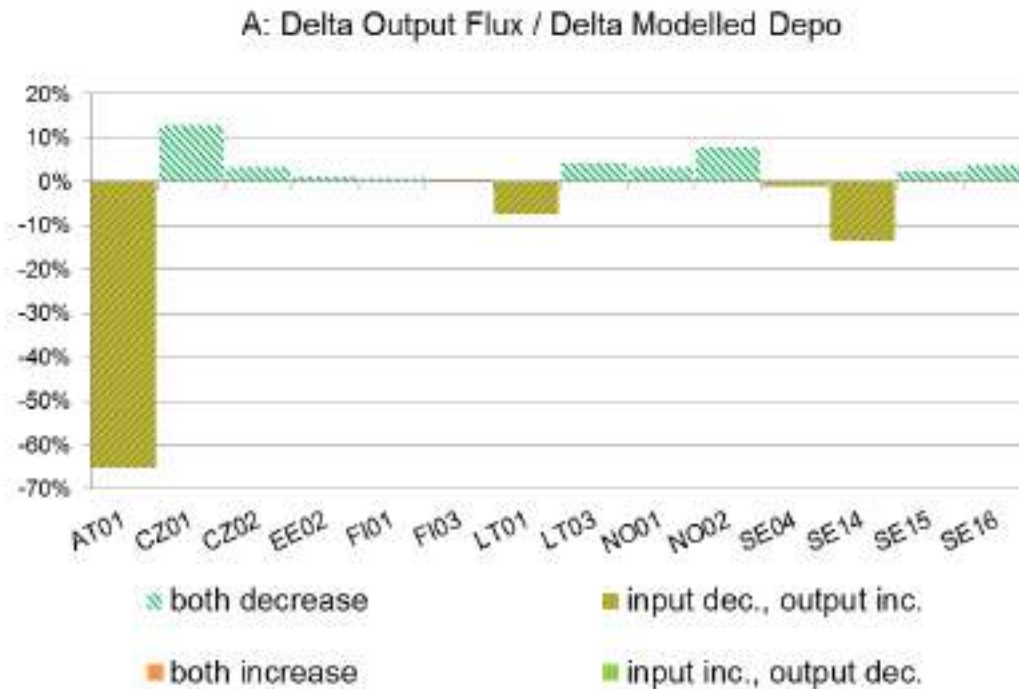
The arrows begin at the locations of the data points for the period 2000-2002 and end at the locations of the data points for the period 2013-2015.



Shift towards less exceedance and lower TIN concentrations in runoff

Comparison of changes in TIN flux in runoff, relative to the change in deposition.

Changes calculated as differences between the values for 2013 – 2015 and those for 2000 – 2002. Relative changes (%) as change in flux divided by change in modelled deposition. This comparison reflects also differences in meteorological and hydrological conditions and forest disturbance regimes for the two periods.



To summarize...



We still conclude that there is a link between modelled critical thresholds and empirical results for nutrient N



Improvement visible

- N in deposition (modelled & measured) and output decreased rather than increased between the two observation periods 2000-2002 and 2013-2015
- A shift towards less exceedance (ExCLnutN)



Work will continue

- A scientific paper in 2019, plans to extend empirical indicators to include vegetation indicators
- Work is also related to the EU/H2020 project eLTER

Thank you



Valkea-Kotinen IM catchment (FI01)
Photo: Jorma Keskitalo

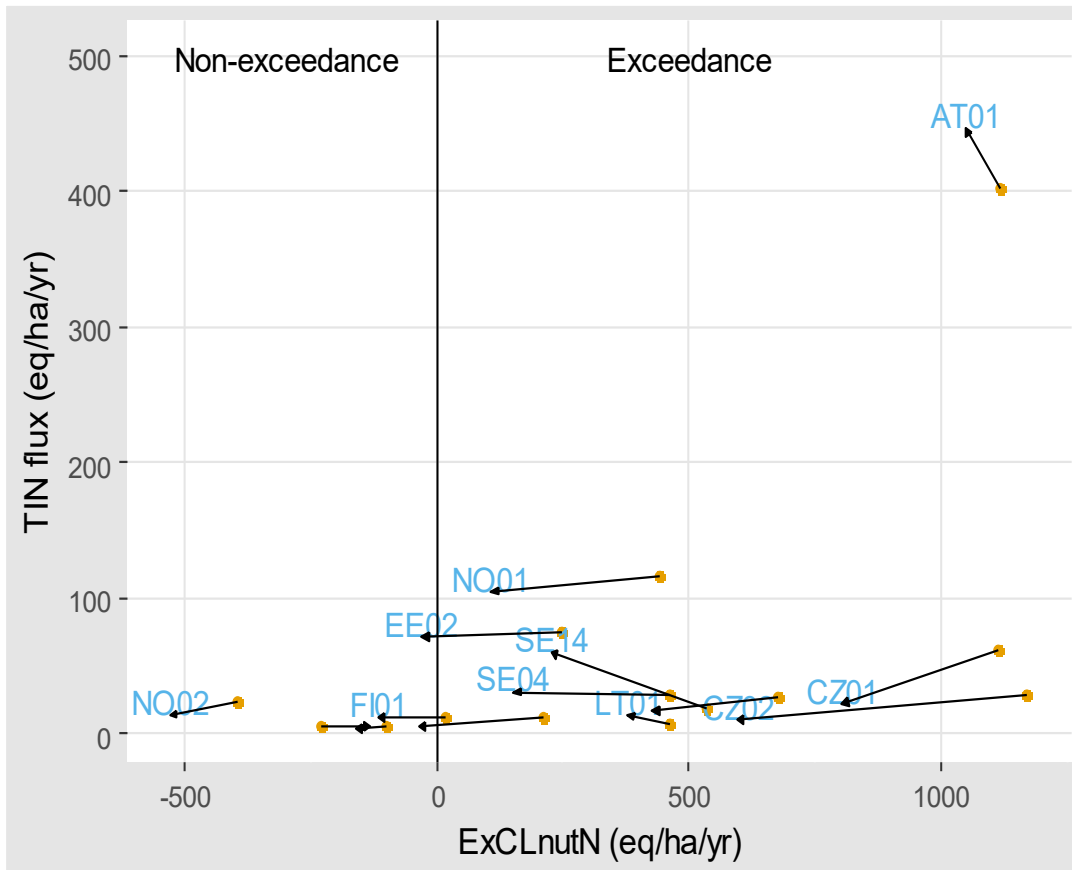


Figure 5b. The observed flux of TIN in runoff (y-axis) versus the calculated exceedance of critical loads of nutrient N (x-axis), using modelled deposition values. The arrows begin at the locations of the data points for the period 2000–2002 and end at the locations of the data points for the period 2013–2015.

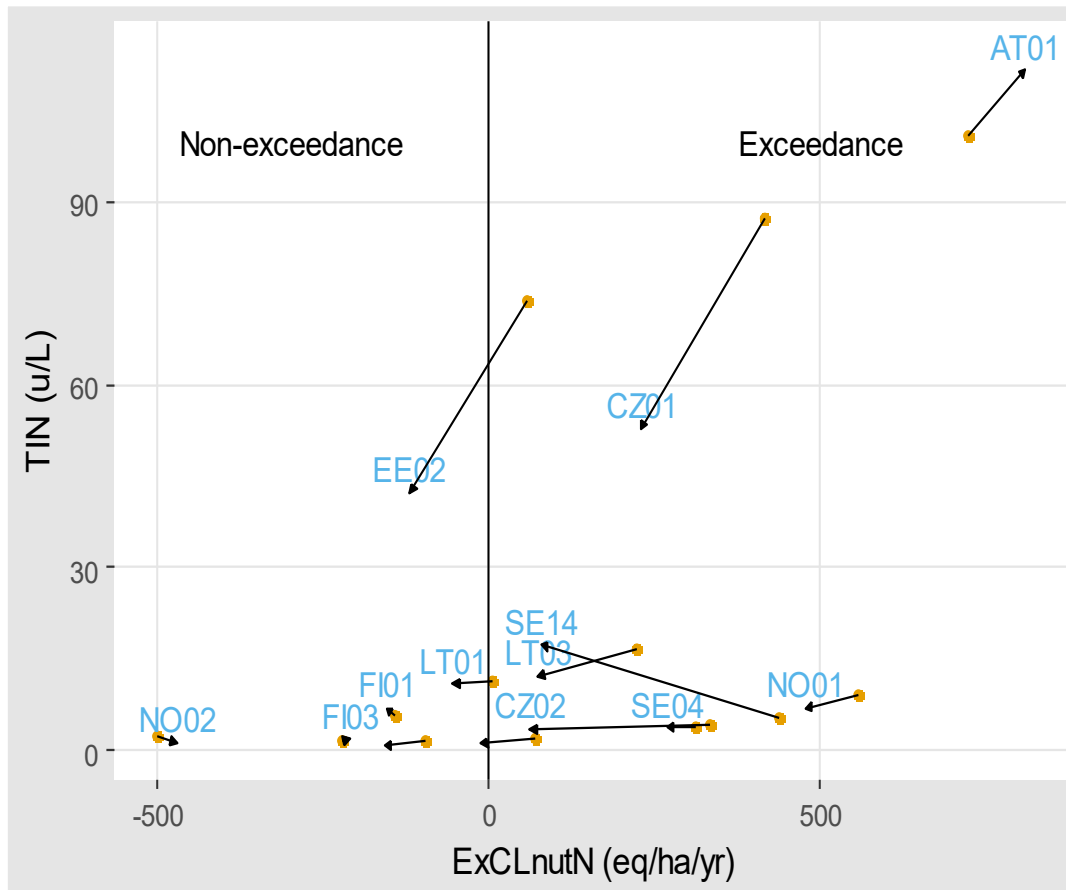


Figure 5c. The observed concentration of TIN in runoff (y-axis) versus the calculated exceedance of critical loads of nutrient N (x-axis), using observed input fluxes for deposition values. The arrows begin at the locations of the data points for the period 2000–2002 and end at the locations of the data points for the period 2013–2015.

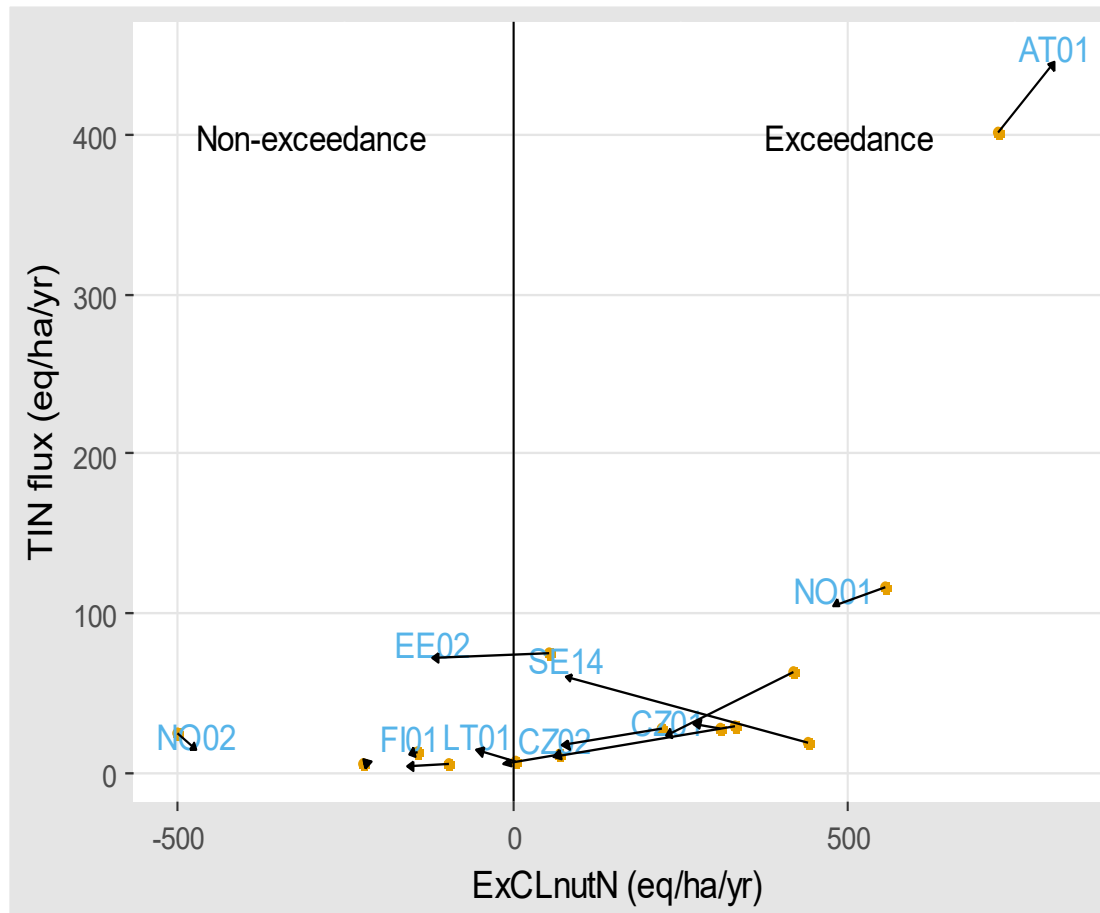


Figure 5d. The observed flux of TIN in runoff (y-axis) versus the calculated exceedance of critical loads of nutrient N (x-axis), using observed input fluxes for deposition values. The arrows begin at the locations of the data points for the period 2000–2002 and end at the locations of the data points for the period 2013–2015.

Intercomparison 1630

pH, conductivity, alkalinity, nitrate, TOC,
major ions and trace metals

ICP Waters report 129/2016

Intercomparison 1630:

pH, Conductivity, Alkalinity, NO₃-N, Cl, SO₄, Ca, Mg, Na, K,
TOC, Al, Fe, Mn, Cd, Pb, Cu, Ni and Zn



The International Cooperative Programme on Assessment
and Monitoring Effects of Air Pollution on Rivers and Lakes
(ICP Waters)
Convention on Long-Range Transboundary Air Pollution



Intercomp number 30



Participation in intercomparison 1630

- 89 laboratories were invited to participate
- 35 laboratories from 20 countries accepted the invitation and submitted results for one or more parameters

Participation

Country	No. of labs.	Country	No. of labs.
Austria	1	Netherlands	1
Canada	1	Norway	1
Czech Republic	1	Poland	2
Estonia	1	Russia	6
Finland	2	Serbia	1
France	1	Spain	1
Germany	6	Sweden	1
Ireland	2	Switzerland	1
Italy	1	UK	3
Lithuania	1	USA	1

Preparation of the samples

- Water (50 L) was collected from Hakadal Verk (north of Oslo)
- Filtered (0,45 μm), stored at room temperature and left to equilibrate with atmosphere
- pH was lowered with addition of HCl and H_2SO_4 (sample set AB). Sample A diluted with MQ water.
- Sample set CD was spiked with metals and conserved (0.33 % HNO_3)

Results

				Acceptable Limit	Acceptable results (%)				
Variable	Sample pair	Sample 1	Sample 2	%	1630	1529	1428	1327	1226
pH	AB	6.00	6.04	3.3	56	64	68	52	59
Conductivity.	AB	4.08	4.59	10	77	89	93	78	72
Alkalinity.	AB	0.042	0.046	20	46	75	26	63	48
Nitrate + nitrite-nitrogen.	AB	276	309	20	71	88	14	0	52
Chloride.	AB	2.45	2.81	20	87	97	93	78	79
Sulphate.	AB	8.07	9.2	20	90	97	87	77	80
Calcium.	AB	3.9	4.42	20	93	97	97	85	75
Magnesium.	AB	0.41	0.47	20	89	100	87	82	74
Sodium.	AB	1.87	2.14	20	96	97	97	91	84
Potassium.	AB	0.43	0.49	20	86	97	97	70	81
Total organic carbon.	AB	4.09	4.65	20	81	70	82	78	76
Aluminium.	CD	255	270	20	78	89	78	89	79
Iron.	CD	186	190	20	91	81	74	72	70
Manganese.	CD	48	51.7	20	91	84	88	78	89
Cadmium.	CD	6.87	7.73	20	90	100	84	85	84
Lead.	CD	7.23	8.34	20	86	77	80	71	77
Copper.	CD	24.6	28	20	86	93	88	84	86
Nickel.	CD	10.2	11.5	20	90	97	92	83	78
Zinc.	CD	34.4	38.1	20	81	83	79	60	61
Total					81	88	80	73	74

Units: Conductivity: mS/m
Alkalinity: mmol/l
Nitrate+nitrite-N: µg N/l

Cl. SO₄. Ca. Mg. Na. K. TOC: mg/l
Al. Fe. Mn. Cd. Pb. Cu. Ni. Zn: µg/l

pH

Analytical variable and method	Sample	TRUE Value		No. lab.		Avg/Std.av.	
		S. 1	S. 2	Total	Omm.	Sample 1	
pH	AB	6.00	6.04	32	0	6.04	0.25
Electrometry				25	0	6.04	0.24
Stirring				4	0	5.89	0.17
Equilibration				2	0	6.37	
Other method				1	0	6.05	
Alkalinity	AB	0.042	0.046	24	8	0.041	0.007
Gran plot titration				9	1	0.044	0.008
End point titration				8	3	0.039	0.005
End point				2	2	0.079	
Other method				2	1	0.040	
Colorimetry				1	1	2.400	
End point 5.4				1	0	0.036	
End point 5.6				1	0	0.042	

Conclusions

- Accuracy in determination of major ions was very good (> 85 % had target accuracy < 20 %)
- Accuracy for trace metals and Fe was good (> 80 % had target accuracy better than 20 %)
- Accuracy for NO₃, TOC and Al was fair (70-80% had target accuracy < 20 %)
- Accuracy for pH and alkalinity was poor (56 % or less had target accuracy better than 0.2 units and 20 %, respectively)

Intercomparison test 1731

- Will include phosphorous
- Free for labs within UN-ECE and EECCA that deliver results to national monitoring programs. Others have to pay a minor fee

Plan for Intercomp 1731

- 24 May : Deadline for registration in the database
- Week 24: Samples delivered to the participants
- 5 September: Deadline for reporting results in the database
- Week 46: Report available
- Interested parties can contact carlos.escudero@niva.no



CONVENTION ON LONG-RANGE TRANSBOUNDARY AIR POLLUTION

International Cooperative Programme on
Integrated Monitoring of Air Pollution Effects on Ecosystems
25th ICP IM Task Force Meeting & Workshop

Organiser and Location
Swedish University of Agricultural Sciences, SLU
Uppsala

Swedish responsible
Swedish Environmental Protection Agency

Thursday 11 May
II. ICP IM Task Force meeting



ICP IM Task Force

Explanatory notes to the agenda

1. Opening of the meeting
2. Approval of minutes of the 24th ICP IM TF in Asker, Norway, 2016

ICP IM Task Force

Explanatory notes to the agenda



3. Report from Chairman and Programme Centre

Presented at the Working Group on Effects 2016

WGE chairman reported from EB

CLRTAP Strategy reviewed 2016

Monitoring specially mentioned

Emphasis on policy oriented messages and further actions by Parties

Joint report presented and accepted

Outreach to EECCA and SEE and translations

The Assessment report Report was presented in Brussels and Batumi



ICP IM Task Force

Explanatory notes to the agenda

Item 3 continues

Highlights in joint report

World Health Assembly resolution road map

Economic losses ozone impacts on wheat yield revealed Parties presented measures for emission reductions

Climate warming may increase the vulnerability of ecosystems

Excess N may have negative effects ; reductions policy relevant

Hg in fish over recommended values for humans

CH₄ on hemispheric scale to reduce ozone damage

Climate measures include reductions of methane and nitrous oxide

CCE meets problems with financing

ICP IM Task Force

Explanatory notes to the agenda



Item 3 continues

Two year workplan

A joint workplan together EMEP and WGE, currently 2016-2017

Next period: 2018 - 2019

Changed meeting structure;

EMEP and WGE together in September 2017

EB somewhat back to earlier schedule but several meetings and a meeting in December 2017

EMEP and WGE homepages should have a common portal. But, now probably WGE homepage under UN ECE.



Item 4

Highlights from ICPs at WGE 2016

Forests - Database and evaluations priority; data used in publications

Waters - biodiversity would improve, in line with chemical recovery

Materials - trends in air pollution, corrosion and soiling showing a 50% decrease in corrosion since 1987 but in recent years poor further improvements

Vegetation - report on revised ozone risk assessment methods for vegetation; report analysis of tropospheric ozone - TOAR

IM - nitrogen still accumulates in the ecosystems and output is very low compared to input. S released from catchment soils. HM still accumulating in ecosystems.



Item 4 continues

Highlights from ICPs

- MandM** - Call for calculating biodiversity critical loads and by updating the acidification and eutrophication critical loads
- TF Health** - WHO AIRQ+ software to quantify the health impacts of air pollution
- JEG** - dynamic modelling in particular with respect to nutrient nitrogen biological response, aquatic biodiversity and climate

Updating of scientific and technical knowledge

The Joint report highlighted impacts of air pollution on human health. Updated scientific and technical knowledge as presented above. Reports with stronger connections to societal needs and also with direct response to the Convention requests.



Joint EMEP SB and WGE Bureau meeting

Item 5 Update scientific knowledge etc.

Joint thematic sessions three issues:

the linkages between climate change and air pollution; benzo(a)pyrene and wood burning; and ozone. Contribution of methane, black carbon and other climate change forcers.

Co-operation AMAP, Climate and Clean Air Coalition, EANET, etc.

CCE is financially problematic and the Netherlands intend to stop 2018.

Election WGE; Ms. Isaura Rabago (Spain) new Chair for a two-year term. Re-elected vice chairs Mr. Jesper Bak (Denmark), Ms. Gudrun Schuetze (Germany), Ms. Sabine Augustin (Switzerland), Mr. Thomas Dirnböck (Austria) and Mr. Nebojsa Redzic (Serbia).

Next WGE, No. 36, is scheduled for 11-15 September 2017.



Joint EMEP SB and WGE Bureau meeting

Item 6

The ICP review concludes the importance of continuing monitoring and enhances the co-operation between ICPs as well as between WGE and EMEP.

CLRTAP launched a summary report, *Towards Cleaner Air*

Next Workplan is for 2 years with content reported by May 20.

WGE homepage; a page under UN ECE; "unece-wge.org"

HTAP, consider interactions between air pollution and climate including ozone, atmospheric compounds, Hg, ozone and POPs.



7. CLRTAP Workplan

Workplan 2016-2017

WGE 2017 joint EMEP and WGE meeting

ICPs present the obligations and monitoring work content for the leading countries and the PCs.

Mandates to be presented related to the CLRTAP strategy and an update of the MoU.



7. CLRTAP Workplan

ICP IM Mandates

There is a template for revised mandate for a WGE
ICP/TF/Centre

Lead country: Sweden
ICP IM PC at SYKE, Finland

Determination and prediction of the state of ecosystems and their long-term changes with respect to the regional variation and impact of selected air pollutants, with special attention to effects on biota.



ICP Integrated Monitoring

Ecosystem investigation

Monitor state and changes with causative explanations

Develop and validate models

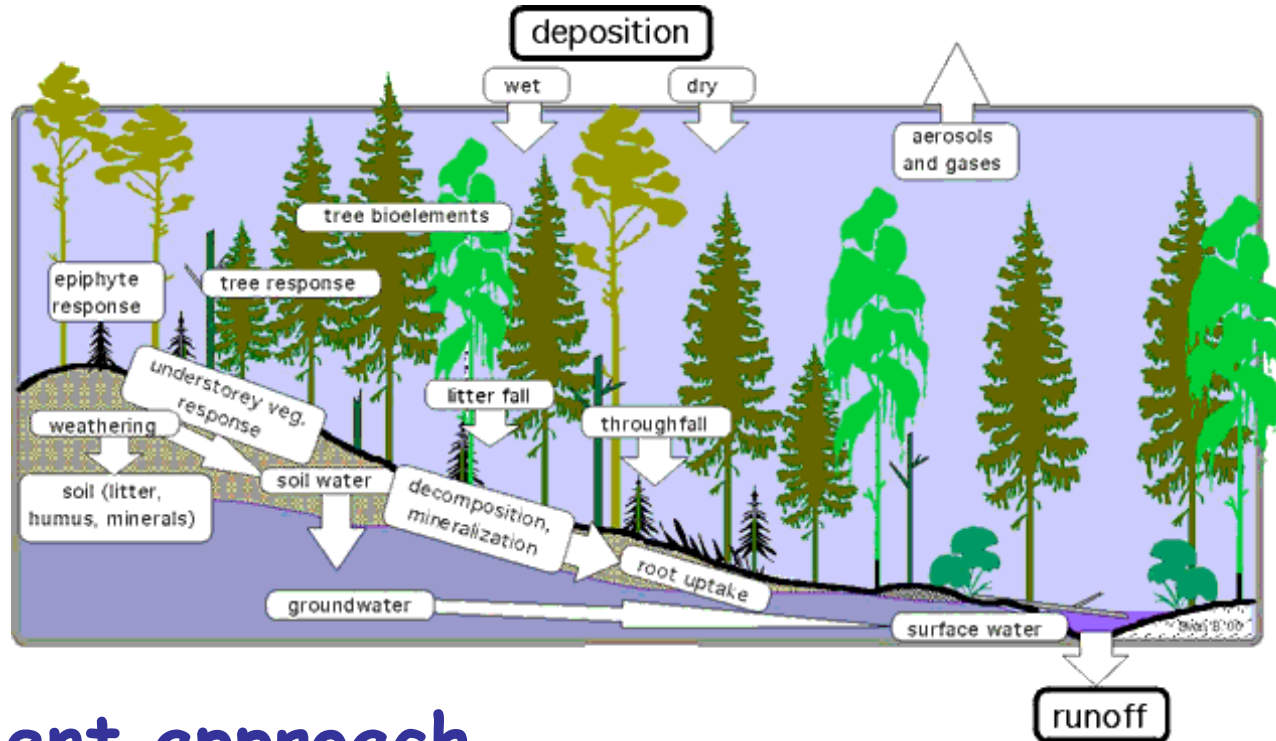
Detect changes by biomonitoring

Cause-effect approach

Cross-media flux approach



Integrated Monitoring



Catchment approach
Budget calculations
Process oriented



Item 8 - ICP IM activities 2016

The ICP IM activities in 2016/2017 included work on S and N in-/output budgets for IM sites, dynamic vegetation modelling, the Annual ICP IM Report No 25 and common items such as outreach and common activities (ECE/EB.AIR/WG.1/2016/8).

Representatives have participated in a number of international meetings.

Database update (later item 10)



Item 9

Reports from WGE and other ICPs

Possibly

Heleen de Wit, ICP Waters; Recent activities of ICP Waters



Item 10

Database

Item 11

Reports 2017

26th ICP IM Annual Report

Progress Report on connections between calculated CL exceedances and observed impacts of N. ICP IM PC.

Progress Report on concentrations of heavy metals in important forest ecosystem compartments (Sweden and ICP IM PC)

Report on mercury in the aquatic environment; Joint Report together with ICP Waters with ICP IM contribution

Comparison of activities across continents and regions; outreach and hemispheric view will relate to HTAP, EECCA, SEE and LTER as regards ICP IM



Item 11

Reports 2017

Common activities

Contributions to WGE/EMEP Joint report with policy relevant results

Revised Mandates

Technical reporting



Item 12

Technical Reports and deliverables

The Joint report should present important results with policy relevance to **May 20, 2017**.

Medium term workplan for ICP IM has plans for the year 2018 and 2019 which will be covered under item 15; technical report submitted **June 1, 2017**,

Co-operation between WGE and EMEP with a joint meeting **11-14 September 2017**,

Preparations of the 36th session of WGE, ordinary ICP activities will be given a short overview with main focus on common items and thematic scientific reporting, especially policy relevance of the ICP results



Item 14

Co-operation with other ICPs and external projects/organizations:
LTER-Europe, ALTER-Net, LifeWatch, Expeer Natura 2000, etc.

More meetings could be found on;

<https://www.unece.org/environmental-policy/conventions/envlrtapwelcome/meetings-and-events.html#/>

LTER Europe

ILTER

ALTER-Net

Other future work priorities



Item 15 Work plan 2017 and 2018-2019

2018

Scientific paper on long-term trends in atmospheric deposition and runoff water chemistry of S and N compounds at ICP IM catchments in relation to changes in emissions and hydrometeorological conditions

Scientific paper on dynamic modelling on the impacts of future deposition scenarios on soil and water conditions in ICP IM catchments

Reporting of ICP IM activities to WGE



Item 15 Work plan 2017 and 2018-2019

2019

Report on dynamic modelling on the impacts of deposition and climate change scenarios on ground vegetation (ICP IM)

Scientific paper on the relationship between critical load exceedances and empirical ecosystem impact indicators (ICP IM).

Reporting of ICP IM activities to WGE



Item 16

Financing/external applications

Trust fund - small amendments, slight decrease

Parties support most important

EU Horizon 2020

FTP and SRA

Item 17

Next Task Force meeting



*ICP Integrated Monitoring of Air
Pollution Effects on Ecosystems -
ICP IM*

*Next Task Force
2018*



in May in Poland





Item 18

Other business



Item 18

Other business

Sensitive issue

Lars Lundin will retire mainly due to Swedish Regulations

and have a suggestion

Ass. Professor Ulf Grandin will take over

This is recognised by SE EPA and SLU



Item 19

Conclusion and end

Thank You



The End



IM data submission and database status

Sirpa Kleemola, SYKE
ICP IM Task Force meeting,
Uppsala, Sweden
May 9-11, 2017

ICP IM network

- data submission for at least part of the ‘data period’ 2011–2015: Austria, Belarus, the Czech Republic, Estonia, Finland, Germany, Ireland, Italy, Lithuania, Norway, the Russian Federation, Spain, Sweden, [Switzerland](#) and site on Crimean peninsula.
- Poland is re-joining and will include one or more sites to the network.
- Total 43 sites from 15 countries.
 - Canada, Latvia and United Kingdom have not reported data for recent years

Data reported in 2016/2017

For data year 2015

- 13 countries have reported 2015 data so far
 - Austria, Belarus, the Czech Republic, Estonia, Finland, Germany, Italy, Lithuania, Norway, the Russian Federation, Spain, Sweden, Switzerland.
 - Some sites/subprogrammes missing from above mentioned countries
- For details please see Draft Annual Report 2017

Data reporting in 2017

- Deadline 1st of Dec 2017 for 2016 values
- No changes in reporting formats
- Some changes to biological data reporting done in 2010, raw data, not averages, examples and reporting guidelines available
- For new Input programme data per area+subprogramme, e.g. FI01_RW_2016.txt or FI01_RW_2016.xls

Thank you for your attention!