

1 **Habitat characteristics between different clusters of wolf (*Canis lupus*)**
2 **activity before and after brown bear (*Ursus arctos*) emergence in**
3 **Central Sweden.**



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36

37 ABSTRACT

38

39 We describe the habitat used by wolves (*Canis lupus*) at different clusters of activity, namely at
40 kill sites, bed sites and other sites, before and after brown bears (*Ursus arctos*) emergence from
41 dens in central Sweden. Field protocols were gathered in previous predation studies from 2010
42 to 2014 of both species in Scandinavia. Wolves of three different packs and brown bears which
43 overlapped their territories were GPS collared. Field protocols resulted in 182 clusters of
44 activity and were further divided into different cluster types: kill sites (n=47), bed sites (n=65)
45 and other sites (tracks and nothing, n=11, n=59, respectively). We established two main
46 thresholds in order to select wolf kill sites and other clusters, (i) first, a cutoff directly related to
47 seasonal variation of brown bears – winter season, when bears are in their winter dens and –
48 spring season, when bears come out of den, (ii) second, a cutoff related to the life cycle of their
49 main prey, moose (*Alces alces*) in Scandinavia. We used multinomial logistic regression and
50 Akaike’s information criterion (AIC) to select the best models. We picked those models within
51 $\Delta AIC < 2$. The results revealed that kill sites and bed sites were common in mature forest (the
52 availability of land use is not considered in the comparison). However, the frequency of other
53 sites tend to be similar in both mature forest and other (clear-cuts and water related areas).
54 Human population density was high in bed sites and other sites compared to kill sites during
55 winter. In contrast, kill sites occurred in places with moderate human density in both seasons
56 and near roads within young forest compared to mature forests ($221 \text{ m} \pm 188$ and 283 ± 207 ,
57 respectively). Clusters type differed mainly due to the vegetation type, followed by human
58 population density, whereas was lower in kill sites than other clusters type. Finally, human
59 population density was the only variable which had effect, showing that kill sites within mature
60 forest were located in areas with higher human density, while those in open areas (clear-cuts and
61 water related) were located in areas with low human density. In conclusion, the wolf habitat use
62 was influenced by vegetation type and human population density, but not by the presence of
63 brown bear.

64 **Keywords:** vegetation type, wolf clusters, kill sites, wolves, brown bears, protocols

65

66 INTRODUCTION

67

68 Wolves are considered habitat generalists, occupying a wide variety of land-use and vegetation
69 types (Mladenoff *et al.* 1995; Mech and Boitani, 2003). Nevertheless, they show clear
70 preferences for specific kill sites and travel routes (Kunkel and Pletscher 2001; Husseman *et al.*
71 2003). A good understanding of how wildlife use available space is important for their
72 management and conservation, as it provides insight into the ecological requirements of the
73 species. For example, Rostro-Garcia *et al.* (2014) found that, for cheetahs, there is a trade-off
74 between prey abundance and predator avoidance in terms of habitat selection within their home
75 range. Ecological requirements such as landscape attributes (Balme *et al.* 2007), resource
76 availability (Yurewicz *et al.* 2004) and inter- and intraspecific interactions (Palomares *et al.*
77 1999) should be considered in order to obtain an integrated understanding of space use.

78

79 Scandinavia's boreal forest is among the most intensively exploited forest in the world, with
80 very small amount of prime forest left (Linnell *et al.* 2000). The fragmentation of the boreal
81 forests results in a set of clear-cuts, transitional woodlands and mature forest (Guraire *et al.*
82 2011), on a 70-80 years cycle with the use of heavily mechanised system (Linnell *et al.* 2000).
83 Such forest modifications have important effects on biodiversity (Essen *et al.* 1992). A clear
84 example of such modifications are the direct impacts on predator-prey relationships (Wittmer *et*
85 *al.* 2007), where habitat loss and fragmentation (Chubbs *et a.* 1993; Smith *et al.* 2000), limiting
86 resource availability whereas animals seeking riskier environments increasing predation
87 (Wittmer *et al.* 2007). Human presence is likely also an important factor for wildlife in modified
88 boreal forests. A large number of gravel roads is built and used for commercial logging and
89 forest management practices (Sand *et al.* 2008). These roads are used for travelling by wolves

90 (Zimmermann *et al.* 2014) and may enhance the chances of encounter with their prey
91 (Kauffman *et al.* 2007, Courbin *et al.* 2013). Forest roads and habitat fragmentation increases
92 the number of transitional habitats and successional growth of deciduous plants used by
93 ungulates browsers (Edenius *et al.* 2002). Human population density is low within the
94 distribution area of the Scandinavian wolf population, but, can affect wolf space use
95 (Zimmermann *et al.* 2014). The main ungulate prey species in the area are moose *Alces alces*
96 and roe deer *Capreolus capreolus* (Sand *et al.* 2005, 2008; Zimmermann *et al.* 2014). Moose is
97 an abundant browser which have preference from young forest, indirectly benefitting from
98 forestry practices (Edenius *et al.* 2002).

99 Ballard *et al.* (2003) argued that interactions between competing species of predators are an
100 important driver of space use. There are two main types of interactions: (i) indirect, through
101 resource competition when both species use the same food resource (Ballard *et al.* 2003;
102 Mattison *et al.* 2011) and (ii) direct, such as interespecific killing (Palomares & Caro, 1999;
103 Ballard *et al.* 2003). Wolves and bears use commom resources in a similar way (Ballard *et al.*
104 2003). In Katmai National Park (Alaska), Smith *et al.* (2003) showed that wolves can harass,
105 and displace brown bears from carcasses and even steal fish from Grizzly bears. There are, on
106 the other hand, many cases documented in which american bears [Grizzly] usurpate carcasses
107 from wolves (Gunther *et al.* 2004). In Scandinavia, the majority of interactions between brown
108 bears and wolves are indirect (Milleret C, 2011) with only a few direct interactions being
109 recorded to date (Steyaert *et al.* 2010). Understanding the ecological interactions between
110 wolves and brown bears in boreal forests can provide insights into their ecological roles as top
111 predators, including the relationship with their prey which may be important to the conservation
112 and management of these carnivores.

113 Nowadays, as a result of studies on the ecology and conservation of large carnivores many
114 people see an both an inherent intrinsic value in carnivores and perceives their role as important
115 in ecosystem functioning (Miller *et al.* 2001). At large scale, large carnivores are threatened
116 with both population size and their distribution ranges declining around the world (Ripple *et al.*

117 2014). At the continental scale, large carnivores may be stable or increasing their abundance in
118 many countries of Europe (Chapron *et al.* 2014) but possibly losing their functional value at
119 ecosystem level (Gilroy *et al.* 2015).

120 At local scale and focusing in the Scandinavian Peninsula (Norway and Sweden), the population
121 of large carnivores declined strongly throughout the 19th and 20th centuries. The eradication
122 was directly caused by persecution, including bounties, and indirectly by elimination of prey
123 (Schwartz *et al.* 2003). Nowadays there is a small but growing wolf population of 450 wolves
124 (winter 2014-2015, Anon. 2013) and a stable brown bear population of 3200 bears in Sweden
125 (for 2008, Kindberg *et al.* 2011). The main prey, the moose, is present all across Sweden
126 (except in the Island of Gotland) being the largest herbivore in the country (Jensen 2004). The
127 Swedish population is estimated to be 200.000–300.000 individuals at the end of the hunting
128 season (Singh *et al.* 2014).

129 Currently, the techniques to track animals are developing very fast (Urbano *et al.* 2010). GPS
130 (Global positioning system) technology make it possible to track individuals everywhere
131 providing greater knowledge of animal behaviour (Frair *et al.* 2010). In Scandinavia, every year
132 wolves and bears are captured and collared with GPS radio collars within the frame of a long
133 term research programs. The spatial locations can be used to identify clusters of local activity,
134 potentially allowing the identification of kill sites (Anderson and Lindszey, 2003; Sand *et al.*
135 2005; Franke *et al.* 2006) and other areas more intensively used by wolves such as resting sites
136 or other types of behaviour.

137 In this project, my aim was to investigate the habitat characteristics of different types of wolf
138 activity sites (kill sites, bed sites and other sites; tracks and nothing) in *(i)* two different seasons,
139 when bears are still in their winter dens (winter season), and after bears come out of their winter
140 dens (spring season); *(ii)* and the habitat characteristics between wolf kill sites in both these
141 seasons.

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145 **Study area**

146 The study was conducted in Jämtland, Gävleborg and Dalarna counties in central Sweden
147 (61°N, 15°E), covering an area of 2.848 km² (Figure 1). The mean daily temperature in winter
148 and summer is -7°C and 15°C, respectively. Mean precipitation during the period of vegetation
149 growth is 350–450 mm (Swenson *et al.* 1999). Snow is present from November to early May
150 (Steyaert and Frank, 2010).

151 The landscape is hilly, with elevations between 200 to 750 m a.s.l. Lakes and bogs are
152 widespread in the area (Fiebre *et al.* 2001). Coniferous forest of Scots pine (*Pinus sylvestris*),
153 Norway spruce (*Picea abies*) and lodgepole pine (*Pinus contorta*) dominates the landscape.
154 Deciduous trees are also common, including birch (*Betula pubescens*), silver birch (*Betula*
155 *pendula*), aspen (*Populus tremula*) and grey alder (*Alnus incana*) (Friebe *et al.* 2001, Solberg *et*
156 *al.* 2006, Nellemann *et al.* 2007). The ground layer is dominated by several species of berries
157 (*Vaccinium myrtillus*, *Vaccinium vitis-idaea*, *Empetrum nigrum* and *Rubus chamaemorus*),
158 grasses, lichen and mosses (Nellemann *et al.* 2007). A large network of gravel roads has been
159 developed due to extensive commercial logging and forest management practices (Sand *et al.*
160 2008). Nevertheless, human density within the distribution of the Scandinavian wolf population
161 is low, having large areas with less than a 1 person per km² (Wabakken *et al.* 2001; Mattison *et*
162 *al.* 2013). Moose are the most important prey of wolves in Scandinavia (Zimmermann *et al.*
163 2014), representing up to 95% of the food consumed of Scandinavian wolf packs (Sand *et al.*
164 2008). Other ungulate prey are roe deer, red deer, mountain and European hares (*Lepus timidus*
165 L., *Lepus europeus* Pallas), capercaillie (*Tetrao urogallus* L.) and black grouse (*Lyrurus tetrix*
166 L.; Sand *et al.* 2008). Other carnivores are also present in the area, including wolverine (*Gulo*
167 *gulo*), boreal lynx (*Lynx lynx*), red fox (*Vulpes vulpes*), Eurasian badger (*Meles meles*) and
168 Marten (*Martes martes*).

169 **Wolf and brown bear data**

170 Research on wolves is carried out by Skandinavian Wolf Research Project and that of brown
171 bears by Skandinavian Brown Bear Research Project. The SKANDULV research project started
172 in Sweden and Norway in the year of 2000 (based at Grimsö Wildlife Research Station,
173 Sweden), while SBBRP works mostly in Sweden and Norway since 1984 (based at the
174 Norwegian University of Life Science)

175 The three wolf packs included in this study overlap the area with the highest brown bear density
176 in Sweden. According to Kindberg *et al.* (2011), the overall estimated population of brown
177 bears present in 2008 in the three counties was 608 individuals which. The three wolf territories
178 were known as Tandjon, Tenskog and Kukumaki. We used information on the spatial location
179 of GPS-collared brown bears and wolves whose home ranges overlapped and obtained in the
180 course of previous studies on predation.

181 Wolves were captured using snow-tracking during winter to locate the pack. Then they were
182 immobilized from a helicopter following standard procedures (Sand *et al.* 2006; Eriksen *et al.*
183 2011; Kreeger and Arnemo, 2012) and equipped with a GPS neck collar (GPS-Simplex, Web-
184 Direct, or Tellus by Followit, Sweden, or GPS-Plus by Vectronic Aerospace, Germany). Brown
185 bears are usually captured in early spring, after they come out of dens. Brown bears were darted
186 from a helicopter using a remote drug delivery system. They are fitted with GPS collars and
187 VHF abdominal implants (Arnemo, 2006; Evans *et al.* 2012).

188 The scheduling of the spatial locations of marked individuals is described by Milleret 2011.
189 Briefly, a wolf male was located at hourly intervals (24 locations per day) and a female wolf
190 every four hours (6 locations per day). In the case of brown bears the schedule varied along the
191 season of the year. For the purpose of this study the schedule was one position per hour (24
192 positions per day). When 7 locations were saved in the internal memory of the GPS collar, they
193 were sent by SMS (Short Message Service) using GSM (Global System for Mobile
194 communications network) through GPS PLUS Manager V3.0.0 software (Vectronic aerospace).

195 The coordinate system used is RT 90 for Sweden. I used data from 2010 to 2014 for kill sites:
196 Tandsjon 2012 and 2014; Kukumaki 2013 and 2014 and Tensskog 2010 and 2011. Data for bed
197 sites and other sites (tracks and nothing) was obtained in during field work in 2014 for the packs
198 in Tandsjon and Kukumaki.

199 **Data processing, clusters and identification of type of activity**

200 The analysis was based on hourly GPS positions of wolves showing spatial aggregation in
201 clusters. In order to create a cluster, 100 and 60 meters buffers were created around all positions
202 of wolves and bears, respectively. The overlapping area between buffers was defined as a
203 cluster with a unique cluster identification number (Sand *et al.*, 2005, 2008; Zimmermann *et al.*,
204 2007). We used clusters of locations to identify potential kill-, bed- and other-sites. Clusters
205 were visually identified including both consecutive positions and other positions as they
206 represent revisits to the same site, potentially representing the re-use of a carcass. Identified
207 cluster sites were visited in the field to identify the type of cluster. Every third day wolf clusters
208 were visited. In each wolf cluster any sign wolf of activity was registered. We classified the
209 clusters in four categories: (i) carcass sites (ii) bed sites, (iii) tracks and (iv) other. For the study
210 we merged tracks and other into a single category other sites (Table 2). GPS coordinates were
211 recorded in every cluster visited, defined as the exact location of carcass sites, bed sites in the
212 case of resting places (always wolf hair present), and sites with only wolf tracks. In clusters
213 with no sign of activity, the position was defined as the center of the defined cluster. We placed
214 camera traps at carcass sites to record predator activity (*shot mode* – Reconyx HC600 or Scout
215 Guard), far enough to take wide angle shots of the site.

216 To estimated date of death of killed prey carcass sites were classified using the first date and
217 time of the GPS collar. For this study, we classified the clusters as belonging to different
218 seasons considering two time thresholds. First, a cutoff directly related with seasonal variation
219 of brown bear winter season, i.e. when bears are in their winter dens (before the first of April)
220 and spring season, when bears come out of their den (after the first of April). The date of
221 emergence of brown bears is different between males and females with male emergence on

222 average between 6 March and 25 April, with a mean in the 4th of April, whereas females emerge
223 on average of 17 days later (Friebe *et al.* 2001; Manchi and Swenson 2005).

224 Second, we used a cutoff related with the life cycle of moose in Scandinavia. Females give birth
225 to one or two calves over a two week period in late spring (Solberg *et al.* 2007; Haydn, 2012);
226 usually between the late May and early June (Jensen, 2004). Previous predation studies show
227 that predation on moose calves by brown bears usually starts during the third week of May.
228 Therefore, we decided to use the 15th of May as cutoff with the purpose identifying the two
229 periods representing predation on adult and yearling moose (before 15th May) and juvenile
230 moose (after 15th May), respectively. The identified clusters were nearly equally distributed
231 between the two seasons.

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233 **Habitat and other environmental data**

234 The clusters visited in the field were described following pertinent procedures, obtaining
235 information on the type of carcass (if present), the habitat and other predator signs (Appendix
236 1). Carcasses were classified according to the: number of carcasses, species, estimated age
237 (days) since death, earliest/latest date, cause of death, animal age, sex, and also including a
238 sample (usually jawbone) from the killed animal. We described the habitat of the clusters using
239 predefined vegetation categories and for the analyses we used the vegetation type assigned in
240 the field (Table 1). Additionally, we used a GIS vegetation layer (SLU Forest Map with 25m x
241 25m pixel resolution, projection RT90 2.5 gon V of the year 2010) to classify the sites into
242 different vegetation types. The layer is based on a combination of data from the Swedish
243 National Forest inventory and satellite data (www.slu.se). We used the layers of roads network
244 of Scandinavia and human density. The minimum distance to the roads was calculated by using
245 Analysis tools, proximity, Near in Arc Toolbox. To extract information from human density
246 layer we used Overlay, Arc Toolbox.

247 **Analyses**

248 We analyzed the information obtained in the field regarding kill sites (only those classified as
249 wolf killed and probably wolf killed), bed sites and other sites (tracks and nothing) to test if
250 there were differences in the habitat use (type of vegetation) for different activities before/after
251 brown bears came out of winter dens. The potential predictors for the type of vegetation used
252 were: old forest, young forest, and other (Table 1), season (factor) was either 0 when bears are
253 in winter dens and 1 when bears are active, distance to nearest road (m), and human population
254 density (persons/km²). In the first analysis we looked for differences between the three types of
255 cluster: kill sites, bed sites and other sites as response variable and using vegetation type, season
256 (factor), distance to nearest road (m) and human density within county (persons/km²) as
257 potential predictors. In a second analysis, we explored, for kill sites only, if the vegetation types
258 (response variable) differed between seasons (factor), distance to nearest road (m) and human
259 density within county (persons/km²). For both analyses we used multinomial logistic regression
260 models (nnet package in R), using Akaike's information criterion (AIC) for model selection
261 (AICcmodavg package, Mazerolle 2015) and Multimodel Inference Based on (Q)AICc (Akaike
262 1974, Shibata 1981). Parameter estimates and standard errors in the set of best top models were
263 examined to assess the reliability of each variable as a predictor of comparison of different
264 clusters and between kill sites before and after bear emergence of winter dens. All statistical
265 analysis was carried out in R studio (<https://www.rstudio.com/>).

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271 RESULTS

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273 In total we visited 182 clusters, and identified the activity performed by wolves in 68% of them
274 (Table 3). The majority of the clusters where a certain activity was identified were classified as
275 *other sites*. This is probably so because the centre of the cluster is located in slightly different
276 positions and the locations lie at the edge between two types of vegetation (Table 3), or due to
277 vegetation growth since 2010 (See Material and methods). In general, kill and bed sites tended
278 to be more frequent in oldforest (note that land use availability is not considered in this
279 comparison). In contrast, the frequency of other sites tended to be proportionally distributed
280 between old forest and other vegetation types. Human population density was higher in both
281 bed sites and other sites compared to kill sites during winter (Table 4). Kill sites tended to be
282 closer to roads when located in young forest in comparison with those in old forests ($221 \text{ m} \pm$
283 188 and 283 ± 207 , respectively).

284 The types of clusters differed in their habitat characteristics mostly due to the type of vegetation
285 (Table 5). The main effect was due to other sites, which were much more frequently located in
286 open areas such as bogs and clear-cuts (other vegetation type, Table 6, Figure 2). Human
287 population density was the second most important variable explaining differences between sites,
288 and being lower in kill sites than in other cluster types of activity (Table 6). Season was the
289 third variable included, with a very weak effect (Table 5).

290 We further explored the role of habitat characteristics on kill sites using vegetation type as
291 dependent variable (Table 7). Human density was the only habitat variable having an effect,
292 showing that kill sites in old forest were located in areas with relatively higher human density,
293 while those in open areas (bogs and clearcuts) were located in areas with lower human density
294 (Table 8, Figure 2). The effect was strong with statistical support (Table 7). Brown bear season
295 had no apparent effect (Table 7).

298 It is known that wolf is a flexible and opportunistic predator (Peterson and Ciucci, 2003) and one
299 of the most adaptable mammals (Mech and Boitani, 2003). Therefore biological and ecological
300 factors, such as moose abundance, may explain where wolf clusters of activity are located in a
301 modified boreal landscape. The distribution of this ungulate depends on a suitable interspersed
302 of food and protective cover (Dussault et al. 2006), with the seasonal changes in foraging
303 patches affecting space use (Heikkilä and Härkönen, 1993). A previous study at Isle Royale
304 showed that moose will use dense coniferous forest to decrease the probability of detection from
305 wolves and to reduce the chance of wolf attacks (Peterson 1977). In a similar study, Creel et al.
306 (2005) showed how elk are likely to use foraging sites in open terrain with no presence of
307 wolves, but change into forest areas in the presence of them, most likely to benefit from
308 increased forest cover protection. Moreover, Kunkel and Pletscher (2001) suggested that wolves
309 prefer areas with significant forest cover, which enhances their chances to remain undetected by
310 prey. The high number of kill sites detected in old forest could be correlated with moose density
311 distribution.

312 The number of kill sites we found within young forest is lower than old forest (Figure 2), but
313 still quite high compared to other (clear-cuts and water related) Previous studies suggest the use
314 of young forest stands as foraging sites by moose (Parker and Morton, 1978). Some findings in
315 boreal regions showed that there is a positive correlation between moose occurrence and young
316 forests, clear cuts and young pine plantations (Cederlund and Okarma 1988; Cederlund 1989),
317 showing that under some conditions moose have preference for young stands (Edenius, 2002;
318 Bergqvist et al., 2003).

319 In terms of forest management, Gauthier et al. (2009) explains that many ecological, economic
320 and biodiversity indicators are related to different combinations of forest units and age classes.
321 Moose prefer to browse in regenerating clear cuts and in open, homogenous stands (Edenius et
322 al., 2002; Potvin et al. 2005a). In contrast, a high proportion of recent clear-cuts (not

323 regenerating ones) decrease moose browse availability (Courtois et al. 1998), potentially
324 reducing the number of wolf prey within remaining forest patches (Potvin et al. 1999). Previous
325 studies suggest that the response by wildlife to clear-cuts varies between species (Potvin et al.,
326 1999; Smith et al., 1999; Simon et al., 2002). Therefore, species can get both benefits and
327 disadvantages in response to clear-cuts. Caribou (*Rangifer tarandus*), for example, may increase
328 foraging but also increase predation risk (Leclerc et al. 2014). In the case of wolves, Ciucci et
329 al. (2003) suggest that travel routes have higher probability of being closer to a forest edge
330 increasing chances of encountering prey in which wolves use forest as hiding place and prey
331 could be more abundant. On Isle Royale, yearling moose were likely to be killed close to
332 shoreline habitats, where they forage for browse (Montgomery et al., 2014).

333 Previous studies in Scandinavia found that wolves modified their behaviour in spring coinciding
334 with brown bears emergence from winter dens (Manchi and Swensson, 2005 in Milleret C.
335 2011). Dahle et al. (2013) documented the successful hunt from a male brown bear towards a
336 radio-collared female adult moose in spring in Scandinavia. Steyaert and Frank (2010) reported
337 two direct interactions between wolves and bears but due to GPS-collar problems was
338 impossible to reveal the outcome of the encounter. Milleret (2011) showed only one direct
339 interaction in which brown bear gained access by being dominant of the carcass. The same
340 author, conclude that most of the interactions of the study were indirect. In Yellowstone
341 National Park, Gunther and Smith (2004), recorded some interactions at wolf-killed carcasses
342 between wolves and female grizzly bears with cubs. A study carried in Slovenia (Krofel and
343 Kos, 2010) showed the high effect of kleptoparasitism by bears on lynx predations. Previous
344 studies showed that habitat selection by wolves is well known to vary between seasons
345 (Mladenoff et al. 1995, Ciucci et al. 2003) and depends on seasonal variation in the habitat
346 selection and body condition of wolves' prey (Mao et al. 2005, Metz et al. 2012). The presence
347 of old forest is an important component in winter habitat for moose (Hamilton et al., 1980;
348 Welsh et al., 1980), when there is excessive snow cover (Thompson and Vukelich, 1981). In

349 spite of all this previous evidence, we found no change in the type of habitat in which wolves
350 located their activities before and after bear emergence.

351 We found that the number of kill sites in old forest in winter was almost double as compared to
352 spring (Table 3). This is most likely due to snow depth, prey physical condition and the increase
353 of wolf pack cohesion near prey wintering areas (Peterson et al. 1984; Fuller 1989). It is well
354 known that wolves increase their hunting success with deeper snow conditions (Kolenosky,
355 1972; Peterson and Allen, 1974), and prey are more vulnerable to predators during late winter
356 (Ciucci et al. 2003).

357 In Bialowieza forest wolves did not select for a particular type of forest for resting places
358 (Theuerkauf et al., 2003). Most of the bed sites we found were within old forest category ,
359 followed by young forest and other. The number of bed sites described in spring season is
360 higher than in winter, most likely due to the difficulty to reach them due to snow depth, e.g. in
361 Tenskog territory the snow reached 105cm in mid February 2011 (Weather station Hamra and
362 Lillhamra, in Milleret 2011). During winter bed sites tend to be in places with higher human
363 density, with predominant old-young forest cover and closer to roads. Most likely they used
364 areas with dense cover either for snow protection or avoiding human encounters. Zimmermann
365 et al. (2014) showed that wolves preferred to rest at intermediate distances to gravel roads.

366 In previous studies, Fuller et al. (1992) found that most wolf packs in Minnesota were located in
367 areas where human density was ≤ 8 humans/km². Light and Fritts (1994) found dispersing
368 wolves in the Dakotas to be in areas with a mean human density of 3.5 humans/km² and with
369 8.2 humans/km². Jedrzejewski et al. (2005) revealed that when anthropogenic impact is not too
370 high, wolves can get used to human presence by spatiotemporal segregation from people. It is
371 remarkable that human density within the distribution of the Scandinavian wolf population is
372 low, having large areas with less than a 1 person per km² (Wabakken et al. 2001; Mattison et al.
373 2013. In contrast, wolves avoid areas where human disturbance is greater (Karttinen et al.,
374 2015; Guraire et al., 2011). We found that human density best explains the differences between

375 kill sites located in areas with more or less vegetation cover (from old forest to open areas).
376 Thus kill sites within vegetation type «other» shows the lowest human density value (Table 4)
377 most likely to avoid human encounters in open areas. In contrast, wolves tend to kill in areas
378 with high human density and dense forest protection.

379 In conclusion , our results show that vegetation type and human density can influence the choice
380 made by wolves to locate different types of activities. The high occurrence of kill- and bed sites
381 in old and young forest during winter and spring from 2010 to 2014, and also the heavily used
382 other habitat category (clear-cuts and water related areas) for other activities (tracks and
383 nothing) most likely relates with wolves movement behaviour (marking, hunting, controlling
384 territory). It is possible to hypothesize that wolves generally used open areas as other sites (clear
385 cuts and water related) most likely during night when human activity decrease but tend to
386 refuse these areas, at least in activities which involve stay longer such as kill sites or resting
387 places, maybe to avoid human encounter. Nevertheless, in order to provide an adequate
388 assessment of the drivers of space use in relation with the interaction with bears, further
389 monitoring of wolf –bear interactions and the consideration of other variables such as wolf pack
390 size, prey density or bear sex and age, would improve our knowledge towards interspecific
391 interaction between large carnivores in boreal landscapes.

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398

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425 REFERENCES

426

- 427 Akaike, H. 1974. A New Look at Statistical Model Identification, IEEE. Transactions on
428 Automatic Control 19: 716-723
- 429 Anderson C.R., Jr, Lindzey, FG. 2003: Estimating cougar predation rates from GPS location
430 clusters. - Journal of Wildlife Management 67: 307-316
- 431 Arnemo J.M, Ahlqvist P, Andersen R, Berntsen F, Ericsson G, Odden J, Brunberg S, Segerström
432 P, Swenson J.E, 2006. Risk of capture-related mortality in large free-ranging mammals:
433 experiences from Scandinavia. Wildlife Biology 12(1):109-113.
- 434 Balme G., Hunter L, Slotow, R., 2007. Feeding habitat selection by hunting leopards *Panthera*
435 *pardus* in a woodland savanna: prey catchability versus abundance. Animal Behaviour 74: 589-
436 598.
- 437 Ballard W.B, Carbyn L.N, Smith D.W, 2003. Wolf interactions with non-prey. In: Mech LD,
438 Boitani L, editors. Wolves: behavior, ecology, and conservation. Chicago: University of
439 Chicago Press. pp. 259–271.
- 440 Bergqvist G, Bergström R, Edenius L, 2003. Effects of moose (*Alces alces*) rebrowsing on
441 damage development in young stands of Scots pine (*Pinus sylvestris*). Forest Ecology and
442 Management. 176: 397-403
- 443 Cederlund G.N, Okarma H, 1988. Home range and habitat use of adult female moose. J Wild
444 Manag 52:336–343
- 445 Cederlund G, 1989. Activity patterns in moose and roe deer in a north boreal forest. Ecography
446 12:39–46
- 447 Creel S., Winnie J.J., Maxwell B., Hamlin K., Cree M., 2005. Elk alter habitat selection as an
448 antipredator response to wolf. Ecology, Ecological Society of America, 86(12), pp.3387 – 3397
- 449 Ciucci P., Masi M., Boitani L., 2003. Winter habitat and travel route selection by wolves in the
450 northern Apennines, Italy. Ecography 26:223-235
- 451 Chapron G., Kaczensky, P., Linnell J.D.C., Von Arx, M., Huber, D., Andrén, H., López-
452 Bao, J.V., Adamec, M., et al., 2014. Recovery of large carnivores in Europe's modern human-
453 dominated landscapes. Science. Volume 346, Issue 6216, Pages 1517-1519
- 454 Chubbs TE, Keith LB, Mahoney SP, McGrath MJ, 1993. Responses of woodland caribou
455 (*Rangifer tarandus caribou*) to clear-cutting in East-Central Newfoundland. Can J Zool 71:
456 487–493.
- 457 Courtois R, Ouellet J.P, Gagne B, 1998. Characteristics of cutovers used by moose (*Alces alces*)
458 in early winter. Alces 34:201–211
- 459 Courbin N., Fortin D, C. Dussault, V. Fargeot, Courtois R. 2013. Multi-trophic resource
460 selection function enlightens the behavioural game between wolves and their prey. Journal of
461 Animal Ecology 82:1062–1071.

- 462 Dahle B, Wallin K, Cederlund G, Persson I.L, Selvaag L.S, Swenson J.E, 2013. Predation on
463 adult moose *Alces alces* by European brown bears *Ursus arctos*. *Wildl. Biol.* 19: 165-169
- 464 Dussault, C., Poulin, M., Courtois, R. & Ouellet, J-P. 2006: Temporal and spatial distribution of
465 moose-vehicle accidents in the Laurentides Wildlife Reserve, Quebec, Canada. - *Wildl. Biol.*
466 12: 415-425.
- 467 Edenius L, Bergman M, Ericsson G, Danell K., 2002. The role of moose as a disturbance factor
468 in managed boreal forest. *Silva Fenn* 36:57-67
- 469 Eriksen A, Wabakken P, Zimmermann B, Andreassen H.P, Arnemo J.M, Gundersen H, Liberg
470 O, Linnell J, Milner J.M, Pedersen H.C, et al., 2011. Activity patterns of predator and prey: a
471 simultaneous study of GPS-collared wolves and moose. *Anim.Behav.* 81:423-431
- 472 Esseen P.A, Ehnström B, Ericson L, Sjöberg K, 1992. Boreal forests – the focal habitats of
473 Fennoscandia. In: Hansson L (ed) *Ecological Principles of Nature Conservation: Applications in*
474 *Temperate and Boreal Environments*, pp 252–325. Elsevier, London
- 475 Evans A.L, Sahlén V, Støen O-G, Fahlman Å, Brunberg S, Madsløen K, et al., 2012. Capture,
476 Anesthesia, and Disturbance of Free-Ranging Brown Bears (*Ursus arctos*) during Hibernation.
477 *PLoS ONE* 7(7): e40520. doi:10.1371/journal.pone.0040520
- 478 Frair J.L, Fieberg J, Hebblewhite M, Cagnacci F, DeCesare N.J and Pedrotti L., 2010.
479 Resolving issues of imprecise and habitat-biased locations in ecological analyses using GPS
480 telemetry data. *Philosophical Transactions of the Royal Society B: Biological Sciences.*
481 365(1550):2187-2200.
- 482 Friebe A, Swenson J.E, Sandegren F, 2001. Denning chronology of female brown bears (*Ursus*
483 *arctos*) in central Sweden, *Ursus* 12:37-46
- 484 Franke A, Caelli T, Kuzyk G, Hudson RJ, 2006. Prediction of wolf (*Canis lupus*) kill-sites using
485 hidden Markov models. *Ecol. Model.*197:237.
- 486 Fuller, T.K. 1989. Population dynamics of wolves in north-central Minnesota. *Wildlife*
487 *Monographs* 105.
- 488 Fuller T.K., Berg W.E., Radde G.L., Lenarz M.S., Joselyn G.B, 1992. A history and current
489 estimate of wolf distribution and numbers in Minnesota. *Wildlife Society Bulletin*, 20:42-55.
- 490 Gauthier S, Vaillancourt S.A, Leduc A, De Grandpé L, Kneeshaw ., Morin H, Drapeau P,
491 Bergeron Y, 2009. Ecosystem management of the boreal forest. *Presses de la l'Université du*
492 *Quebec*
- 493 Gilroy J.J., Ordiz, A., Bischof, R., 2015. Carnivore coexistence: Value the wilderness . *Science*
494 *Volume* 347, Issue 6220, Page 382
- 495 Gurarie E, Suutarinen J, Kojola I, Ovaskainen O, 2011. Summer movements, predation and
496 habitat use of wolves in human modified boreal forests. *Oecologia* 165: 891-903
- 497 Gunther K.A, Smith D.W, 2004. Interactions between wolves and female grizzly bears with
498 cubs in Yellowstone National Park. *Ursus* 15(2): 232-238

- 499 Hayes RD, Baer AM, Wotschikowsky U, Harestad AS, 2000. Kill rate by wolves on moose in
500 the Yukon. *Can J Zool* 78:49–59
- 501 Haydn A, 2012. Calving site selection by moose (*Alces alces*) along a latitudinal gradient in
502 Sweden. University of Natural Resources and Applied Life Sciences. Master Thesis.
- 503 Hamilton G.D, Drysdale P.D, Euler D.L, 1980. Moose winter browsing patterns on clear-cuts in
504 northern Ontario. *Canadian Journal of Zoology* 58: 1412-1416
- 505 Heikkilä R, Härkönen S, 1993. Moose (*Alces alces* L.) browsing in young scots pine stands in
506 relation to the characteristics of their winter habitats. *Silva Fennica*. Vol 27 No 2: 127-143
- 507 Husseman, J.S, Murray D.L, Power G, Mack C, Wegner C.R, and Quigley H., 2003. Assessing
508 differential prey selection patterns between two sympatric large carnivores. *Oikos* 101:591-601.
- 509 Jensen B, 2004. Nordens däggdjur. 2ed. - Stockholm: Prisma.
- 510 Jedrzejewski W, Niedzialkowska M, W. Mysiajek R, Nowak S, Jedrzejewska B., 2005. Habitat
511 selection by wolves *Canis lupus* in the uplands and mountains of southern Poland. *Acta*
512 *Theriologica* 50 (3): 417-428
- 513 Kauffman M. J, Varley N, Smith D. W, Stahler D. R, MacNulty D. R, Boyce M. S, 2007.
514 Landscape heterogeneity shapes predation in a newly restored predator-prey system. *Ecology*
515 *Letters* 10:690–700
- 516 Kaartinen S, Antikainen H, Kojola I, 2015. Habitat model for recolonizing Wolf (*Canis lupus*)
517 population in Finland. *Annales Zoologici Fennici*, 52(1-2):77-89.
- 518 Kreeger T.J, Arnemo J.M, 2012. Handbook of wildlife chemical immobilization. 4th ed. Sybill
519 (Wyoming): Terry J. Kreeger
- 520 Kindberg J, Swenson J.E, Ericsson G, Bellemain E, Miquel C, Taberlet P, 2011. Estimating
521 population size and trends of the Swedish brown bear (*Ursus arctos*) population. *Wildlife*
522 *Biology* 17: 114-123
- 523 Kolenosky G.B, 1972. Wolf predation on wintering deer in east-central Ontario. *Journal of*
524 *Wildlife Management*, 36: 357–369
- 525 Krofel M, Kos I., 2010. Modeling potential effects of brown bear kleptoparasitism on the
526 predation rate of Eurasian lynx. *Acta Biologica Slovenica*. Vol. 53, 1: 47–54
- 527 Kunkel K.E, Pletscher D.H, 2001. Winter hunting patterns of wolves in and near Glacier
528 National Park, Montana. *Journal of Wildlife Management* 65(3):520-530
- 529 Lecrec M, Dussault C, St-Lauren M.H, 2014. Behavioural strategies towards human
530 disturbances explain individual performance in woodland caribou. *Oecologia*. Vol. 176. (1):
531 297-306
- 532 Linnell J.D.C, Swenson J.E, Andersen R, 2000. Conservation of biodiversity in Scandinavian
533 boreal forest: large carnivores as flagship, umbrellas, indicators or key stones? *Biodiversity and*
534 *Conservation* 9: 857–868

535 Light D.S, Fritts L.H, 1994. Gray wolf (*Canis lupus*) occurrences in the Dakotas. *American*
536 *Midland Naturalist*, 121:387-389.

537 Mao J. S., Boyce M. S, Smith D.W, Singer F.J, Vales D.J, Vore J.M, Merrill E.H. 2005.
538 Habitat selection by elk before and after wolf reintroduction in Yellowstone National Park.
539 *Journal of Wildlife Management* 69:1691–1707.

540 Manchi S, Swenson J.E, 2005. Denning behaviour of Scandinavian brown bears (*Ursus arctos*).
541 *Wildlife Biology* 11:123-132

542 Mech L.D, Boitani L, 2003. *Wolves: behaviour, ecology and conservation*. The University of
543 Chicago Press, Chicago and London.

544 Mattison J, Persson J, Andren H, Segerström P, 2011. Temporal and spatial interactions
545 between an obligate predator, the Eurasian lynx (*Lynx lynx*) and a facultative scavenger, the
546 wolverine (*Gulo gulo*). *Canadian Journal of Zoology* 89 (2): 79-89.

547 Mattisson J, Sand H, Wabakken P, Gervasi V, Liberg O, Linnell J.D, Rauset G.R, Pedersen
548 H.C, 2013. Home range size variation in a recovering wolf population: evaluating the effect of
549 environmental, demographic, and social factors. *Oecologia*, 1-13.

550 Mazerolle M.J, 2015. Model Selection and Multimodel Inference Based on (Q)AIC(c). Package
551 ‘AICcmodavg’. CRAN

552 Montgomery R.A, Vucetich J., Roloff G.J, Bump J.K, Peterson R.O, 2014. Where wolves kill
553 moose: The influence of prey life history dynamics on the landscape ecology of predation. *PLoS*
554 *ONE* 9(3): e91414. doi:10.1371/journal.pone.0091414

555 Mladenoff D.A, Sickley T.A, Haight R.A, Wydeven A.P, 1995. A regional landscape analysis
556 and prediction of favorable gray wolf habitat in the northern Great Lakes region. *Conservation*
557 *Biology* 9(2):279-294.

558 Milleret C, 2011. Estimating wolves (*Canis lupus*) and brown Bear (*Ursus arctos*) interactions in
559 Central Sweden. Does the emergence of brown bears affect wolf predation patterns? Master
560 Thesis

561 Miller B, Dugelby B, Foreman D, Martinez Del Rio C, Noss R, Phillips M, Reading R, Soulé
562 M.E, Terborgh J, Willcox L, 2001. The importance of large carnivores to healthy ecosystem.
563 *Endangered species updates*. 18:202-210

564 Metz M.C, Smith W.D, Vucetich J.A, Stahler D.R, Peterson R.O, 2012. Seasonal patterns of
565 predation for gray wolves in the multi-prey system of Yellowstone National Park. *Journal of*
566 *Animal Ecology*. doi: 10.1111/j.1365-2656.2011.01945.x

567 Nellemann C, Stoen O-G, Kingberg J, Swenson J.E, Vistnes I, Ericsson J, Katajisto J,
568 Kaltenborn B.P, Martin J, Ordiz A, 2007. Terrain use by an expanding brown bear population in
569 relation to age, recreational resorts and human settlements. *Biological Conservation* 138:157-
570 165

571 Palomares F, Caro TM. 1999. Interspecific killing among mammalian carnivores. *American*
572 *Naturalist*. 153: 492-508

- 573 Parker G.R, Morton L.D, 1978. The estimation of winter forage and its use by moose on clear-
574 cuts in northcentral Newfoundland. *Journal of Range Management* 31(4):300-30
- 575 Peterson, R.O, Ciucci P, 2003. The wolf as a carnivore. Pages 104-130 in D.L. Mech, and L.
576 Boitani, editors. *Wolves : behaviour, ecology and conservation*. The University of Chicago
577 Press, Chicago and London.
- 578 Peterson R.O, Allen D.L, 1974. Snow conditions as a parameter in moose-wolf relationships.
579 *Canadian Naturalist*, 101 (1974), pp. 481–492
- 580 Peterson RO, 1977. Wolf ecology and prey relationship on Isle Royale. U.S Nat. Park Serv.
581 *Sci.Monogr.No.11*
- 582 Peterson RO, Woolington JD, Bailey TN, 1984. Wolves of the Kenai Peninsula, Alaska. *Wildl.*
583 *Monogr. No.88*
- 584 Potvin F, Courtois R, Belanger L, 1999. Short – term response of wildlife to clear-cutting in
585 Quebec boreal forest: multiscale effects and management implications. *Can.J.For.Res.* 29: 1120-
586 1127
- 587 Potvin F, Breton L, Courtois R., 2005a. Response of beaver, moose and snowshoe hare to clear-
588 cutting in a Quebec boreal forest: a reassessment 10 years after cut. *Can J For Res* 35:151–160
- 589 R Development Core Team 2010: R: A language and environment for statistical computing. - R
590 Foundation for Statistical Computing. Vienna, Austria. Available at: <http://www.r-project.org>
- 591 Ripple W.J., Estes, J.A., Beschta, R.L., Wilmers, C.C., Ritchie, E.G., Hebblewhite, M., Berger,
592 J., Elmhagen, B., Letnic, M., Nelson, M.P., Schmitz, O.J., Smith, D.W., Wallach, A.D., 2014.
593 Status and ecological effects of the world's largest carnivores. *Science*. Volume 343, Issue 6167,
594 Page 1241484
- 595 Rostro-García S, Kamler JF, Hunter LTB. 2015. To Kill, Stay or Flee: The Effects of Lions and
596 Landscape Factors on Habitat and Kill Site Selection of Cheetahs in South Africa. *PLoS ONE*
597 10(2): e0117743. doi:10.1371/journal.pone.0117743
- 598 Shibata R., 1981. An Optimal Selection of Regression Variables. -*Biometrika* 68: 45-54.
- 599 Sand H, Zimmermann B, Wabakken P, Andren H, 2005. Using GPS – technology and GIS-
600 cluster analyses to estimate kill rates in wolf-ungulate ecosystems. *Wildlife Society Bulletin*.
601 33: 914-925
- 602 Sand H, Wabakken P, Zimmermann B, Johansson O, Pedersen H, Liberg O, 2008. Summer kill
603 rates and predation pattern in a wolf-moose system: can we rely on winter estimates? *Oecologia*.
604 156:53-64
- 605 Smith S, Hardin P.J, Flinders J.T 1999. Response of bighorn sheep to clear-cut logging and
606 prescribed burning. *Wildl. Soc. Bull.*, 27 (3) pp. 840–845
- 607 Smith KG, Ficht EJ, Hobson D, Sorensen TC, Hervieux D, 2000. Winter distribution of
608 woodland caribou in relation to clear-cut logging in west-central Alberta. *Can J Zool* 78: 1433–
609 1440

- 610 Smith T.S, Partridge S.T, Schoen J.W, 2003. Interactions of Brown Bears, *Ursus arctos*, and
611 Gray Wolves, *Canis lupus*, at Katmai National Park and Preserve, Alaska. *Canadian Field-*
612 *Naturalist* 118(2): 247-250
- 613 Smith DW, Drummer TD, Murphy KM, Guernsey DS, Evans SB, 2004. Winter prey selection
614 and estimation of wolf kill rates in Yellowstone National Park, 1995–2000. *Journal of Wildlife*
615 *Management.*; 68:153–66.
- 616 Schwartz C.C, Swenson J.E, Miller S.D, 2003. Large carnivores, moose and humans: a
617 changing paradigm of predator management in the 21st century. *Alces* Vol. 39: 41-63.
- 618 Singh N.J, Leonardsson K., 2014. Partial migration and transient coexistence of migrants and
619 residents in animal populations. *PLOS ONE*. Volume 9. Issue 4. E94750
- 620 Solberg, K. H., Bellemain E., Drageset O.-M., Taberlet P, Swenson J.E, 2006. An evaluation of
621 field and non-invasive genetic methods to estimate brown bear (*Ursus arctos*) population size.
622 *Biological Conservation* 128:158–168
- 623 Solberg E.J, Heim M, Grøtan V, Sæther B-E, Garel M. 2007. Annual variation in maternal age
624 and calving date generate cohort effects in moose (*Alces alces*) body mass. *Oecologia*. 154:
625 259-271.
- 626 Steyaert S, Frank C. S, 2010. Wolf-brown bear interactions in Central Sweden. Pilot Study.
627 *Scandinavian Brown Bear Research Project & Scandinavian Wolf Project*. 19 p
- 628 Swenson J.E, Jansson A, Riig R, Sandegren F., 1999. Bears and ants: myrmecophagy by brown
629 bears in central Scandinavia. *Can. J. Zool.* 77 (4), 551–561.
- 630 Theuerkauf J, Rouys S, Jedrzejewski W, 2003. Selection of den, rendezvous, and resting sites
631 by wolves in the Bialowieza Forest, Poland. *Can. J.Zool.* 81: 163-167
- 632 Thompson I.D, Vukelich M.F, 1981. Use of logged habitat in winter by moose with calves in
633 northeastern Ontario. *Canadian Journal of Zoology* 59:2103-2114
- 634 Urbano F, Cagnacci F, Calenge C, Dettki H, Cameron A, Neteler M., 2010. Wildlife tracking
635 data management: a new vision. *Philosophical Transactions of the Royal Society B: Biological*
636 *Sciences*. 365(1550):2177–2185
- 637 Wabakken P, Sand H, Liberg O, Bjärvall A, 2001. The recovery, distribution, and population
638 dynamics of wolves on the Scandinavian peninsula, 1978-1998. *Can. J. Zool.* 79: 710-725
- 639 Welsh D.A, Morrison K.P, Oswald K, Thomas E.R, 1980. Winter utilization of habitat by
640 moose in relation to forest harvesting. *Proceedings of North American Moose Conference*
641 *Workshop* 16:398-428
- 642 Wittmer H.U, McLellan B.N, Serrouya R, Apps C.D., 2007. Changes in landscape composition
643 influence the decline of threatened woodland caribou population. *J Anim Ecol* 76:568–579
- 644 Yurewicz KL, Wilbur HM, 2004. Resource availability and cost of reproduction in the
645 salamander *Plethodon cinereus*. *Copeia* (1): 28-36

646 Zimmermann B, Wabakken P, Sand H, Pedersen HC, Liberg O., 2007. Wolf movement
647 patterns: a key to estimation of kill rate? *Journal of Wildlife Management* 71:1177-1182

648 Zimmermann B, Nelson L, Wabakken P, Sand H, Liberg O, 2014. Behavioral responses of
649 wolves to roads: scale – dependant ambivalence. *Behavioral Ecology*, 25(6), 1353-1364

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667 TABLES AND FIGURES

668

669 **Table 1.** Main habitat types used in field protocols. Re-classification of habitat into three main
670 vegetation types -other, young and old forest used in the study.

Habitat type	Description
RZ - Riparian zone	A river > 3m wide within the 30-meter plot.
B - Bog	Peat ground with very low productivity. No flow or ground water
TRB – Tree rich bog	Similar to bog but rich in trees.
I - Impediment	Natural barriers, rift, holes and other related.
SW/SF - Swamp/Swamp forest	Waterlogged ground (not on peat) often with broadleaf grasses and herbs. There is outflow of water and high productivity.
W – Water	River or lake within 30 m radius of the plot
Rd – Road	Small gravel roads, roadsides and main roads
OT – Other	Any habitat not described previously
Vegetation types	
Old forest (mature forest)	(G1/S1) with interval of age from 25 – 30 years and > 30 years old
Young forest	(R1/R2) with interval of age in between 5 – 25 years old
Other (clear-cuts/water related)	(K1/K2) with interval of age 0 – 5 years old and water related.

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672

673 **Table 2.** Definition of clusters types used in the study.

Cluster type	Description
<p>Kill sites (prey remains)</p>	<p>Clusters defined as kill site place with the presence of a carcass. Evidences of fight between wolf-prey. Often the rumens prey is spread out. Often strong smell. Pipe bleeding. Ground scratches.</p> <p>Wolf killed prey: clear evidences of fight. Warm prey blood pumped out of arteries/veins make little tubes called pipes through the cold snow which is called “pipe-bleeding” (Sand <i>et al.</i>, 2008). Blood clots upon tree trunks or branches. Often rumens prey is spread out. Occasionally, we could observe clear tracks, scratches and urine typical of chasing prey (Hayes <i>et al.</i>, 2000, Sand <i>et al.</i>, 2005, 2008).</p> <p>Probably wolf killed: the estimation of the time of prey death coincided with time of the GPS-positions (Sand <i>et al.</i> 2008). The degree of consumption and decomposition depends on the time elapsed and the weather.</p> <p>Not wolf killed prey: the prey was bounded by hunting or road killed.</p> <p>Other potential killers: difficult to discern between wolf and other predators, but with some evidence of other predator. For example, evidence of bear behavior around at similar date/time of collar positions.</p> <p>Unknown: not classified into the previous categories.</p>
<p>Bed sites</p>	<p>Clusters related with resting behavior. Beds always confirmed with the presence of wolf hair on it.</p>
<p>Other sites</p>	<p>Combination of tracks and nothing clusters. In both type of clusters wolves spent time for behavioral reason.</p> <p>Tracks: Clusters with clear wolf tracks and evidences of being a way of passing back and forth through the same spot.</p> <p>Nothing: Clusters with no clear evidences at human view but wolves used these places.</p>

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675

676 **Table 3.** Vegetation types in wolf different types of clusters in Scandinavia, as described in field protocols that were filled in during wolf predation studies
 677 and from GIS layers.

678

Cluster type	Field sampling						GIS data					
	Kill sites (n=47)		Bed sites (n=65)		Other sites (n=70)		kill sites (n=47)		Bed sites (n=65)		Other sites (n=70)	
Period	0	1	0	1	0	1	0	1	0	1	0	1
Vegetation type												
Old forest	16	9	14	19	12	15	17	15	14	27	15	14
Young forest	6	9	12	13	7	8	8	0	9	4	3	19
Other	3	4	3	4	5	23	1	6	4	7	6	13
TOTAL	25	22	29	36	24	46	26	21	27	38	26	44

Period: 0 (winter season), 1 (spring season)

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680

681 **Table 4.** The table represents the values (mean \pm S.E) of independent variables used in the study in relationship with different cluster types (n=182). Results
 682 of human density values (mean \pm S.E) in kill sites (n=47) in relationship with vegetation type.

683

Period	Distance to roads (m)		Human density (person/km ²)				
	0	1	0	1	Old forest	Young forest	Other
Cluster type							
Kill sites	247 \pm 168	302 \pm 232	3.44 \pm 2.01	3.36 \pm 2.18	3.98 \pm 1.92	3.61 \pm 1.97	0.87 \pm 0
Bed sites	243 \pm 152	261 \pm 250	4.34 \pm 2.01	3.64 \pm 2.28	-	-	-
Other sites	219 \pm 179	286 \pm 368	4.05 \pm 1.92	3.84 \pm 2.32	-	-	-

684

Period: 0 (winter season), 1 (spring season)

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686

687 **Table 5.** Model selection of different cluster types (kill sites, bed sites and other sites) before
 688 and after bear den emergence in Scandinavia based on AICc, Δ AIC and wAIC.

689

Kill sites, bed sites and other sites

	K	cAIC	ΔAIC	wAIC
Vegetation type	6	388.30	0.00	0.33
Vegetation type + Human density	8	388.52	0.23	0.30
Vegetation type + Human density + Season	10	390.07	1.78	0.14
Vegetation type + Season	8	390.37	2.08	0.12
Vegetation type + Distance roads	8	392.44	4.14	0.04
Vegetation type + Distance roads + Human density	10	392.95	4.66	0.03
Vegetation type + Distance roads + Season	10	394.48	6.18	0.02
Vegetation type + Distance roads + Human density + Season	12	394.56	6.26	0.01
Vegetation type x Season	12	395.43	7.14	0.01
Season	4	398.86	10.56	0.00
Intercept only	2	398.95	10.65	0.00
Human density + Season	6	399.46	11.16	0.00
Human density	4	400.26	11.96	0.00
Distance roads + Season	6	402.83	14.54	0.00
Distance roads	4	402.93	14.64	0.00
Distance roads + Human density + Season	8	403.70	15.71	0.00
Distance roads + Human density	6	404.43	16.13	0.00
Distance roads x Human	8	408.13	19.83	0.00

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691

692 **Table 6.** Results of multinomial logistic regression with a set of top models ($\Delta AIC < 2$) to compare wolf kill sites, bed sites and other sites (nothing and
 693 tracks), before and after bear emergence in Scandinavia.

	Predictors						
	Intercept	Vegetation type		Human density		Season	
		other	young	old		winter	spring
Model 1: Vegetation type							
Kill site	- 0.277 (0.265)	0.277 (0.596)	-0.233 (0.420)	0	-	-	-
Other site	-0.200 (0.259)	1.586 (0.495) ***	-0.310 (0.417)	0	-	-	-
Bed site	0	0	0	0	-	-	-
Model 2: Vegetation type + Human density							
Kill site	0.168 (0.437)	0.172 (0.604)	-0.264 (0.423)	0	-0.114 (0.089)	-	-
Other site	-0.499 (0.458)	1.642 (0.503) ***	-0.290 (0.418)	0	0.069 (0.087)	-	-
Bed site	0	0	0	0	0	-	-
Model 3: Vegetation type + Human density + Season							
Kill site	-0.023 (0.473)	0.267 (0.610)	-0.251 (0.425)	0	-0.124 (0.090)	0.430 (0.394)	0
Other site	-0.409 (0.478)	1.596 (0.507) ***	-0.301 (0.419)	0	0.076 (0.088)	-0.259 (0.375)	0
Bed site	0	0	0	0	0	0	0
Estimates (S.E). ($P \leq .001$) ***							

695 **Table 7.** Model selection of wolf kill sites before and after bear den emergence in Scandinavia
 696 based on AICc, Δ AIC and wAIC.

Kill sites	K	cAIC	ΔAIC	wAIC
Human density	4	83.44	0.00	0.65
Human density + Season	6	86.06	2.62	0.17
Human density + Distance roads	6	87.06	3.61	0.11
Human density x Distance roads	8	88.76	5.32	0.05
Human density + Distance roads + Season	8	89.90	6.46	0.03
Intercept only	2	96.76	13.32	0.00
Season	4	98.90	15.45	0.00
Distance roads	4	99.38	15.94	0.00
Distance roads + Season	6	101.72	18.28	0.00

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708 **Table 8.** Results of multinomial logistic regression with the best top model ($\Delta AIC < 2$) to
709 compare wolf kill sites before and after bear emergence in Scandinavia.

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Model: Human density	Predictor	
	Intercept	Human density
Vegetation type		
Old forest	0	0
Young forest	-0.116 (0.723)	-0.103 (0.171)
Other	4.146 (23.93)	-4.370 (27.49)

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Estimates (S.E)

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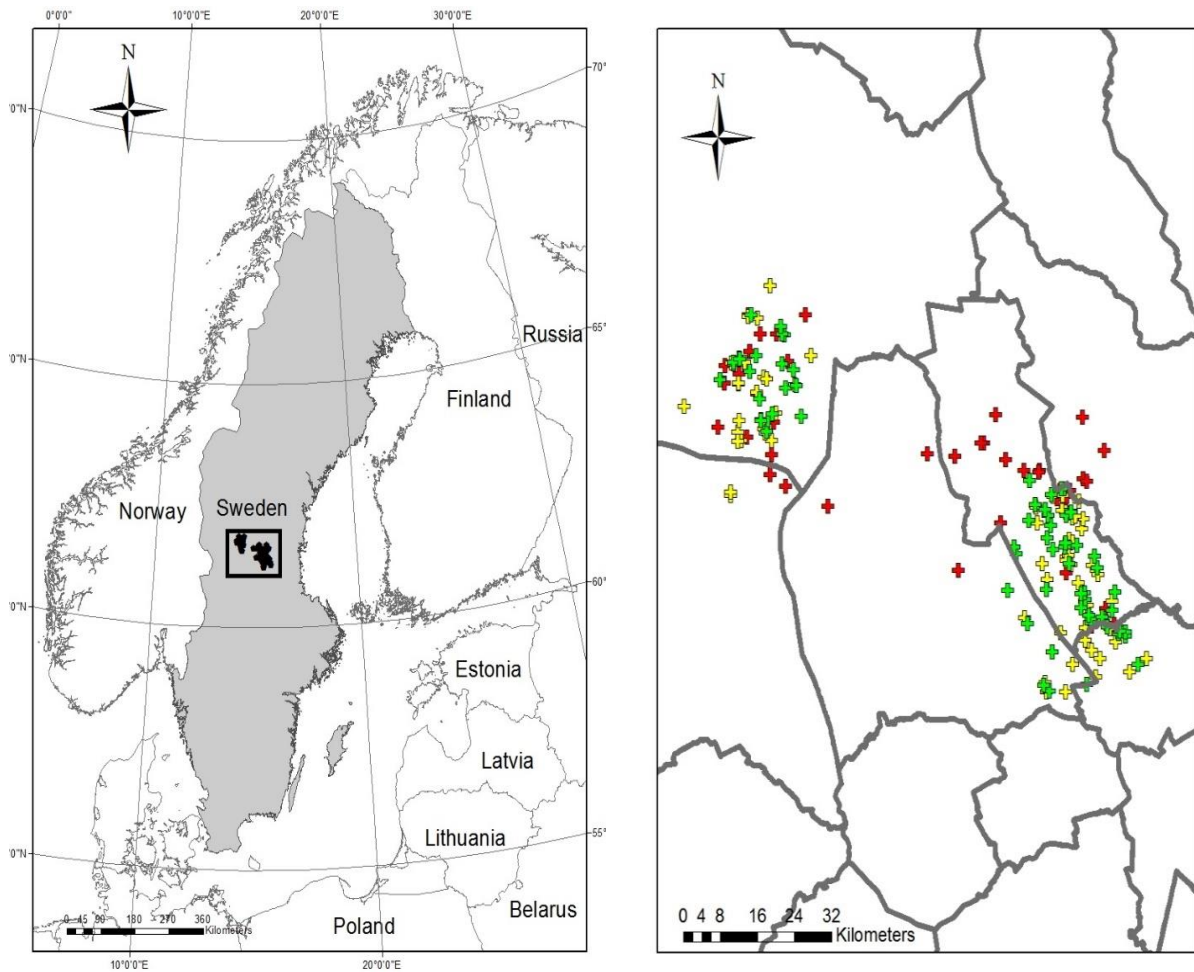
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725 **Figure 1.** Map of the study area in Scandinavian Peninsula (grey area, black square) and
 726 wolf territories (black dots) within three counties – Jämtland, Gävleborg and Dalarna in central
 727 Sweden. Total of wolf kill sites (red crosses), bed sites (green crosses) and other sites (yellow
 728 crosses) in the study area.

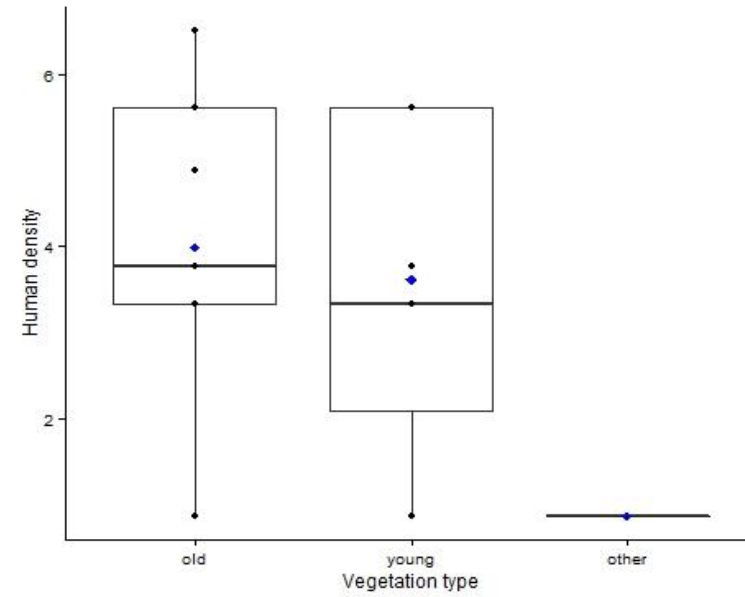
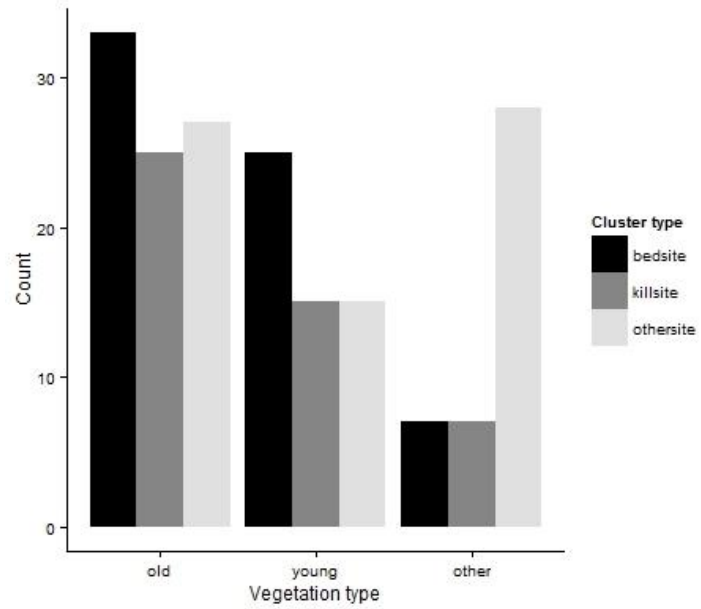


Figure 2. Bar plot representing cluster type counts resulted in field protocols (n=182) in relationship with vegetation type. Box plot representing the values of human density on wolf kill sites (n=47) in relationship to vegetation type.

731 APPENDIX

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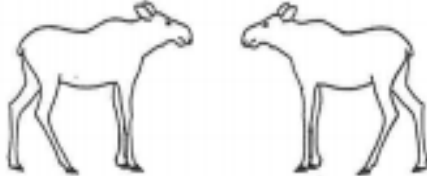
733 **Appendix 1:** Field protocol used for predation studies in Scandinavia.


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Wolf-Bear Pilot Study: Carcass Protocol 2011

(last updated on 2011-05-27 by SHF, JøR, GRRs, and Krvd)

Protocol Information	Protocol ID: <small>(e.g. PR00070-1-003)</small>		Observer:	Year ___ Month ___ Day ___	Start: Stop:	
	Related to:			Plot Selection:		
	Cluster Coord. N: E:		<input type="checkbox"/> Wolf-bear cluster <input type="checkbox"/> Bear cl <input type="checkbox"/> Wolf cl Other: ___			
	Wolf Name(s)		ID:	Date:	Time:	
Bear Name(s)		ID:	Date:	Time:		
Carcass	No. of Carcasses:		Carcass Coord. N: E:			
	Species:		Carcass Est. Age (days):			
	Earliest / Latest Date: / /		Animal Age:		Sex:	
	Cause of Death: <input type="checkbox"/> Wolf <input type="checkbox"/> Bear <input type="checkbox"/> Other predators (sp.) _____ <input type="checkbox"/> Non predator _____ <input type="checkbox"/> Unknown _____		<input type="checkbox"/> 0-1 <input type="checkbox"/> >1 <input type="checkbox"/> unknown		<input type="checkbox"/> M <input type="checkbox"/> F <input type="checkbox"/> Unknown	
Carcass # of Pieces (<input type="checkbox"/> sampled):						
Mandible: (<input type="checkbox"/>) Bone(s): (<input type="checkbox"/>) Skull: (<input type="checkbox"/>) Hair: (<input type="checkbox"/>) Other: (<input type="checkbox"/>)						
Habitat	In Habitat:	Ground/Field:	Moisture:	Ruggedness:	Picture Taken: <input type="checkbox"/> Yes <input type="checkbox"/> No	





Ingested/Missing Part:

Consumed/Missing of Carcass: ___ %

Fresh Bleeding on Snow/Ground: Yes No

Concentrated "pipebleeding": Yes No

Predator Signs	Sign Type	Bear	Wolf	Otherspecies: _____
	Tracks (Picture <input type="checkbox"/> Y <input type="checkbox"/> N)	<input type="checkbox"/> Y <input type="checkbox"/> N <input type="checkbox"/> Unsure	<input type="checkbox"/> Y <input type="checkbox"/> N <input type="checkbox"/> Unsure	<input type="checkbox"/> Y <input type="checkbox"/> N <input type="checkbox"/> Unsure
	Scat (Sampled <input type="checkbox"/> Y <input type="checkbox"/> N)	<input type="checkbox"/> Y <input type="checkbox"/> N <input type="checkbox"/> Unsure	<input type="checkbox"/> Y <input type="checkbox"/> N <input type="checkbox"/> Unsure	<input type="checkbox"/> Y <input type="checkbox"/> N <input type="checkbox"/> Unsure
	Hair/Fur (Sampled <input type="checkbox"/> Y <input type="checkbox"/> N)	<input type="checkbox"/> Y <input type="checkbox"/> N <input type="checkbox"/> Unsure	<input type="checkbox"/> Y <input type="checkbox"/> N <input type="checkbox"/> Unsure	<input type="checkbox"/> Y <input type="checkbox"/> N <input type="checkbox"/> Unsure
	Bed (confirmed with hair)	<input type="checkbox"/> Y <input type="checkbox"/> N #	<input type="checkbox"/> Y <input type="checkbox"/> N #	<input type="checkbox"/> Y <input type="checkbox"/> N #
Other sign: _____				

Other: (clarification, comments, sketch of area/hunt)

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